

Geothermal ecosystems as natural climate change experiments: The ForHot research site in Iceland as a case study

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ABSTRACT

This article describes how natural geothermal soil temperature gradients in Iceland have been used to study terrestrial ecosystem responses to soil warming. The experimental approach was evaluated at three study sites in southern Iceland; one grassland site that has been warm for at least 50 years (GO), and another comparable grassland site (GN) and a Sitka spruce plantation (FN) site that have both been warmed since an earthquake took place in 2008. Within each site type, five ca. 50 m long transects, with six permanent study plots each, were established across the soil warming gradients, spanning from unwarmed control conditions to gradually warmer soils. It was attempted to select the plots so the annual warming levels would be ca. +1, +3, +5, +10 and +20 °C within each transect. Results of continuous measurements of soil temperature (Ts) from 2013-2015 revealed that the soil warming was relatively constant and followed the seasonal Ts cycle of the unwarmed control plots. Volumetric water content in the top 5 cm of soil was repeatedly surveyed during 2013-2016. The grassland soils were wetter than the FN soils, but they had sometimes some significant warming-induced drying in the surface layer of the warmest plots, in contrast to FN. Soil chemistry did not show any indications that geothermal water had reached the root zone, but soil pH did increase somewhat with warming, which was probably linked to vegetation changes. As expected, the potential decomposition rate of organic matter increased significantly with warming. It was concluded that the natural geothermal gradients at the ForHot sites in Iceland offered realistic conditions for studying terrestrial ecosystem responses to warming with minimal artefacts.

Keywords: geothermal soil warming; subarctic grasslands; climate change; spruce forest; decomposition

ÚTDRÁTTUR

Jarðhitavistkerfi sem náttúrulegar loftslagsbreytingartilraunir: ForHot verkefnið á Íslandi sem sýnidæmi.

Þessi grein lýsir því hvernig jarðhitasvæði hérlandis hafa verið notuð til að rannsaka áhrif hlýnunar á norðlæg þurrlendisvistkerfi. Rannsóknirnar fóru fram á þremur stöðum í Ölfusi, í næsta nágrenni Hveragerðis: i) í graslendum sem hafa verið undir áhrifum jarðvegshlýnunar í langan tíma (GO), eða allavega í 50 ár, ii) í samskonar graslendum sem byrjuðu fyrst að hitna vorið 2008 eftir Suðurlandsskjálftann (GN) og iii) í gróðursettum 50 ára sitkagrenisskógi sem einnig byrjaði að hitna vorið 2008 (FN). Á hverjum stað voru lögð út fimm um 50 m löng snið, þvert á hitastigla svæðanna, og sex fastir mælireitir lagðir út á hverju sniði þannig að einn var á úpphituðum jarðvegi (samanburðarreitur) en hinir á síheitari jarðvegi. Reynt var að velja þannig að upphitunin yrði sem næst +1, +3, +5, +10 og +20 °C. Samfelldar mælingar á jarðvegshita reitanna 2013-2015 sýndu að upphitunin hélst tiltölulega stöðug og jarðvegshiti þeirra sveiflaðist líkt og í úpphituðum reitum milli árstíða. Vatnsinnihald yfirborðsjarðvegs (0-5 cm) var mælt reglulega yfir tímabilið 2013-2016. Graslending höfðu að jafnaði rakari yfirborðsjarðveg en skógurinn, en þau, ólíkt skóginum, sýndu einnig stundum marktæka uppbörnun á heitustu reitunum. Efnagreiningar sýndu engin merki þess að jarðhitavatn næði upp í jarðveg svæðanna. Sýrustig jarðvegs hækkaði aðeins (varð basískara) með auknum hita sem tengdist líklega gróðurfarsbreytingum. Niðurbrotsgeta á lífrænu efni jókst með jarðvegshita í öllum reitunum, eins og búist var við. Lokaniðurstaðan var að jarðhitasvæði ForHot verkefnisins framkölluðu aðstæður sambærilegar við ýmsar stýrðar jarðvegsupphitunartilraunir erlendis sem notaðar eru til að rannsaka áhrif hlýnunar á þurrlendisvistkerfi.

INTRODUCTION

The impacts of changing climatic conditions on terrestrial ecosystems at high latitudes (>60 °N) are a very active research field today (Hyvönen et al. 2007, Way & Oren 2010, Kayler et al. 2015). The typical cold conditions at high latitudes strongly constrain decomposition of organic material (McGuire et al. 2009) and over time this has led to the build-up of organic-matter rich soils with relatively slow nutrient cycles in this region. As a consequence, northern high latitude soils store almost 30% of the global soil carbon (C) stocks (Scharlemann et al. 2014), even if they only cover ca. 5% of the terrestrial global soil surface (CAFF 2013). It is still uncertain how fast these ecosystems will change in a warmer world (IPCC 2013), which makes them especially important to study.

Many different experimental approaches have been used to evaluate the effects of future warming on ecosystem structure and function including: i) monitoring of natural variation in temperature and ecosystem variables, ii) climate gradient studies, iii) process-based modelling of plant and ecosystem responses to warming, and iv) experimental indoor or outdoor warming of plants, soils or whole ecosystems (cf. Rustad 2008, De Boeck et al. 2014). Each method

has its pros and cons, but there is a consensus among most ecosystem ecologists that *in situ* warming of ecosystem components and/or whole ecosystems is the most powerful tool that allows for the elucidation of cause-and-effect relationships and provides a mechanistic understanding of ecosystem responses to climate change (Rustad, 2008; Way & Oren 2010, De Boeck et al. 2014). Manipulation experiments tend, however, to be technically challenging and expensive to run, which most often limit their duration to only a few years and include only one or a few warming levels (De Boeck et al. 2014).

Given that anthropogenic warming is occurring on a relatively short time scale, transient biological responses are likely in the initial stages (O’Gorman et al. 2014). Rustad (2008) stated in her review that there is a great need to conduct longer-term warming studies (decadal responses) in order to better understand changes that occur on multiannual time scales. Recently it has also been stressed that far too few *in situ* ecosystem warming experiments have been designed to impose warming gradients to identify possible response thresholds in ecosystem responses (De Boeck et al. 2014, Kayler et al. 2015). One possibility to meet both

these requirements, i.e. studying decadal *in situ* effects to warming and to include large warming gradients, without requiring excessive funds and technological complexity, is to use ecosystems affected by natural geothermal activity.

There are, however, some challenges to using geothermal gradients as warming experiments, and of those we feel that four issues are especially important: i) Geothermal gradients can have existed for a very long time, even centuries or millennia, during which their communities and soils may have been shaped not only by temperature, but also by a myriad of other ecological and evolutionary factors operating at longer time scales (De Boeck et al. 2014, O’Gorman et al. 2014). Under such conditions, the geothermal gradients can function similarly to larger elevational or latitudinal gradients, even if they are spatially confined. Sometimes the time of the onset of the geothermal warming is known (Kayler et al. 2015), and then such gradients can offer comparable conditions to ecosystem manipulation studies that focus on annual to decadal responses to warming. ii) An advantage of geothermal gradients compared to most other natural temperature gradients is that they are much more confined in space, which reduces the potential influence of other confounding environmental factors (O’Gorman et al. 2014). This is only true, however, if geothermal water is not able to reach the soil and the root zone of the warmed plots, since it is typically rich in dissolved minerals that can have ecotoxic effects (Wetang’ula & Snorrason 2005). iii) Geothermal gradients only warm the soil and thereby their effect on processes dominated by air temperature rather than soil temperature may be different from climate warming. This is also an issue with other types of soil warming experiments commonly used to study the effects of warming on ecosystems (Streit et al. 2013, Schindlbacher et al. 2015). iv) Geothermal gradients normally start with a stepwise change in soil temperature, which differs from the gradual changes of Earth’s climate. This may affect how the ecosystem responses develop (De Boeck et al. 2014) and therefore it is important

to include multiannual or decadal responses when such results are extrapolated in relation to climate change. Again, this is something that geothermal gradients have in common with other manipulation studies.

