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Reference:

Alves Andreia, Vanermen Guido, Covaci Adrian, Voorspoels Stefan.- Ultrasound assisted extraction combined with dispersive liquid-liquid microextraction (US-DLLME)-a fast new approach to measure phthalate metabolites in nails
Analytical and bioanalytical chemistry - ISSN 1618-2642 - 408:22(2016), p. 6169-6180
Full text (Publisher's DOI): <http://dx.doi.org/doi:10.1007/S00216-016-9727-1>
To cite this reference: <http://hdl.handle.net/10067/1357040151162165141>

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2 Ultrasonid Extraction Combined with Dispersive

3 Liquid-Liquid microextraction (US-DLLME) for

4 biomonitoring of phthalate metabolites in nails

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22 **Abstract**

23 A new, fast and environmental friendly approach, ultrasound extraction combined with
24 dispersive liquid-liquid microextraction (US-DLLME) was developed for assessing the levels
25 of seven phthalate metabolites (including the mono(ethyl hexyl) phthalate (MEHP), mono(2-
26 ethyl-5-hydroxyhexyl) phthalate (5-OH-MEHP), mono(2-ethyl-5-oxohexyl) phthalate (5-oxo-
27 MEHP), mono-*n*-butyl phthalate (MnBP), mono-isobutyl phthalate (MiBP), monoethyl
28 phthalate (MEP) and mono-benzyl phthalate (MBzP)) in human nails by UPLC-MS/MS. The
29 optimization of the US-DLLME method was developed according to the Taguchi
30 combinatorial design (L_9 array). Several parameters such as extraction solvent nature,
31 extraction solvent volume, extraction time, acid media, acid concentration and vortex time
32 were studied. The optimal extraction conditions achieved were 180 μL of trichloroethylene
33 (extraction solvent), 2 mL trifluoroacetic acid in methanol (2M), 2 h extraction and 3 min
34 vortex. Under the most favorable conditions, good intra-day (RSD % between 6 % and 14 %)
35 and inter-day (RSD % between 2 % and 17 %) precisions were obtained. The accuracy ranged
36 from 79 % to 108 % and the limit of method quantification (LOQ_m) was below 14 ng/g for all
37 compounds. The developed US-DLLME method was applied to determine the target
38 metabolites in ten Belgian individuals. Average levels of the analytes measured in nails
39 ranged between 3.6 ng/g and 918 ng/g. The MEHP, MBP isomers and MEP are the major
40 metabolites (100 % detection). Despite the miniaturization (low volumes of organic solvents),
41 low costs and simplicity on extraction are main advantages of the proposed US-DLLME.

42

43 **Keywords:** nails, US-DLLME, UPLC-MS/MS, biomonitoring, non-invasive, phthalate
44 metabolites

46 **1. Introduction**

47 Since 1920s, the phthalate esters (PEs) have been used worldwide in several consumer
48 products (e.g. personal care products, perfumes, upholstery, food containers, toys, medical
49 devices, insects repellents, antifoaming agents, building products, etc.) mainly as plasticizers,
50 additives, emollients, humectants or sealants [1–6]. PEs are composed by a group of aromatic
51 chemicals containing a phenyl ring with two attached and extended ester groups (i.e., alkyl
52 chain). Depending on the length of the alkyl chain, PEs can be classified as high molecular
53 weight (> 7 carbons) or low molecular weight (3 to 6 carbons) compounds [1, 7]. In addition,
54 toxicity of PEs is associated with the length and branching of alkyl side-chain [8, 9], however
55 several studies reported that the high molecular weight PEs are quickly metabolized and excreted
56 than low molecular weight (short chain), therefore high molecular PEs have rather low
57 bioaccumulation in human body [10, 11]. Yet, the PEs (and in some cases their metabolites) are
58 considered toxic for animals and/or humans, where the most common effects include
59 hepatotoxicity, carcinogenicity, endocrine disruption, fertility decrease (reproductive system), low
60 birth weight and decreased anogenital distance in children among other effects [5, 12-15].

61 However, the ubiquitous presence of PEs combined with the unexpected external
62 contamination (e.g. for food during processing, handling, packaging or storage) makes the
63 identification of the main routes (ingestion, dermal contact, hand-mouth contact) and factors
64 (food, water, environment) that contribute for the total human exposure much more difficult [1,
65 5]. Also, it is known that PEs are constantly released into the environment by direct release,
66 migration, evaporation, leaching and abrasion of and from the products they are used in,
67 therefore they can be detected in different types of environmental samples such as indoor dust
68 [16, 17] and/or (personal) ambient air [3].

69 Still, after all (preventive) regulation aspects imposed by law for controlling and/or reducing
70 human exposure to PEs [18-22], there has been an increasing attention for human biomonitoring
71 (HBM) and in particular for these substances and/or metabolites in target populations (e.g.
72 pregnant/breastfeeding women, young/adolescents, etc.) [3, 6, 23, 24].

73 Thus, due to the fast metabolism (short half-life) of PEs in human body, urine is seen as a
74 great matrix for assessing PEs, where the measurements in this matrix are considerable easy and
75 fast [2, 4, 6, 7, 12, 23, 27-30]. Nevertheless, the dilution effect of certain metabolites even after
76 creatinine-adjustment in urine (e.g. mono (ethyl hexyl) phthalate, MEHP) [1, 4] can be a
77 considerable disadvantage [30] when a study comparison within participants and between
78 matrices (blood, plasma) is needed. Thus, to overcome this gap alternative non-invasive samples
79 like hair [31], saliva [12, 32] or nails (explored in this study), seems to be promising alternatives

80 in HBM studies [1], although not yet very explored due to the high frequency of non-detects or
81 low capacity to translate significant internal exposure (> LOQ).