Some earlier manipulation experiments have shown that N limited northern ecosystems may respond to warming mainly through effects on soil processes (cf. Way & Oren, 2010). For example, in a recent large-scale warming experiment on mature Norway spruce (*Picea abies*) forest in northern Sweden it was discovered that only increasing the air temperature did not significantly change tree volume growth (Sigurdsson et al. 2013), whereas, when only the soils were warmed at the same site, the forest productivity responded strongly (Strömgren & Linder 2002). This was explained with soil warming enhancing N availability through increased soil organic matter decomposition and the strong N limitation that exists in most high-latitude ecosystems, which can override the direct effects of air temperature on aboveground processes. Recently, there have been various extensive experiments warming only the soil in different alpine and northern ecosystems (cf. Rustad 2001, Strömgren & Linder 2002, Patil et al. 2013, Streit et al. 2013, Schindlbacher et al. 2015) as well as a few experiments warming both soils and aboveground air separately (Bronson & Gower 2010, Krassovski et al. 2015).

Some research on either experimental or natural *in situ* warming has been carried out previously in Iceland, but never on such a large spatial scale or looking at as many terrestrial ecosystem levels as the ForHot project (e.g. Sigurdsson 2001, Bergh et al. 2003, Elmarsdóttir et al. 2003, Jónsdóttir et al. 2005, Dalebeler et al. 2014, 2015). The most comparable work in Iceland is an ongoing project on naturally warmed stream ecosystems in the nearby Hengill area (e.g. Woodward et al. 2010, O’Gorman et al. 2014).

In this paper we address how stable the hourly to annual geothermal soil warming was at 10 cm depth at three different geothermal gradients in southern Iceland, as well as how

the warming was distributed vertically, both in the soil and in the air above the study plots. Moreover, we address how other potential confounding environmental drivers were affected by the geothermal warming, including changes in soil pH, soil humidity and if there were any indications of potential pollution by geothermal water in the soil profile, all of which can determine how suitable those sites are for

studying the effects of warming on terrestrial ecosystem structure and function.

METHODS

Experimental setup

The study sites were located in south Iceland, close to the village of Hveragerði (64.008°N, 21.178°W; 83-168 m a.s.l.), on the grounds of the Agricultural University of Iceland

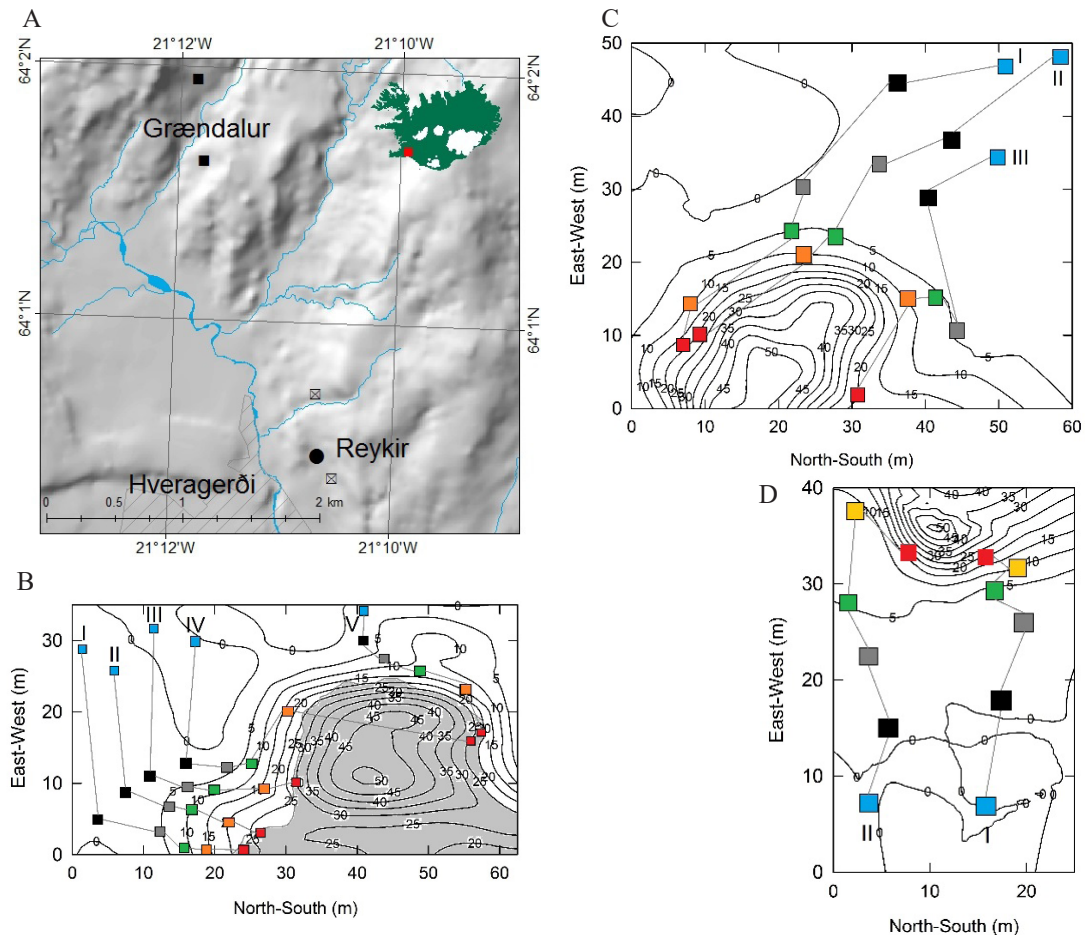


Figure 1. a) Location of the ForHot study sites in S Iceland. Filled circle is the recently warmed forest site (FN), crosshatched squares are the recently warmed grasslands (GN) and the filled squares are the near-by grasslands with long-term soil warming (GO) in Grændalur. The hatched area represents the village of Hveragerði. b) Soil warming isotherms (°C) at five transects within the FN site in spring 2012 and at c) three transects of the GN and d) two transects of the GO sites in spring 2014. Warming levels are A (unwarmed control; blue), B (black), C (grey), D (green), E (orange) and F (red). For evaluation of warming levels see Table 2. In FN, the natural background soil temperature (T_s) varied from -2 to +2 °C because of stand density variation in the unwarmed forest stand and the actual geothermal warming was therefore ca. 2-4 °C lower than the isolines indicate. The shaded area for FN represents the area where Sitka spruce had died off in 2012.

campus at Reykir (Figure 1). Between 2003 and 2015, the closest synoptic station at Eyrabakki (9 km S of Hveragerdi) had a mean annual air temperature (MAT), mean annual precipitation (MAP) and mean wind speeds of +5.2 °C, 1457 mm and 6.6 m s⁻¹, respectively (Icelandic Meteorological Office, 2016). The mean temperature of the warmest and coldest months, July and December, were 12.2 °C and -0.1 °C for the same period. The mean monthly precipitation during May-July was 75 mm month⁻¹, while it was on average 135 mm month⁻¹ for the remaining months at Eyrabakki (Icelandic Meteorological Office, 2016). During the period of 1972-1999 both precipitation and air temperature were measured on the Reykir campus. During that period the MAT was similar (0.1 °C warmer), but the MAP was on average 13% higher at Reykir (1616 mm year⁻¹) than at Eyrabakki (Icelandic Meteorological Office, 2016). The growing season normally starts in late May and ends in late August. Snow cover is not permanent during winters due to the mild oceanic climate, but the soil typically freezes for at least a couple of months during mid-winter.

On the 29 May, 2008, a major earthquake (magnitude 6.3 on the Richter scale) occurred in S Iceland (Halldorsson & Sigbjörnsson 2009). The earthquake caused substantial damage to infrastructures and affected geothermal systems close to its epicentre. One such geothermal system at Reykir moved to a previously unwarmed area (Þorbjörnsson et al. 2009), where the new geothermal bedrock channels resulted in increasing temperature (Ts) in the soil above by radiative heating (O’Gorman et al. 2014).

The recently warmed area is covered by two different site types: a) a Sitka spruce forest (*Picea sitchensis*, provenances Seward and Homer from Alaska) that was planted in 1966-1967 (Böðvar Gudmundsson, pers. comm.), hereafter termed “FN” (Forest New), and b) unmanaged treeless grasslands dominated by *Agrostis capillaris* grass, some herbs and moss (Table 1), hereafter termed “GN” (Grassland New). The soil type at both sites is Silandic Andosols (IUSS Working Group WBR 2015;

a volcanic soil type, also known as Brown Andosol; Arnalds, 2015). It is silty loam in texture and has the typical characteristics of such soils in Iceland (Table 1; Arnalds 2015). The FN plantation was established mainly for shelter and has never been thinned. Therefore, it had a relatively high stand density, basal area and leaf area index (LAI) compared to typical managed spruce forests in Iceland or Scandinavia (Table 1; Snorrason & Einarsson 2002, Weslien et al. 2009).

The third study site “GO” (Grassland Old) is located 2.0-2.5 km NW of GN and FN on older geothermal Ts gradients, in Grændalur (Figure 1). It is covered by the same grassland type as GN and on the same soil type (Table 1). There, the earliest survey of geothermal hot spots was made in 1963-1965 (45 years prior to the 2008 earthquake; Kristján Sæmundsson, pers. comm.). In autumn 2008, after the 2008 earthquake, the locations of the new and old geothermal hot spots in the area were remapped (Þorbjörnsson et al. 2009). This survey was used to choose the GO, GN and FN sites for the ForHot study. Some hot spots at GO have been monitored since 2005 by regular field measurements of Ts in another study in Grændalur (Daebeler et al. 2014, 2015). The geothermal activity has most likely been persistent in Grændalur (Green valley) for centuries, as according to local knowledge its name comes from the fact that the subarctic grasslands on the warmest hot spots remain green during early winter and turn green sooner after the worst of winter has passed. The oldest historical document that mentions this place name was written in 1708 (Magnússon & Vídalín, 1918-1921). Additional evidence for persistent geothermal warming at GO includes the geothermal clay layers found at various depths in the subsoil profile, thus indicating that over longer time periods, the warming may have fluctuated somewhat, as was observed at other nearby hot spots following the 2008 earthquake (Daebeler et al. 2014).

In autumn 2012 and spring 2013, twenty-five permanent study plots were established in each of the three site types (FN, GN and GO), around one main hot spot in FN and in

Table 1. Main plant and soil textural and chemical characteristics of the control plots (A plots) in the recently warmed forest (FN), grassland (GN), and long-term warmed grassland (GO) in 2013. Data from Cilio (2014), Guðmundsdóttir et al. (2014), Michielsen (2014) and Leblans (2016).

| | FN | GN | GO |
|--|--|---|--|
| Soil type | Silandic Andosol | Silandic Andosol | Silandic Andosol |
| Soil texture | Silt loam | Silt loam | Silt loam |
| Clay:Silt:Sand ratio ^a | 8:61:31% | 6:53:41% | 8:62:30% |
| Stoniness in top 10 cm | 2.1% | 1.6% | 0.4% |
| Bulk density in top 10 cm | 0.62 g cm ⁻³ | 0.70 g cm ⁻³ | 0.55 g cm ⁻³ |
| Topsoil C concentration ^b | 7.1% | 5.4% | 5.1% |
| Topsoil N concentration ^b | 0.47% | 0.49% | 0.44% |
| Topsoil C/N ratio ^b | 15.1 | 10.9 | 11.5 |
| Three most dominant vascular plant species | <i>Picea sitchensis</i> Understory: <i>Equisetum arvense</i> – <i>Geranium sylvaticum</i> | <i>Agrostis capillaris</i> – <i>Galium boreale</i> – <i>Anthoxanthum</i> <i>odoratum</i> | <i>Agrostis capillaris</i> – <i>Galium boreale</i> – <i>Ranunculus acris</i> |
| Vascular plant cover ^c | 7% | 46% | 79% |
| Moss cover | 5% | 88% | 62% |
| LAI ^{max} of veg. > 3 cm ^d | 5.4 | 6.0 | 5.8 |
| Dominant height | 10.3 m | - | - |
| Diameter at breast height | 12.6 cm | - | - |
| Stand basal area | 49 m ² ha ⁻¹ | - | - |
| Stand density | 4.461 trees ha ⁻¹ | - | - |

^a The standard methodology used here gives an underestimation for the true clay fraction in Andosol (Arnalds 2015); ^b 5–10 cm depth (only mineral soil); ^c Only including the ground vegetation in FN; ^d Determined by a LAI2200 instrument on 15–20 Sept. 2016.

two separate hot-spots for GN and GO (Figure 1). The plots were placed along five ca. 50 m long transects placed perpendicular to the soil temperature gradients ranging from ambient soil temperature to $\sim +10^{\circ}\text{C}$, placing five replicate plots at different warming levels (WLs) on each transect ($\sim +0$, 1, 3, 5 and 10°C warming; termed levels A (unwarmed control), B, C, D, and E, respectively). In GN and GO the plots were 2×2 m in size, but in FN they were 1×1 m and placed in between trees. In spring 2014 one additional WL was installed at each transect at $\sim +20^{\circ}\text{C}$ (termed level F), but those plots were all 1×1 m in size due to the steeper soil warming gradients at the highest temperatures. This increased the number of permanent study plots to 30 per site type, or 90 across all three.

Both FN and GN transects were in areas which had been previously fenced and protected

from livestock grazing, while Grændalur and its neighbouring areas are used as grazing commons and typically have 10–20 sheep from late May to late August. The areas containing the study plots at GO were therefore fenced off in spring 2013.