82 In fact, the major advantages of non-invasive over the invasive samples are remarked by the
83 broader detection time window for human exposure (days to years) and by the easier, faster
84 sampling and handling, long term storage and high stability [1].

85 Another point of concern which has been slightly overlooked is the lack of novel
86 extraction/analytical methods able to detect PE metabolites at trace levels in biological samples.
87 Most of the methods presented in the current literature [2, 6, 23, 28, 29, 31] for assessing PE
88 metabolites either in liquid or solid matrices, involve a considerable time-consuming, tedious
89 and an expensive approach based on solid phase extraction (SPE) or liquid- liquid extraction
90 (LLE) followed by liquid chromatography tandem mass spectrometry (LC-MS/MS) analysis.

91 From our knowledge, the miniaturization of sample preparation and/or extraction can lead to
92 reach the low (expected) detection levels in ng/g or pg/ μ L range, where there is a little even
93 foreseen and applied to the solid non-invasive samples such as nails. The main feature of
94 DLLME is the miniaturization of the extraction, leading therefore to enormous advantages
95 including fastness, easiness, use of an environmental friendly approach, less waste of organic
96 (and toxic) solvents, the use of low extraction solvent volumes (μ L), combination of ultra-
97 concentration and clean-up, low costs, simplicity of operation [33]. Due to these advantages,
98 DLLME and/or its combination with other techniques (solid phase extraction, ultrasonication,
99 solidification of single drop, for example) has gained an increasingly importance in distinct
100 fields of application (environmental, food, pharmaceutical, inorganic, etc.) [34-38]. In particular,
101 DLLME-based techniques have been also widely used to determine phthalate esters in water,
102 wine, food, cosmetics, etc [39-42]. In addition, the use of microextraction techniques for
103 assessing their metabolites seems very rare. So far, only one study by Sun et al. (2013) [43] that
104 assessed the DEHP metabolites in urine using a DLLME-based approach was disclosed.

105 Thus, in the present study we proposed the use of the dispersive liquid-liquid microextraction
106 (DLLME) in combination with an ultrasound extraction (US) for assessing low ng/g levels of
107 seven PE metabolites (including, MEHP, 5-oxo-MEHP, 5-OH-MEHP, mono-*n*-butyl phthalate
108 (MnBP), mono-isobutyl phthalate (MiBP), monoethyl phthalate (MEP) and mono-benzyl
109 phthalate (MBzP)) in human nails. For the first time, the fast and more environmental friendly
110 US-DLLME approach based on a combinatorial Taguchi design is developed.

111

112 **2. Experimental section**

113 **2.1 Chemicals, material and Equipment**

114 The neat phthalate metabolites including MEHP (99.9 %), MiBP (97.8 %), MnBP (97.4 %),
115 MBzP (99.8 %) and MEP (100 %) were supplied from Accustandard Inc. (Connecticut, USA).
116 The neat MEHP oxidative metabolites (5-oxo-MEHP (>95 %) and 5-OH-MEHP (>96.5 %))
117 were provided by BCP instruments (Irigny, France). Mass-labelled internal standard (IS)
118 solutions for all metabolites (^{13}C -MEHP, ^{13}C -5-oxo-MEHP, ^{13}C -5-OH-MEHP, ^4d -MiBP, ^{13}C -
119 MnBP, ^{13}C -MBzP and ^{13}C -MEP) were obtained from Cambridge Isotope Laboratories
120 (Andover, USA) (purity 95 %). Individual stock standard solutions of PE metabolites and IS
121 were prepared on a weight basis in acetonitrile for final concentrations approximately of 20
122 $\mu\text{g/mL}$. Working standard solutions of both native and IS were prepared separately by mixing
123 the individual standards in Ultra Performance Liquid Chromatography (UPLC)-grade
124 acetonitrile. The spike solutions (PEs and IS) and the standard used for the standard calibration
125 curves were prepared in ultra-pure water by appropriated dilution of the working solutions.

126 The organic solvents such as trichloroethylene, chloroform, toluene, nitric acid (65% purity)
127 and the acetic acid (100 %) were all purchased from Merck (Darmstadt, Germany). Acetone,
128 acetonitrile (ACN) and methanol (MeOH) by UPLC-grade were purchased from Fisher
129 Scientific (Loughborough, UK). Trifluoroacetic acid (TFA, 99% purity) and hydrochloric acid
130 (HCl, 37% purity) were supplied by Sigma-Aldrich (Steinheim, Germany).

131 The ultra-pure water was provided by the Millipore S.A (Advantage A10 system, Overijse,
132 Belgium). Before use, all glassware was prior washed and heated in the oven at 450°C
133 (overnight) in order to reduce the environmental contamination deposited in the glass surface.
134 Afterwards, the materials were stored and protected with aluminium foil against the contact
135 with air and dust. The use (and contact) with plastic materials was minimized or eliminated
136 (e.g. septa and caps of centrifuge vials were covered with aluminum foil prior use).

137

138 2.2 Samples collection and preparation

139 Nail samples were collected among volunteers in our institute (VITO) during 2014. All
140 participants were duly informed about the purpose of this study giving their consent to
141 participate (register N. ° B300201316329). Before the extraction, all nails were rinsed twice
142 with acetone in order to remove dust particles and some residues of polish material. Then the
143 samples were grinded and homogenized in a mixer mill (30 min, 20 Hz) and stored afterwards
144 in a glass vial. For the method development, a pool sample was prepared mixing nails from
145 different individuals, since the amount collected from an single participant was not enough for
146 performing the assigned experiments (development and validation). After validation of the US-

147 DLLME, the evaluation of method and its applicability was performed in Belgian individuals
148 (n= 10).

149 For analysis, 30 mg of powder nails were weighted into glass vials and spiked with IS (5-6
150 ng) (and PE metabolites, i.e., 8 ng of MEP, 9 ng of 5-oxo-MEHP and MBzP, 10 ng of MEHP
151 and 5-OH-MEHP and 18 ng of MBP isomers in the method development and validation).