Field measurements

Soil temperatures were measured hourly adjacent to each plot at 10 cm soil depth using HOBO TidbiT v2 Water Temperature Data Loggers (Onset Computer Corporation, USA). Air temperature was measured close to the surface (at 2 cm and 15 cm) at two plots at each warming level in each site type and at 2 m height in one to two places in each system, using the same type of loggers and logging frequency, but protected from direct sunlight with radiation shelters. Vertical soil temperature

profiles were measured in late June 2014 by 90 cm long temperature probe placed at different depths (Digi-Sense Type K Thermometer Probe, Oakton Instruments, IL, USA). Volumetric soil water content in the top 0–5 cm of soil was measured during campaigns in 2013 to 2016 by a handheld Theta Probe (Model ML3, Delta-T Devices Ltd., Cambridge, England), with a “mineral soil factory calibration curve” that has been found to give realistic results for Icelandic Andosols in S Iceland (Berglind Orradottir, pers. comm.). Soil depth was measured using a 1 m long metallic rod pushed down until hitting a rock at 11 places along the S edge of each permanent plot, but recorded as 100 cm when deeper. A relative measure for exchangeable sulphur (S) was obtained using exchange membranes (PRSTM probes, Western Ag Innovations Inc., Saskatoon, SK, Canada). The membranes continuously absorb charged ionic species over the burial period, and the S availability is calculated as soil S flux over time. Four sets of membranes were inserted at 0–10 cm depth for 89 days (23 May to 20 August 2013) in each permanent plot. Afterwards, they were sent to Western Ag Innovations Inc. (Saskatoon, SK, Canada) for further analyses. Soil pH in H₂O and 1M KCl was determined from sieved (mesh size 2 mm) soil samples taken from 0–10 cm layers in all permanent plots in July 2014. The samples were dissolved in a 1:2.5 (per mass) solution, shaken for 20 min and shaken shortly again after two hours before measuring pH with a Two Channel Benchtop pH/mV/ISE Meter (Hanna Instruments, Temse, Belgium). Finally, the potential decomposition rate of easily decomposable organic matter was determined with the TBI method (Keuskamp et al. 2013). Four Lipton Green teabags were incubated at 5–7 cm depth in each permanent plot from late May to the middle of September (110 days) in 2014 and then dried at 85 °C for 48 hours, weighed and their mass loss compared to stored control bags.

Statistical analysis

Individual permanent study plots in each site type were used as the unit of replication (n = 25

or 30, without and with the F treatment plots, respectively), except when standard deviations of all observations were calculated. Then all individual measurements were included. One-Way ANOVA was used to test for differences in mean annual values between the three site types, and when significant followed by Fisher’s Least Significant Difference (LSD) pairwise tests (SAS, version 9.4; SAS Institute Inc., Cary, NC, USA). The potential effect of soil warming on different parameters was evaluated with a

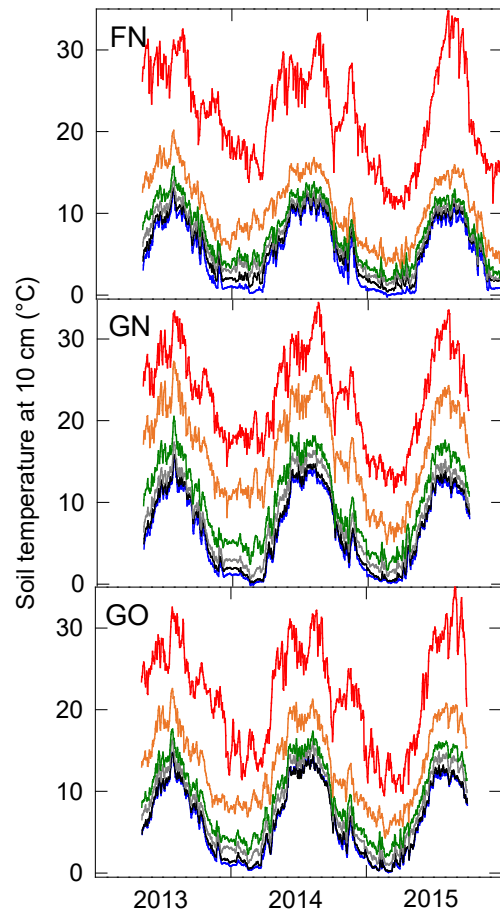


Figure 2. Changes in daily mean soil temperature (Ts) at 10 cm depth in the six warming levels (A–F) of recently warmed forest (FN; top), grassland (GN; middle) and long-term warmed grassland (GO; bottom). Warming levels are A (unwarmed control; dark blue), B (black), C (grey), D (green), E (orange) and F (red). For evaluation of warming levels see Table 2.

Table 2. Mean annual soil temperature at 10 cm depth (MATs, °C; \pm Sd) and mean annual warming (W, °C) in 2014 and the maximum and minimum hourly warming (W_{min}, W_{max} ; °C) relative to average MATs on unwarmed (A) plots and different soil warming levels (B-F). Fraction of hourly MATs data from 2014 within 1, 3, 5 and 10 °C of the annual mean (Variation; \pm 1, \pm 3, \pm 5 and \pm 10 °C), July mean Ts in 2013, 2014, 2015, and January mean soil temperature (Ts) in 2014 and 2015 of the recently warmed forest (FN), grassland (GN), and long-term warmed grassland (GO). All averages and standard deviations are for n=5 plots per warming level (WL).

| | WL | 2014 | | | | Variation (%) in 2014 | | | | July Ts | | | January Ts | |
|----|----|----------------|-------------|------------------|------------------|-----------------------|---------|---------|----------|---------|------|------|------------|------|
| | | MATs | W | W _{min} | W _{max} | \pm 1 | \pm 3 | \pm 5 | \pm 10 | 2013 | 2014 | 2015 | 2014 | 2015 |
| FN | A | 5.3 \pm 0.2 | 0.0 | -0.9 | 1.0 | 99 | 100 | . | . | 10.0 | 10.5 | 9.4 | 1.0 | 0.7 |
| | B | 6.2 \pm 0.5 | 1.0 | -0.2 | 3.3 | 88 | 100 | . | . | 10.6 | 11.0 | 9.9 | 2.0 | 1.6 |
| | C | 7.2 \pm 0.5 | 1.9 | 0.2 | 4.3 | 73 | 99 | 100 | . | 11.6 | 11.5 | 10.5 | 3.6 | 2.3 |
| | D | 8.0 \pm 0.2 | 2.7 | 0.7 | 5.4 | 73 | 96 | 100 | . | 13.1 | 12.4 | 11.5 | 4.5 | 3.2 |
| | E | 11.1 \pm 0.7 | 5.8 | 2.7 | 10.3 | 50 | 86 | 98 | 100 | 17.2 | 15.1 | 14.0 | 7.9 | 5.4 |
| | F | 22.8 \pm 2.2 | 17.5 | 9.4 | 25.0 | 25 | 49 | 85 | 99 | 27.8 | 25.7 | 28.3 | 17.8 | 15.7 |
| GN | A | 6.3 \pm 0.3 | 0.0 | -1.5 | 1.4 | 95 | 100 | . | . | 12.5 | 12.9 | 12.5 | 1.1 | 0.8 |
| | B | 6.8 \pm 0.3 | 0.5 | -1.4 | 1.9 | 96 | 100 | . | . | 12.9 | 13.5 | 13.0 | 1.7 | 1.2 |
| | C | 8.3 \pm 1.4 | 2.1 | -0.2 | 4.1 | 35 | 93 | 100 | . | 14.3 | 15.0 | 14.2 | 2.8 | 2.3 |
| | D | 10.2 \pm 0.3 | 3.9 | 0.3 | 7.4 | 49 | 97 | 100 | . | 16.9 | 16.4 | 16.5 | 5.1 | 3.6 |
| | E | 16.7 \pm 2.5 | 10.5 | 3.7 | 15.0 | 22 | 68 | 97 | 100 | 23.3 | 22.1 | 22.5 | 11.5 | 8.1 |
| | F | 23.6 \pm 1.3 | 17.3 | 8.8 | 24.2 | 30 | 75 | 93 | 100 | 29.4 | 28.3 | 30.5 | 18.2 | 14.6 |
| GO | A | 6.3 \pm 0.6 | 0.0 | -1.8 | 1.9 | 82 | 99 | 100 | . | 11.8 | 12.9 | 11.9 | 1.1 | 1.1 |
| | B | 6.5 \pm 0.6 | 0.2 | -1.8 | 2.6 | 77 | 100 | . | . | 12.0 | 12.7 | 12.4 | 1.6 | 1.2 |
| | C | 7.8 \pm 0.5 | 1.6 | -0.4 | 3.5 | 74 | 100 | . | . | 13.4 | 14.0 | 13.8 | 3.0 | 2.2 |
| | D | 9.1 \pm 0.3 | 2.9 | 0.2 | 5.2 | 71 | 100 | . | . | 14.7 | 15.0 | 15.4 | 4.2 | 3.6 |
| | E | 13.0 \pm 2.0 | 6.8 | 2.7 | 10.4 | 28 | 82 | 99 | 100 | 19.4 | 18.6 | 19.2 | 8.3 | 7.3 |
| | F | 21.9 \pm 3.3 | 15.6 | 6.5 | 25.7 | 19 | 53 | 70 | 95 | 27.7 | 26.3 | 28.7 | 15.6 | 14.2 |

linear regression analysis using mean annual Ts' measured for each plot.

RESULTS

The mean annual soil temperature (MATs) of the unwarmed soil was 6.3 °C in the two grasslands, but 5.3 °C under the dense forest cover in 2014 (Figure 2; Table 2). The July and January Ts' were on average 12.4 °C and 1.0 °C in the unwarmed grassland, but 10.0 °C and 0.9 °C at FN, respectively (Table 2). Frozen soil below 10 cm depth in winter, indicated with Ts being stable around 0 °C, did occur. While this only lasted for a limited time in the unwarmed

grasslands, the unwarmed forest (FN) showed longer periods with frozen soil during winter (Figure 2).

All the warming levels (WLs) were similar during the three years of study and followed the natural Ts seasonal cycle, with relatively stable offset and no periods with frozen soil for the WLs of C to F (Figure 2). We were not successful in placing the permanent plots at exactly the same MATs in each site type, but the average WL across the three site types was +0.6, +1.9, +3.2, +7.7 and +16.8 °C for B, C, D, E and F, respectively (Table 2). Both natural unwarmed Ts and the WLs fluctuated, but for

A, B, C and D the hourly mean Ts was more or less always within ± 3 °C of the annual average WL for each site type. The fluctuation became somewhat larger for E and F, where the hourly mean Ts was within ± 3 °C for 79% and 59% of the time across the three site types (Table

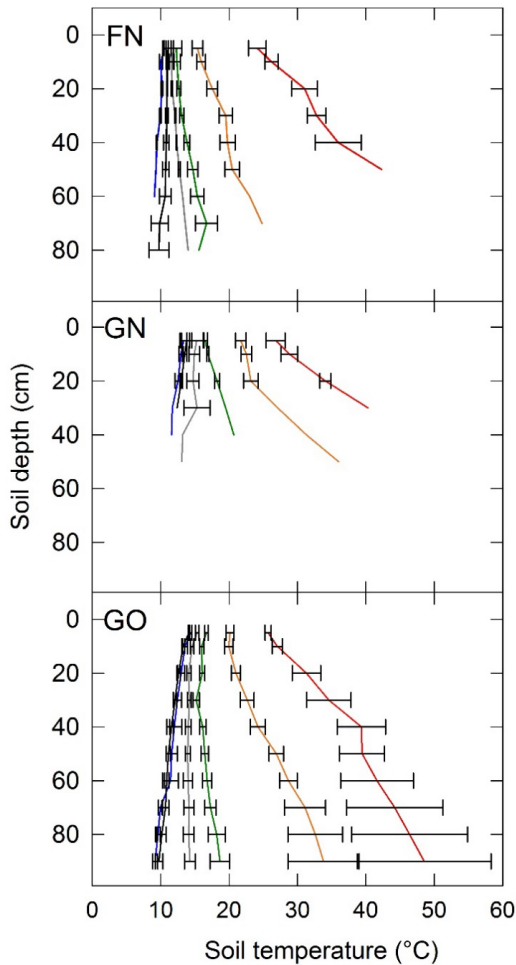


Figure 3. Vertical profiles (5, 10, 20, ... 90 cm) of soil temperature in the six warming levels (A-F) of the recently warmed forest (FN; top) and grassland (GN; middle) and the long-term warmed grassland (GO; bottom) at 27 June 2014. Vertical lines represent loess fits between measured averages and the warming levels are A (unwarmed control; dark blue), B (black), C (grey), D (green), E (orange) and F (red). SEs of Ts measurements from 2-5 plots are shown as lateral bars.

2). The absolute annual peak values for hourly maximum and minimum warming during 2015 for each WL are also shown in Table 2.

A survey measuring vertical Ts profiles at all WLs at all three site types showed that the Ts measured at 10 cm depth were relatively constant in all WLs down to ca. 20-25 cm depth, which represents the most active root layer. Only in the warmest level (F; Figure 3), did the Ts increase more rapidly with depth and there the Ts, which were measured at 10 cm depth, would clearly underestimate the soil warming in the 15-30 cm layer. The spatial variability among the five replicated plots within each WL increased with Ts (Figure 3), indicating

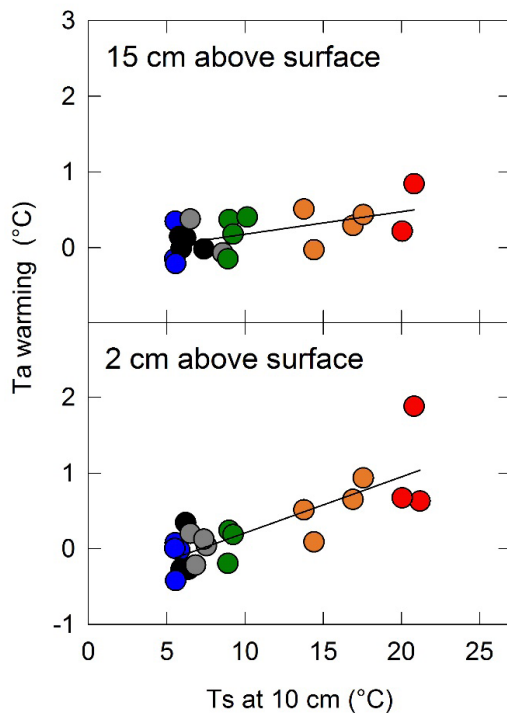


Figure 4. The relationship between mean soil temperature (Ts) at 10 cm depth and average air temperature (Ta) warming compared to Ta measured at 2 m height for the period 15 February to 30 June 2015 in the ForHot grasslands. Warming levels are A (unwarmed control; dark blue), B (black), C (grey), D (green), E (orange) and F (red). The lines represent significant regression relationships.

that it became increasingly difficult to place all plots on exactly similar Ts' as the geothermal warming gradients became steeper. Figure 3 illustrates the differences in soil depth between the three site types, where GN had significantly shallower soils than both GO and FN, and GO had significantly the deepest soils (Table 3).