152

153 2.3 UPLC-MS/MS analysis

154 The instrumental analysis was performed by Ultra Performance Liquid Chromatography
155 (UPLC) coupled to tandem mass spectrometer (Quattro Premier XE Micromass) from Waters
156 (Milford, MA, USA). The chromatographic separation was performed in an Acquity UPLC
157 BEH phenyl column (100 mm x 2.1 mm I.D; 1.7 µm particle size) with a Van Guard Acquity
158 UPLC BEH C₁₈ pre-column (5 mm x 2.1 mm I.D; 1.7 µm particle size). The column
159 temperature was kept at 40 °C.

160 Optimal separation was obtained using a gradient elution program: water (solvent A) and
161 ACN (solvent B), both acidified with 0.1 % acetic acid (v/v). The gradient elution program was:
162 0-0.5 min: 95 % A; 0.5-8 min: 95-10 % A; 8-8.5 min: 10-95 % A (return to initial conditions);
163 8.5-10 min: equilibration of the column. The mobile phase was discarded in the first 3 min to
164 avoid the interference of proteins or other hydrophilic substances into the analyte's ionization.
165 The flow rate was set in 0.4 mL/min and the injection volume was 10 µL. The total run time
166 was 10.5 min.

167 The UPLC system was coupled to the mass spectrometer (MS) interfaced with an
168 electrospray ionisation source operating in negative ionization mode (ESI-). The MS operation
169 conditions were as follows: the cone voltage was set at 25 V. The collision cell energy varied
170 between 14 and 20 eV (Table 1). The dwell time varied between 0.05 and 0.25 sec. The
171 electrospray source block and the desolvation temperature were 120 °C and 350 °C,
172 respectively. The argon collision gas flow was kept constant at 0.25 mL/min. The cone and
173 desolvation nitrogen gas flow were set at 50 L/h and 800 L/h, respectively. The characteristic
174 precursor and daughter ions that were selected for detection of PE metabolites in multiple
175 reaction monitoring mode (MRM) are presented in Table 1. The daughter ions provided from
176 the transitions were used for the quantification.

177

178 2.4 Method development

179 2.4.1 Extraction solvent nature selection for the DLLME

180 The extraction conditions of PE metabolites were carried out as follows: 2 mL of
181 TFA:MeOH solution (1M) were added to 30 mg nails spiked with IS (5-6 ng) and native PE
182 metabolites (8-18 ng). Then, the extraction was performed by US at 45°C. After 2h, the extracts
183 were collected and together with 120 µL of each extraction solvent (trichloroethylene,
184 chloroform, tetrachloroethylene, chlorobenzene or toluene were tested) were dispersed in 1 mL
185 of water prior placed in a conic screw-cap vial. Then, the samples were centrifuged (6 min,
186 3500 rpm) and the settled organic phase was collected. The volume was completely evaporated
187 under a gentle nitrogen flow and the residue was resuspended in 100 µL H₂O: ACN (75:25,
188 v/v). The selection criteria for the suitable extraction condition was based on the instrumental
189 response factor (Area analyte × (IS Concentration/IS Area) and in the extraction efficiency (%
190 recoveries).

191

192 2.4.2 Optimization of the extraction parameters based on the Taguchi design

193 After the selection of the extraction solvent for the DLLME, other parameters such as the
194 extraction time, acid media, vortex time and trichloroethylene volume (µL) were tested based
195 on an orthogonal experimental array, namely the Taguchi design [44].

196 Thus, a Taguchi design composed by 4 factors and 3 levels summarized in the Table 2 was
197 tested (1st and 2nd design). A total of nine experiments were carried out by following the L₉
198 orthogonal array. Each experiment was conducted in triplicate and the averages of the peak
199 areas were analyzed. Also, the S/N ratio was calculated according with the following equation:

$$\frac{S}{N} = -\log_{10} \frac{1}{n} (\sum y^2) \quad (\textbf{Equation 1})$$

200

201 Where n is the number of observations and y the data (average of peak area).

202 Moreover, in the method optimization two sets of combinatorial designs were tested (1st and
203 2nd Taguchi design) which the main goal in the first set of experiments was to find the best
204 condition, leading us afterwards to the fine tuning and achievement of the optimal (steady state)
205 set of extraction in the 2nd design model.

206

207 2.5 Final US-DLLME procedure for extraction of PE metabolites from human nails

208 Two mL of TFA:MeOH mixture (2M) was added to the reaction vials containing the powder
209 nails (30 mg) and 5-6 ng of IS. The samples were extracted in the US during 2h at 45°C. After,
210 the methanolic extract (~ 2 mL) was collected to a clean vial and used as dispersion solvent in
211 DLLME. The extraction solvent, i.e., trichloroethylene (180 µL) was mixed with the dispersion

212 solvent and both were simultaneously transferred at once to 5 mL screw cap conic vials
213 containing 1 mL of ultra-pure water. Afterwards, a cloudy (triphasic phase) solution was
214 formed by the fine organic droplets dispersed in the aqueous solution and the samples were
215 vortex during 3 min for promoting a better enrichment of the analytes in the organic phase.
216 After, the samples were centrifuged (6 min, 3500 rpm) for separation of the small μ L drop
217 deposited at the bottom. The organic phase was collected using a μ L syringe (Hamilton,
218 Bonaduz, Switzerland) into 300 μ L conical glass inserts (Supelco, Bellefonte, USA). Further,
219 the organic solvent was completely evaporated under a gentle nitrogen flow, the extracts were
220 reconstituted in 100 μ L of H₂O: ACN (75:25, *v/v*) and 10 μ L was injected in the UPLC system
221 (Figure 1).