Within each site type there was, however, no systematic difference in soil depth across the WLs (Table 3; regressions not significant). Measurements of air temperature (Ta) at 2 m height at different places in FN showed no effects of the soil warming (data not shown) and therefore Ta at 2 m was only measured at one

Table 3. Mean soil depth, pH_{H_2O} , ΔpH_{KCl-H_2O} , exchangeable sulphur, and potential decomposition rate for unwarmed (A) plots and different soil warming levels (B-F) in the recently warmed forest (FN), grassland (GN), and long-term warmed grassland (GO). Significant differences between site-type means are indicated by different letters (post-ANOVA/LSD tests). The lower part of the table shows regression analyses of the effect of soil temperature on the different variables. *** = $P < 0.001$, ** = $P < 0.01$, * = $P < 0.05$, (ns) = $P < 0.10$, ns = $P > 0.10$. Significant changes across Ts levels are indicated in bold.

| Site type | Soil depth (cm) | | | $pH_{H_2O, 0-10\text{ cm}}$ | | | $\Delta pH_{KCl-H_2O, 0-10\text{ cm}}$ | | | Exchangeable S ^a | | | Potential decomposition ^b | | |
|-----------------------------|-----------------|------|------|-----------------------------|--------------|-------|--|--------------|--------------|-----------------------------|-------|-------|--------------------------------------|---------------|---------------|
| | FN | GN | GO | FN | GN | GO | FN | GN | GO | FN | GN | GO | FN | GN | GO |
| A | 76.3 | 54.1 | 98.2 | 5.3 | 5.6 | 5.7 | -1.35 | -1.49 | -1.53 | 51.6 | 77.7 | 222.6 | 47 | 53 | 54 |
| B | 72.4 | 38.6 | 91.5 | 5.1 | 5.8 | 5.7 | -1.19 | -1.63 | -1.52 | 108.7 | 89.1 | 123.0 | 45 | 58 | 55 |
| C | 68.1 | 42.7 | 75.0 | 5.1 | 5.8 | 5.5 | -1.25 | -1.62 | -1.56 | 102.4 | 78.2 | 142.4 | 51 | 58 | 55 |
| D | 68.6 | 28.7 | 90.5 | 5.0 | 5.9 | 5.7 | -1.13 | -1.56 | -1.62 | 73.5 | 56.5 | 177.7 | 49 | 56 | 59 |
| E | 69.3 | 27.4 | 82.9 | 5.2 | 6.3 | 6.1 | -1.25 | -1.95 | -1.73 | 87.2 | 61.7 | 197.3 | 52 | 69 | 58 |
| F | - | - | - | 5.7 | 6.3 | 6.1 | -1.81 | -2.00 | -2.02 | - | - | - | 64 | 69 | 61 |
| Mean | 70.9 | 38.3 | 87.6 | 5.25 | 5.95 | 5.82 | -1.33 | -1.71 | -1.66 | 84.7 | 72.6 | 172.6 | 51 | 61 | 57 |
| LSD | a | b | c | a | b | b | a | b | b | a | a | b | a | b | c |
| Regression analysis (ANOVA) | | | | | | | | | | | | | | | |
| Intercept | 79.8 | 60.1 | 97.6 | 4.905 | 5.524 | 5.532 | -0.99 | -1.39 | -1.42 | 74.8 | 96.47 | 97.57 | 0.415 | 0.491 | 0.527 |
| Slope | -1.2 | -2.2 | -1.2 | 0.034 | 0.036 | 0.026 | -0.03 | -0.02 | -0.02 | 1.30 | 8.90 | -1.16 | 0.0098 | 0.0096 | 0.0039 |
| n | 25 | 25 | 25 | 30 | 30 | 30 | 30 | 30 | 30 | 25 | 25 | 25 | 30 | 30 | 30 |
| r ² | 0.02 | 0.16 | 0.02 | 0.22 | 0.37 | 0.06 | 0.36 | 0.42 | 0.19 | 0.01 | 0.02 | 0.02 | 0.64 | 0.73 | 0.22 |
| P | ns | (ns) | ns | * | *** | ns | *** | *** | * | Ns | ns | ns | *** | *** | ** |

^a Unit: $\mu\text{g S } 10\text{ cm}^{-2}$ 89 summer days⁻¹ ^b = Relative mass loss after 110 days of incubation (spring-autumn; %)

location in GN and GO. At 2 and 15 cm height above the surface in GN and GO the regression relationships between average T_s ($^{\circ}\text{C}$) and average T_a warming (ΔT_a ; $^{\circ}\text{C}$) were significant (Figure 4; $\Delta T_{a_{15}}$: $r^2 = 0.32$; $P = 0.01$; $\Delta T_{a_{02}}$: $r^2 = 0.65$; $P < 0.001$):

$$\Delta T_{a_{15}} = 0.03 \times T_s - 0.12, \quad (1)$$

$$\Delta T_{a_{02}} = 0.07 \times T_s - 0.53. \quad (2)$$

The above relationships indicate that air warming was substantially less than the soil warming, and a 20°C increase in T_s only elevated the average T_a by 0.95°C and 0.48°C at 2 cm and 15 cm height above the surface, respectively (Figure 4).

Measurements of $\text{pH}_{\text{H}_2\text{O}}$ of the top 10 cm of mineral soil did not show any indications that geothermal water had reached the root zone in any of the sites (Table 3). In the unwarmed soil, the pH was significantly lower in the coniferous forest plantation than in the two grasslands, which did not differ (pH 5.2 in FN vs. ca. 5.9 in GN and GO). Soil pH did, however, increase somewhat with T_s and this change was significant in the recently warmed sites (FN and GN), but not in the long-term warmed site (GO). $\Delta \text{pH}_{\text{KCl-H}_2\text{O}}$, which was also significantly lower in FN than in the two grasslands, increased significantly with WL in all three site types (Table 3). A lack of geothermal contamination was further supported by the lack of significant regressions between T_s and exchangeable sulphur (S) in the soil in any of the site types (Table 3). The overall level of exchangeable S was, however, higher in GO, where a higher number of geothermal vents were found within the same valley than in the recently warmed sites.

All site types had the highest volumetric water content (WC) early in the spring and the lowest measured surface WCs in the middle of summer (Figure 5). The average surface WCs of the unwarmed control treatments (WC_{meanA}) during the period of 11 April to 1 September 2016 were 52.2%, 38.8% and 31.1% in GO, GN and FN, respectively (Figure 5), and similar values were obtained in 2013–2015 (Table 4). When

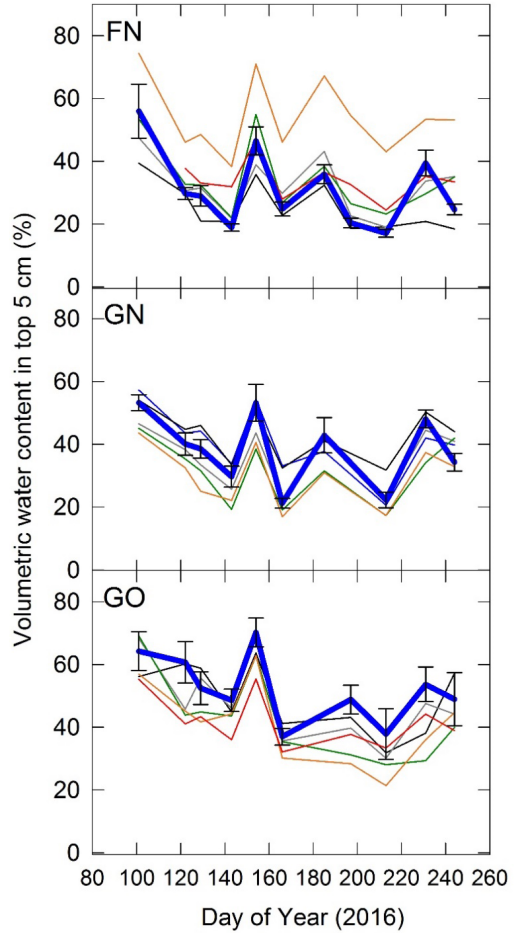


Figure 5. Changes in mean volumetric water content (WC) in the top 5 cm in the six warming levels (A–F) of the recently warmed forest (FN; top) and grassland (GN; middle) and the long-term warmed grassland (GO; bottom) from April to August 2016. Warming levels are A (unwarmed control; dark blue, thicker line, including SE of the mean), B (black), C (grey), D (green), E (orange) and F (red).

the plot-level WCs for individual measurement days were compared across the soil warming gradients in each site type, a significant linear drying effect with warming was observed for the surface soil for a part of the dates in both GN and GO, but only a few times in FN, which was the driest site (regression analysis; data not shown).