222

223 2.6 US-DLLME method validation and nail samples analyses

224 The method precision was evaluated by the intra- (repeatability; $n = 6$) and inter-day
225 (intermediate precision; $n = 6$, during 3 consecutive days) variation for two spike levels (low
226 spike level (LQ) of 4-12 ng and high level (HQ) of 7-23 ng in nails, depending of the
227 compound). The relative standard deviations within and between days (% RSD_{within} and %
228 RSD_{between}) were determined.

229 Since there is no certified reference material available for the analysis of PE metabolites in
230 nails, the method accuracy was evaluated based on the recoveries (%) calculated for non-spiked
231 pool nail sample, LQ and HQ spike levels. The uncertainty was expressed as the expanded
232 uncertainty U using a coverage factor $k = 2$, corresponding to a confidence level of 95 %.

233 The quality control (QC) was ensured by analyzing the levels of the target analytes in
234 procedure blanks (4 blanks extracted per day and injected in parallel with the nail extracts) that
235 were spiked with same amount of IS used for spiking the nails. The variation of the instrument
236 performance (RSD %) during the injection of individual samples was evaluated by the injection
237 a middle range standard (μ g/L) in each ten samples.

238 The instrumental limit of quantification (LOQ_i) was defined as the concentration of analyte
239 in the lowest calibration standard (0.3-0.7 μ g/L) at which the signal-to-noise ratio of the
240 quantification ion is greater than 10 times. The method limit of quantification (LOQ_m) was
241 defined as 3 times the standard deviation (SD) of the PE metabolites's concentration that could
242 be measured in procedure blanks (99 % confidence level). If the analytes were not detected in
243 the blanks, thus the LOQ was defined as the LOQ_i.

244 In the end, the levels of the analytes detected in the procedure blanks were afterwards
245 subtracted from those detected in the nail samples of ten volunteers.

246

247 **3. Results and Discussion**

248 In the US-DLLME method development, a set of parameters like the extraction solvent
249 nature, extraction solvent volume, vortex time as well as acid media nature, its concentration
250 and extraction time were optimized. The first parameter that was optimized is the one which
251 has more significance on the DLLME step, i.e., extraction solvent nature. This step is crucial
252 once only ternary mixtures that form stable emulsions and that are separable into two distinct
253 phases after centrifugation, are a result of a successful DLLME. If no stable cloudy state and no
254 sedimented phase is observed, it means that the triphasic phase in DLLME was not successfully
255 reached, therefore the analyte's transfer did not occur.

256

257 **3.1 Optimization of the extraction solvent nature in DLLME**

258 In DLLME, halogenated and non-halogenated solvents can be used depending on the nature
259 of the target analytes [33-36]. Thus, before choosing the suitable extraction solvent, it is
260 important to consider in advance its chemical properties which should comply with the
261 essential DLLME requirements [33].

262 Based on this criteria, a set of five halogenated organic solvents were selected for the
263 DLLME in this study (see Table SI-1 in Supporting Information) using the mixture
264 TFA:MEOH (1M) as dispersion solvent (see experimental conditions in section 2.4.1).

265 Stable cloudy solutions were formed immediately after mixing both extraction and
266 dispersion solvents (containing the metabolites) in the aqueous phase. Nevertheless, when
267 using chloroform (120 µL) a bad dispersion was observed and consequently a small and
268 instable settle phase was formed (8-28 µL) after centrifugation. In fact, chloroform has the
269 highest water solubility among the extraction solvents selected, thus a higher volume was
270 needed for better dispersion and subsequent drop formation. However, the addition of a higher
271 volume was not tested here because: 1) in DLLME the use of organic solvent should be
272 minimized as much as possible depending on the overall conditions (in µL range); 2) add more
273 chloroform would promote less enrichment of the analytes in the final extract resulting in a
274 formation of higher drop volume (not desirable) and less enrichment of the analytes. The
275 dispersion (cloudy state) using other solvents was well performed, therefore the settle phase
276 volume ranged from 80 to 100 µL (except for toluene which volumes between the replicates
277 were more variable and around 62 to 100 µL).

278 According to the results (Figure 2), high and low response factors were obtained for
279 trichloroethylene and toluene, respectively. Additionally, the recoveries of the metabolites were

280 better for trichloroethylene, i.e., the % R ranged between 80 and 108 %. Due to the low
281 response factor and non-reproducible replicates observed for chloroform, then this solvent was
282 excluded from further method development. The trichloroethylene was considered the best
283 extraction solvent in the US-DLLME.

284

285 3.2 Outputs of the Taguchi design during method development

286 3.2.1 First Taguchi design

287 The analysis of data resulted from a combinatorial design such as the analysis of (mean)
288 peak areas, the S/N ratio or the analyses of the variance (ANOVA), are the recommended
289 parameters to identify the influence of each factor in the Taguchi model [44]. Nevertheless, the
290 conceptual approach of analyzing the output by plotting the factors (and levels) and analyzing
291 their effect based on the increase of peak area/response, seems to lead to a fast and easy visual
292 identification of the most significant factors in the extraction.

293 In this study the analysis of the results provided by the Taguchi method were firstly based
294 on plotting the peak areas of each significant factor for the first and second models (i.e.,
295 extraction time, acid media type and concentration, vortex time and trichloroethylene volume)
296 against each level. Later on, the S/N was also evaluated in both models for confirmation of the
297 chosen optimal conditions.

298 Firstly, the results presented in Figure 3, show that extraction time, acid media and the TC
299 volume are the most significant factors in the first design. The best peak areas were represented
300 for 2h of extraction, TFA in the acid media and 140 µL of TC. Accordingly, there is on one
301 hand a more predominant increase of the peak areas after increasing the extraction time and TC
302 volume, mostly for MEHP and for MBP isomers. On the other hand, for MEP, 5-oxo-MEHP,
303 5-OH-MEHP and MBzP, there was only a slightly increase of the peak areas observed for these
304 two parameters. Thereby, it is noticed that the optimal conditions were not yet reached,
305 therefore further optimization is needed for extraction time and TC volume, as well as for the
306 vortex time which no clear trend was observed for most of the analytes (except for MEHP).

307 In addition, the use of different acids (i.e. the acid media) were tested in the 1st Taguchi
308 design. Here, strong acid such as the TFA (pKa of -0.25) favored the protonation of the
309 analytes before the ionization in negative mode in the LC-MS/MS. The other two acids tested,
310 HCl and HNO₃ (pKa -6 and -1.45, respectively) although are considered strong acids, do not
311 have the same dissociation degree that TFA has, making it a much more powerful acid [45].
312 Also, some authors have already reported that TFA promotes an efficient extraction of the PE

313 metabolites from a similar matrix, i.e. the hair [31]. Nevertheless, any other extraction method
314 (and/or condition) were so far explored for these PE metabolites in human nails.

315 Thus, 2h extraction, 140 μ L TC volume, 2 min vortex and TFA: MeOH (1M) were the
316 conditions selected as starting point for further optimization (in the 2nd Taguchi design).

317

318 3.2.2 Second Taguchi design

319 The 2nd Taguchi was performed for fine-tuning the method extraction. Therefore, the best
320 conditions extracted from the first model were used as initial optimization conditions. In
321 addition, two more levels (i.e., 2h30 and 3 h (extraction time); 0.5 M and 2 M (TFA:MeOH
322 concentration); 3 and 4 min (vortex time); 160 μ L and 180 μ L (TC volume)) for each
323 parameter were tested.

324 With regard to the results (Figure 3), there was in general a significant increase from the
325 level 1 to level 3 for each factor (parameter), except for the extraction time where there was no
326 significant contribution to carry on the extraction for more than 2h. The same was observed for
327 the vortex time, i.e. 3 min was considered enough time to promote interaction between the TC
328 and the analytes in the cloudy system. This step is important to accelerate the transfer of the
329 analytes and subsequent the drop formation. However after 3 min, we observed that the peak
330 areas did not improve and beyond that it would make the method undesirable more time-
331 consuming.

332 The responses based on the acid media concentration increase were substantially high for all
333 metabolites, except for MEHP. From our knowledge this metabolite is one of the major (among
334 the studied PE metabolites) present in nails, therefore its detection in nails using TFA:MeOH
335 (2M) will promote the release of relatively high amounts still.

336 The increase from 140 to 180 μ L promoted better enrichment of the analytes in the final
337 extracts (high areas, except for MBP isomers where a slightly decreased is noticed), thereby
338 the optimal condition selected was 180 μ L.

339

340 3.2.3 The S/N evaluation-Taguchi design

341 Another factor which can be representative of the extraction efficiency in the Taguchi
342 design, is the S/N ratio (Equation 1) which results are presented in Figure 4.

343 Here we explained that for the 1st model (section 3.2.1), the level 3 represents the lowest
344 S/N values for the extraction time (120 min), vortex time (3 min) and TC volume (140 μ L)
345 while for the acid media level 1 (TFA) shows better S/N (i.e. lower value).

346 Further, the optimal conditions (2nd Taguchi, section 3.2.2) are confirmed by the level 3
347 which is visibly the best for TFA concentration (i.e., 2M) and also for the TC volume (180 µL);
348 the level 2 showed lower S/N than level 3 for vortex time (i.e. for 3 min), therefore again
349 increasing the vortex time seems do not contribute positively for the extraction of the analytes
350 to the organic phase (as shown before in Figure 3).

351 The steady-state condition resulting from the variation of the extraction time, showed that
352 extending the extraction time more than 2 h is not beneficial. Also, the peak areas for some
353 compounds (Figure 3) represented by 2.5 h and 3 h of extraction had quite similar response.
354 Therefore, the final extraction conditions for the US-DLLME are set as 2h extraction time,
355 TFA:MeOH (2M), 3 min vortex and 180 µL of TC.

356

357 3.3 US-DLLME method validation

358 The linearity method was determined for each compound by injecting seven standard
359 solutions in the concentration range of 0.3-490 µg/L. The concentration of the IS was constant
360 in all standard solutions, however dependent on the analyte (13-18 µg/L) once the standard
361 solutions were made on mass weight basis (and not volume). The calibration curves fulfilled
362 the requirements of a linear fit, where the square regression coefficients (R^2) were higher than
363 0.99 for all calibration curves.

364 Since there are no existing certified reference materials (CRM) for the analysis of PEs in
365 nails, the accuracy of the method (% recoveries) was assessed based on spiking the nails with
366 two levels of the target compounds (high (HQ) and low (LQ) spike levels) and evaluating the
367 levels of PE metabolites in a non-spiked pool sample. After, the accuracy levels for each
368 analyte was determined based on a curve fit of the corrected concentrations (difference between
369 detected concentration and theoretical spike in nails) considering as points: the zero (non-
370 spiked sample), LQ (4-12 ng) and the HQ (7-23 ng) levels. Therefore, good accuracy was
371 achieved in the present method for all compounds (% R ranged between 79 % and 108 %). The
372 precision (intra and inter-day) was expressed as relative standard deviation (RSD %) for both
373 LQ and HQ spike levels. The intra-day variation did not exceed 14 % and 9 % for LQ and
374 HQ, respectively. Also, a good inter-day precision was determined (RSD % < 17 %). The
375 uncertainty for these experiments were acceptable (ranging from 14 % to 24 % (LQ) and
376 between 13 % to 26 % (HQ)), except for MEHP (Table 3).