Table 4. Mean volumetric soil water content (%) in the top 5 cm in unwarmed soil (WC_{meanA}) and across all soil temperature (Ts) levels (WC_{mean}) and the corresponding standard deviation (WC_{Sd}) in 2013 to 2016 in the recently warmed forest (FN), grassland (GN), and long-term warmed grassland (GO). The lower part of the table shows regression analyses of the effect of mean soil temperature on the mean seasonal plot-wise WC. Int. = intercept, Sl. = slope. *** = $P < 0.001$, ** $P < 0.01$, * = $P < 0.05$, (ns) = $P < 0.10$, ns = $P > 0.10$. Significant seasonal increases or reductions in WC with Ts are indicated in bold.

| Site type | 2013 | | | 2014 | | | 2015 | | | 2016 | | |
|-----------------------------|--------------|--------------|-------|-------|--------------|--------------|-------|--------------|--------------|-------|--------------|--------------|
| | FN | GN | GO | FN | GN | GO | FN | GN | GO | FN | GN | GO |
| Campaigns | 4 | 4 | 4 | 12 | 8 | 8 | 4 | 6 | 6 | 11 | 10 | 10 |
| WC _{meanA} (%) | 31.2 | 38.3 | 36.4 | 33.3 | 49.4 | 53.5 | 28.6 | 46.4 | 54.0 | 31.1 | 38.4 | 52.3 |
| WC _{mean} (%) | 33.9 | 36.6 | 33.9 | 44.4 | 47.9 | 52.1 | 37.5 | 42.4 | 51.5 | 35.5 | 36.6 | 46.0 |
| WC _{Sd} (%) | 9.0 | 7.3 | 9.0 | 15.3 | 13.7 | 14.1 | 15.6 | 13.5 | 14.9 | 15.9 | 13.6 | 15.7 |
| Regression analysis (ANOVA) | | | | | | | | | | | | |
| Int. (%) | 15.25 | 43.8 | 37.89 | 39.4 | 56.7 | 60.0 | 30.6 | 51.6 | 58.4 | 31.8 | 44.9 | 52.0 |
| Sl. (% °C ⁻¹) | +2.46 | -0.75 | -0.20 | +0.50 | -0.74 | -0.73 | +0.68 | -0.77 | -0.65 | +0.33 | -0.71 | -0.56 |
| n | 25 | 25 | 25 | 30 | 30 | 30 | 30 | 30 | 30 | 30 | 30 | 30 |
| r ² | 0.61 | 0.40 | 0.02 | 0.10 | 0.44 | 0.22 | 0.12 | 0.57 | 0.24 | 0.03 | 0.37 | 0.15 |
| P | *** | *** | ns | (ns) | *** | ** | (ns) | *** | ** | ns | *** | * |

During 2013-2016, the surface WC did not change significantly (2014-2016) or it was even significantly increased with increasing Ts (2013) in FN, when averaged over the whole growing season (ranging between +0.33 to +2.46%WC °C⁻¹; Table 4). The increase in WC at FN was especially pronounced at warming level E. In GN, however, the seasonal mean surface WC was always significantly reduced across the Ts gradients (-0.71 to -0.77%WC °C⁻¹), while in GO it shifted between no significant change to significantly reduced (-0.20 to -0.73%WC °C⁻¹; Figure 5; Table 4). It should, however, be noted that the observed surface WCs in the warmest treatments of GO were still substantially higher than the unwarmed control FN soil (e.g. 41.8% vs. 31.1% WC in 2016). The warmest GN plots were usually similar to the unwarmed FN soil, on average (30.0% vs. 31.1% WC, respectively, in 2016; Figure 5).

To indicate how ecosystem processes responded to the soil warming, the potential decomposition rate of organic matter is shown (Table 3). Green tea decomposed faster in the two grasslands than in the forest site in unwarmed soil. The soil warming, significantly

increased the decomposition rate in all three site types, but to a different degree. The slope of the temperature response was almost identical for the two recently warmed sites (FN and GN), but it was ca. 60% lower in the long-term warmed grassland. The net result was therefore that the observed mean decomposition potential across all WLs was significantly highest in GN, second in GO, and significantly lowest in FN (Table 3).

DISCUSSION

Realism of the soil warming

The main concern when starting the ForHot project was whether geothermal water, which is rich in various dissolved minerals, reached the rooting zone in the soils and thus created difficult conditions for many organisms (Wetang'ula & Snorrason 2005). If this was the case, observed changes in processes and ecosystem structure might have been driven by chemical factors rather than by temperature changes. Our results for pH and exchangeable S showed no indication of contamination by geothermal water in any of the ForHot sites. The lack of contamination was further confirmed in a recent thesis on seasonal soil water chemistry below the root zone at the

FN site (Edlinger 2016). A second concern was that the natural seasonal patterns of Ts would flatten out at the warmed plots. This was not the case, as can clearly be seen in Figure 2. Lastly, we also expected the vertical Ts gradients in geothermally affected plots to become very different from the ones in unwarmed conditions, i.e. that Ts would increase dramatically with soil depth. The difference was, however, marginal (except for the warmest F plots), especially for the top 20–30 cm of soil containing most of plant roots.

All the above findings support the hypothesis that the ForHot project geothermal gradients offer conditions similar to, for example, manipulation studies with soil heating cables for studying the effects of soil warming (Rustad 2001, Strömberg & Linder 2002, Bronson & Gower 2010, Patil et al. 2013, Streit et al. 2013, Krassovski et al. 2015, Schindlbacher et al. 2015).

Stability of the soil warming

It is neither expensive nor difficult to maintain the soil warming in the ForHot sites, but since the warming is entirely passive it cannot be controlled. During the study period of 2013–2015 the geothermal gradients remained relatively stable at all three sites. Geothermal systems, however, tend to be dynamic in nature (Carotenuto et al. 2016), as was also witnessed in the present study by the creation of the new geothermal gradients during the large 2008 earthquake (Þorbjörnsson et al. 2009). This complicates the estimation of the duration of unchanged warming in Grændalur (GO). Based on the survey in autumn 2008 by Þorbjörnsson et al. (2009), showing little change in distribution of hotspots and geothermal vents in those areas (especially for transects 1–4 of GO), and Kristjánsson's first mapping of the geothermal hotspots in the same area during 1963–1965, it was assumed that they have existed for at least 50 years. However, even though Grændalur has had these geothermal hot spots for centuries, they may have changed both spatially and thermally over time. To further study the temporal and spatial history of the geothermal warming at

GO, there is an ongoing activity using HPLC (High Performance Liquid Chromatography) analysis to study recalcitrant soil bacteria lipids in soil profiles from GO, which may be used to reconstruct historical soil temperature (De Jonge, pers. comm.). Such measurements have been successfully used as paleo-climate proxies of soil temperature elsewhere (De Jonge et al. 2014).