377 Although, all efforts were done for reducing or eliminating the analyte levels' from the
378 procedure blanks, thereby this was not completely possible and some metabolites were still
379 detected. Albeit, we are analyzing metabolites and not the phthalate precursor compounds, the

380 probability of detecting false positives provided by external contamination is minimized in this
381 study. As preventive actions, all glassware was decontaminated in the over (450 °C) prior
382 extraction, reducing the amount of PEs that could come from environmental contamination
383 (being deposited in glass surface and which by chemical hydrolysis in TFA could form some
384 monoesters that are not exclusively enzymatically formed (e.g. MEP)). Thereby, a set of
385 procedure blanks was done in the same day as the sample's extraction, subtracting the direct
386 external contamination effect possibly occurred during sample preparation. The concentration
387 of the analytes in the procedure blanks was relatively low (0.03 to 0.18 ng/g). In the US-
388 DLLME, the LOQ_m was determined based on the levels in procedure blanks, therefore LOQ_m
389 varied between 2 and 14 ng/g (Table 3). Also, QC standards (middle point calibration standard)
390 and solvent blanks were analyzed for instrumental quality control and assessment of external
391 contamination. The RSD % between standard replicates did not exceed 11 %, thus good
392 analytical quality on the measurements was ensured.

393

394 3.4 Method applicability: Nail Analyses

395 The US-DLLME method was applied to the determination of the PE metabolites in nails of
396 ten Belgian individuals. The average levels ranged between 3.6 ng/g and 918 ng/g (Table 4)
397 showing the good applicability of the developed method and consequently the importance of
398 nails as matrix of choice in HBM field. The major metabolites were the MEHP, the MBP
399 isomers and the MEP (detected in 100 % of the studied individuals) (see chromatograms in
400 Supporting Information). The MEHP oxidative metabolites were minority detected in nails
401 (<LOQ-3.8 ng/g) and their detection seems not relevant in this matrix, instead MEHP seems a
402 better biomarker of DEHP if nails are considered in HBM studies. Although, MBzP was not the
403 major metabolite detected, their exposure in nails should not be neglected (40 % of detection
404 frequency).

405 The comparison of the present method with other extraction methods for nails was not
406 possible due to the lack of literature in this field. Thus, the only possible comparison was done
407 between methods that determined sorts of PE metabolites in matrices that are not usually
408 explored (e.g. hair, saliva, serum, plasma). The methods reported in Table 5, suggest that in
409 other matrices (instead urine), enzymatic (for serum and plasma, for instance) or acid (for hair)
410 hydrolysis [7, 23, 31] followed by solid phase extraction (SPE) or liquid-liquid extraction
411 (LLE) need to be carried out, firstly to destroy the matrix and release the target metabolites and
412 then to concentrate them (and remove interferences) in the final extracts. Nevertheless, we
413 pointed out some disadvantages of those methods such as the extraction time exceeds largely

414 2h, the extraction (and clean-up) is labour extensive (time consuming), an higher volume of
415 solvents (organic, acids/bases, etc.) is used especially for SPE where a higher volume of waste
416 solvents, increase of overall costs.

417 Instead, the optimized US-DLLME method presents good potential in the determination of
418 PE metabolites in nails (high detection in ng/g for some metabolites), comparable or lower
419 extraction time, low costs because less solvent are used in the extraction (μ L) and less waste is
420 produced, so in this sense this method can be consider more environmental friendly. Therefore,
421 the proposed method seems of high relevance in HBM for determining and quantifying PE
422 metabolites via nails.

423

424 **4. Conclusions**

425 For the first time a new, fast and more environmental friendly extraction approach than the
426 conventional solid-phase extraction (SPE) or liquid-liquid extraction (LLE) was developed,
427 combining the US and the DLLME for assessing seven PE metabolites in human nails. Taking
428 advantage of the well-known miniaturization features in DLLME, thus less organic solvents
429 were used (μ L of trichloroethylene), high preconcentration of the target analytes in final μ L
430 drop was successfully obtained and the extraction efficiency was increased by the vortex
431 association upon the cloudy phase formation.

432 The Taguchi design enabled the simplification of the number of experiments ensured by the
433 linear combination of the levels and factors, which therefore reduces the workflow, solvents
434 and sample consumption, etc. Also, the US-DLLME optimal conditions were achieved for a
435 reduced amount of nail samples (30 mg). The good precision, accuracy and low quantification
436 (instrumental or method) limits of the method and its applicability for determining the levels of
437 the metabolites in ten individuals is successfully demonstrated. Also, the first insights of this
438 study showed that nails can be a valuable alternative matrix for measuring human exposure to
439 certain chemicals.

440

441 **Acknowledgements**

442 The authors wish to acknowledge the [European Union] Seventh Framework Programme
443 ([FP7/2007-2013] under *grant agreement n°* [316665] (A-TEAM) by funding support of this
444 research and A. Alves PhD grant (Marie Curie). The authors gratefully acknowledge the
445 participants for the contribution of nail samples.

446

447

448 **Figure Captions**

449
450 **Figure 1.** US-DLLME scheme for analysis of PE metabolites in human nails.

451
452 **Figure 2.** Optimization of the extraction solvent nature in DLLME.

453
454 **Figure 3.** Output from 1st and 2nd Taguchi models during method optimization (parameters
455 tested: extraction time, acid media and concentration, vortex time, trichloroethylene volume).

456
457 **Figure 4.** S/N values for the 1st and 2nd Taguchi designs.

458
459 **Table 1.** MRM transitions, collision energy, cone voltage and retention times for the target
460 metabolites and respective IS.

461
462 **Table 2.** Selected factors and levels in the 1st and 2nd Taguchi design for the US-DLLME.

463
464 **Table 3.** Analytical performance of the US-DLLME in nails.

465
466 **Table 4.** Levels of the 7 PE metabolites in ten individuals.

467
468
469 **Table 5.** Comparison between the US-DLLME method and other extraction methods for
470 determining PE metabolites in non-conventional matrices (nails, hair, serum and seminal
471 plasma).

472
473 **Supporting Information**

474
475 **Table SI-1.** Density, solubility in water and polarity index for five halogenated solvents.

476
477 **Figure SI-1.** Chromatograms of Σ (MnBP, MiBP), MEHP, MEP and MBzP detected in nails of
478 two participants (participant 8 and 3).

480 **References**

- 481 1. Alves A., Kucharska A., Erratico C., Xu F., Den Hond E., Koppen G., Venermen G., Covaci
482 A., Voorspoels S. (2014) Human biomonitoring of emerging pollutants through non-invasive
483 matrices: state of the art and future potential. *Anal Bioanal Chem* 406:4063–4088.
- 484 2. Braun J.M., Just A.C., Williams P.L., Smith K.W., Calafat A.M., Hauser R. (2013) Personal
485 care product use and urinary phthalate metabolite and paraben concentrations during pregnancy
486 among women from a fertility clinic. *J Expo Sci Environ Epidemiol* 24:459–66.
- 487 3. Adibi J.J., Perera F.P., Jedrychowski W., Camann D.E., Barr D., Jacek R., Whyatt R.M.
488 (2003) Prenatal Exposures to Phthalates among Women in New York City and Krakow, Poland.
489 *Environ Health Perspect* 111:1719–1722.
- 490 4. Barr DB, Silva MJ, Kato K, Reidy J.A., Malek N.A., Hurtz D., Sadowski M., Needham L.,
491 Calafat A.M. (2003) Assessing Human Exposure to Phthalates Using Monoesters and Their
492 Oxidized Metabolites as Biomarkers. *Environ Health Perspect* 111:1148–1151.
- 493 5. Wittassek M., Koch H.M., Angerer J., Brüning T. (2011) Assessing exposure to phthalates -
494 the human biomonitoring approach. *Mol Nutr Food Res* 55:7–31.
- 495 6. Romero-Franco M., Hernández-Ramírez R.U., Calafat A.M., Cebrián M.E., Needham L.,
496 Teitelbaum S., Wolff M.S., López-Carrillo L. (2011) Personal care product use and urinary
497 levels of phthalate metabolites in Mexican women. *Environ Int* 37:867–71.
- 498 7. Koch H.M., Lorber M., Christensen K.L.Y., Pälmke C., Koslitz S., Brüning T. (2013)
499 Identifying sources of phthalate exposure with human biomonitoring: Results of a 48h fasting
500 study with urine collection and personal activity patterns. *Int J Hyg Environ Health* 216:672–
501 681.
- 502 8. Hauser R., Calafat A.M. (2005) Phthalates and Human health. *Occup Environ Med* 62:806–
503 818.
- 504 9. Foster P.M.D., Lake B.G., Thomas L.V., Cook M.W., Gangolli S.D. (1981) Studies on the
505 testicular effects and zinc excretion produced by various isomers of monobutyl-ortho-phthalate
506 in the rat. *Chem Biol Interact* 34:233–8.
- 507 10. Saravanabhan G., Murray J. (2012) Human biological monitoring of diisononyl phthalate
508 and diisodecyl phthalate: a review. *J Environ Public Health* 2012:810501.
- 509 11. Ventrice P., Ventrice D., Russo E., De Sarro G. (2013) Phthalates: European regulation,
510 chemistry, pharmacokinetic and related toxicity. *Environ Toxicol Pharmacol* 36:88–96.
- 511 12. Koch H.M., Christensen K.L.Y., Harth V., Lorber M., Brüning T. (2012) Di-n-butyl
512 phthalate (DnBP) and diisobutyl phthalate (DiBP) metabolism in a human volunteer after single
513 oral doses. *Arch Toxicol* 86:1829–39.
- 514 13. Peck C.C., Albro P.W. (1982) Toxic potential of the plasticizer Di(2-ethylhexyl) phthalate in
515 the context of its disposition and metabolism in primates and man. *Environ Health Perspect*
516 45:11–7.
- 517 529

- 530 14. Wittassek M., Angerer J. (2008) Phthalates: metabolism and exposure. Int J Androl 31:131–
531 8.
- 532
- 533 15. Blount B.C., Silva M.J., Caudill S.P., Needham L.L., Pirkle J.L., Sampson E.J., Lucier G.W.,
534 Jackson R.J., Brock J.W. (2000). Levels of Seven Urinary Phthalate Metabolites in a Human
535 Reference Population. Environ Health Perspect 108:979-82.
- 536
- 537
- 538 16. Bornehag C-G., Lundgren B., Weschler C.J., Singsgaard T., Hagerhed-Engman L., Sundell J.
539 (2005) Phthalates in Indoor Dust and Their Association with Building Characteristics. Environ
540 Health Perspect 113:1399–1404.
- 541
- 542 17. Abb M., Heinrich T., Sorkau E., Lorenz W. (2009) Phthalates in house dust. Environ Int
543 35:965–70.
- 544
- 545 18. Directive 2005/90/EC relating to restrictions on the marketing and use of certain dangerous
546 substances and preparations (substances classified as carcinogenic, mutagenic or toxic to
547 reproduction — c/m/r). L33/28.
- 548
- 549 19. Directive 2003/36/EC relating to restrictions on the marketing and use of certain dangerous
550 substances and preparations (substances classified as carcinogens, mutagens or substances toxic
551 to reproduction). L156/26
- 552
- 553 20. Directive 2005/84/EC relating to restrictions on the marketing and use of certain dangerous
554 substances and preparations (phthalates in toys and childcare articles). L 344/40.
- 555
- 556 21. Directive 2004/93/EC amending Council Directive 76/768/EEC for the purpose of adapting
557 its Annexes II and III to technical progress. L 300/13.
- 558
- 559 22. Commission Regulation (EC) N° 552/2009 amending Regulation (EC) N° 1907/2006 of the
560 European Parliament and of the Council on the Registration, Evaluation, Authorisation and
561 Restriction of Chemicals (REACH) as regards Annex XVII. 2009.
- 562
- 563 23. Frederiksen H., Jørgensen N., Andersson A-M. (2010) Correlations between phthalate
564 metabolites in urine, serum, and seminal plasma from young Danish men determined by isotope
565 dilution liquid chromatography tandem mass spectrometry. J Anal Toxicol 34:400–10.
- 566
- 567 24. Huang P-C., Kuo P-L., Guo Y-L., Pao-Chi L., Ching-Chang L. (2007). Associations between
568 urinary phthalate monoesters and thyroid hormones in pregnant women. Hum Reprod 22:2715–
569 2722.
- 570
- 571 25. Griffiths C., Camara P., Saritelli A., Gentile J. (1988). The *in vitro* serum protein-binding
572 characteristics of bis-(2-ethylhexyl) phthalate and its principal metabolite, mono-(2-ethylhexyl)
573 phthalate. Environ Health Perspect 77:151-56.
- 574
- 575 26. Lake B.G., Phillips J.C., Linell J.C., Gangolli S. (1977). The *in vitro* hydrolysis of some
576 phthalate diesters by hepatic and intestinal preparations from various species. Toxicol Appl
577 Pharmacol 39:239-248.