Warming-induced drying

A general concern with all terrestrial warming experiments is their effect on soil and plant water status, since their warming treatments will inevitably increase evapotranspiration and therefore potentially induce drought, which could confound the “warming responses” (Lu et al. 2012, De Boeck et al. 2015). Whether such drying will induce some strong biological or biogeochemical responses depends largely on the hydrological conditions of each site. The annual precipitation tends to be high at Reykir (between 1134 and 2023 mm in 1972–2000; Icelandic Meteorological Office, 2016) and due to the relatively short summer and the high water storage capacity of the sites' soil type (Arnalds 2015), drought was not expected to be a major driver at the ForHot sites.

The driest surface soils were in the FN forest, where the dense forest stand likely had both higher transpiration and much higher evaporation from intercepted rainfall than the grasslands (cf. Koivusalo et al. 2006). However, no signs of water stress were observed when the stomatal conductance of the Sitka spruce trees was measured (André & Bondesson 2014). Further, no significant drying was observed at higher WLs in FN. On the contrary, there was a strong “wetting”, especially at ca. +10 °C soil warming. This occurred exactly at the interface where warming-induced tree mortality was taking place (O’Gorman et al. 2014), and where the stand had thinned, but where only a little ground vegetation had colonized (Gudmundsdottir et al. 2014, Cilio 2014). At the F plots, all Sitka spruce trees had died and the ground was covered with lush herbaceous vegetation (Gudmundsdottir et al. 2014). We

interpret this wetting as a relatively higher reduction in tree transpiration and intercepted rainfall in the warmest treatments in FN compared to the warming-induced drying.

The grasslands did, however, show a significant over-all drying of the surface layer with warming in most years, especially GN, which had thinner soils and therefore less water storage potential. This drying was not always significant and apparently became stronger during and just after the infrequent dry spells in mid- to late summer, when the water content of the unwarmed control plots was also reduced. Such drying of surface soil, however, may not necessarily cause physiological stress for plants, especially not those that have roots that extend below 5 cm. In fact, the soil water status of the “driest” treatments in the grasslands was still similar to the unwarmed control plots in FN. Further, when stomatal conductance (g_s) was measured for *Ranunculus acris*, *Poa pratensis* and *Agrostis capillaris* in GO and GN, no significant (drought-induced) stomatal closure was observed across the WLs (Michielsen 2014). The more shallow-rooted *Poa* and *Agrostis* grasses had, however, on average significantly lower g_s in GN than in GO, which fits well with our measurements of surface water content.

Soil warming vs. air warming

All soil warming techniques have only limited effects on air temperature (Lu et al. 2012, Streit et al. 2013). Patil et al. (2013) observed only 0.20 °C warming at 10 cm height above the soil surface in an experiment where T_s was elevated by 5.0 °C in agricultural soils by buried heating cables at 10 cm depth. By applying Eq. 1 we found that the average air warming at 15 cm height was very similar in our case where the T_s was exactly 5 °C warmer, or 0.15 °C warmer. The surface air warming was stronger, but still it was an order of magnitude smaller than the warming at 10 cm soil depth. This calls for some caution when findings of aboveground processes measured in soil-warming experiments are extrapolated in relation to future climate warming. An example is the

effect of warming on tree canopy gas exchange which was measured at ca. 8 m height above the soil surface (André & Bondesson 2014). There the effects of the soil warming are interesting in terms of better understanding the potential effects soil and root temperatures can have on such aboveground processes, but it is highly doubtful that the responses could be used to predict photosynthesis or transpiration of spruce in a future climate. This is a general issue with soil warming experiments, and ForHot is neither better nor worse in that respect than other recent or ongoing soil warming manipulation studies (Patil et al. 2013, Streit et al. 2013).

Effects on potential decomposition

At higher latitudes (>60 °N), soils store the largest part of the total ecosystem organic matter (Scharlemann et al. 2014). This is also the case for the Andosols of Iceland (Arnalds 2015). Understanding the dynamics of organic matter (C) pools, including turnover of litter (Davidson & Janssens 2006), roots (Leppälammi-Kujansuu et al. 2014) and biota in soil (Clemmensen et al. 2013) is therefore key to making sound predictions of future ecosystem C balance under changing climatic conditions. In this paper, which has mainly focused on methodological issues, we only show warming effects on one such process; the potential decomposition rate of easily decomposable organic material. As expected, the potential decomposition rate was significantly enhanced with increasing T_s in all three site types. Interestingly, however, the slope of the temperature response was much lower in GO, than in GN and FN. This might indicate some acclimation in the GO grasslands in terms of the rate of organic matter breakdown. A recent review of Lu et al. (2012) found indications that the temperature sensitivity of microbes declines at higher temperature levels in manipulation experiments or that the microbes themselves may acclimatize to high temperatures. Further studies on the different aspects of the C-cycle of the ForHot grassland sites are being conducted (Leblans 2016, Poeplau et al. 2016).

Other observed effects of warming

The soil $\text{pH}_{\text{H}_2\text{O}}$ increased slightly, but significantly so, with warming in both GN and FN, but not in GO. This might have been driven by a decrease in vascular plant root-litter inputs at the higher temperatures (Way & Oren 2010) and/or a gradual depletion of partly decomposed humic materials in the warmer soils of FN and GN after the warming was initiated in 2008. The reverse process, i.e. a gradual decrease of $\text{pH}_{\text{H}_2\text{O}}$ as vegetation cover or productivity increases, is well known in Icelandic grassland and forest soils (Sigurðsson et al. 2005, Sigurðsson & Magnusson 2010, Vilmundardóttir et al. 2015).

Another interesting finding was the gradually increased $\Delta\text{pH}_{\text{KCl-H}_2\text{O}}$ with warming in all three site types. This so-called “reserve acidity” is generally caused by additional Al^{3+} and H^+ being released from exchange sites in the soil (Arnalds 2015). This increase may therefore indicate certain soil structural change due to warming, which increased the amount of such exchange sites being exposed. Indeed, a warming-induced change in soil structure and breakdown of soil aggregates has recently been found to be an important driver for the observed alterations in soil organic matter at GN, apart from the increasing decomposition potential caused by higher temperatures (Poepplau et al. 2016). This is a very interesting finding and may change our understanding of how climate change is likely to affect soil organic matter.

CONCLUSIONS

We conclude that the large natural geothermal gradients at the ForHot sites in Iceland have offered realistic conditions to study terrestrial ecosystem responses to warming with minimal artefacts. This conclusion was supported by the findings that the soil warming was relatively stable over multiple years and the seasonal patterns of T_s in the warmed plots closely followed the unwarmed plots. In this respect the geothermal warming simulated the natural conditions better than expected. Also, it was found that, even if other potential environmental drivers such as surface soil water content and

soil pH were not insensitive to the warming, their changes were relatively mild and should not have overridden the direct warming effects on biological processes.

A certain caution must, however, be used when findings from such geothermal gradients are interpreted in relation to the possible large-scale effects of future climate warming: i) Such gradients are only found in the volcanic areas of the world and various soil and ecosystem conditions there may differ from other areas. ii) Since geothermal gradients mainly warm the belowground parts of the ecosystem, they may underrepresent responses in aboveground processes that are more controlled by air temperature. iii) The stepwise increase in soil temperature at the initiation of a geothermal gradient may cause some differences in short-term responses to warming compared to gradual climate warming. The last two issues are also a general problem with all ongoing soil warming manipulation experiments.

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