- 578
- 579 27. Preuss R., Koch H.M., Angerer J. (2005) Biological monitoring of the five major metabolites
580 of di-(2-ethylhexyl)phthalate (DEHP) in human urine using column-switching liquid
581 chromatography-tandem mass spectrometry. *J Chromatogr B Analyt Technol Biomed Life Sci*
582 816:269–80.
- 583
- 584 28. Kato K., Silva M.J., Wolf C., Gray L.E., Needham L.L., Calafat A.M. (2007) Urinary
585 metabolites of diisodecyl phthalate in rats. *Toxicology* 236:114–122.
- 586
- 587 29. Kondo F., Ikai Y., Hayashi R., Okumura M., Takatori S., Nakazawa H., Izumi S., Makino
588 T. (2010) Determination of five phthalate monoesters in human urine using gas chromatography-
589 mass spectrometry. *Bull Environ Contam Toxicol* 85:92–6.
- 590
- 591 30. Servaes K., Voorspoels S., Lievens J., Noten B., Allaerts K., Van De Weghe H., Vanermen
592 G. (2013) Direct analysis of phthalate ester biomarkers in urine without preconcentration:
593 method validation and monitoring. *J Chromatogr A* 1294:25–32.
- 594
- 595 31. Chang Y.J., Lin K.L., Chang Y.Z. (2013). Determination of Di-(2 ethylhexyl)phthalate
596 (DEHP) metabolites in human hair using liquid chromatography-tandem mass spectrometry. *Clin
597 Chim Acta*. 420:155-9.
- 598
- 599 32. Hines E.P., Calafat A.M., Silva M.J., Mendola P., Fenton S.E. (2009) Concentrations of
600 phthalate metabolites in milk, urine, saliva, and Serum of lactating North Carolina women.
601 *Environ Health Perspect* 117:86–92.
- 602
- 603 33. Rezaee M., Assadi Y., Milani Hosseini M-R., Aghaee E., Ahmadi F., Berijani S. (2006)
604 Determination of organic compounds in water using dispersive liquid-liquid microextraction. *J
605 Chromatogr A* 1116:1–9.
- 606
- 607 34. Han D., Row K.H. (2011). Trends in liquid-phase microextraction, and its application to
608 environmental and biological samples. *Microchim Acta* 176:1-22.
- 609
- 610 35. Alves A.C.H., Gonçalves M.P.B., Bernardo M.M.S., Mende B. (2012). Dispersive liquid–
611 liquid microextraction of organophosphorous pesticides using nonhalogenated solvents. *J. Sep.
612 Science*, 35: 2653–2658.
- 613
- 614 36. Alves A.C.H., Gonçalves M.P.B., Bernardo M.M.S., Mende B. (2011), Validated dispersive
615 liquid–liquid microextraction for analysis of organophosphorous pesticides in water. *J. Sep.
616 Science*, 34: 1326–1332.
- 617
- 618 37. Ghani J.A., Choudhury I.A. (2004). Application of Taguchi method in the optimization of
619 end milling parameters. *J. Mater. Process. Technol.*, 145:84-92.
- 620
- 621 38. Albero B., Sánchez-Brunete C., García-Valcárcel A.I., Pérez R.A., Tadeo J.L. (2015)
622 Ultrasound-assisted extraction of emerging contaminants from environmental samples. *Trends
623 Anal Chem* 71: 110-118.

- 624 39. Guo L., Lee H.K. (2013) Vortex-assisted micro-solid-phase extraction followed by low-
625 density solvent based dispersive liquid-liquid microextraction for the fast and efficient
626 determination of phthalate esters in river water samples. *J Chromatogr A* 1300: 24-30.
- 627
- 628 40. Cinelli G., Avino P., Notardonato I., Centola A., Russo M.V. (2013). Rapid analysis of six
629 phthalate esters in wine by ultrasound-vortex-assisted dispersive liquid–liquid micro-extraction
630 coupled with gas chromatography-flame ionization detector or gas chromatography-ion trap
631 mass spectrometry. *Anal Chim Acta* 769: 72-78.
- 632
- 633 41. Pérez-Outeiral J., Millán E., Garcia-Arrona R. (2016) Determination of phthalate in food
634 simulants and liquid samples using ultrasound-assisted dispersive liquid-liquid microextraction
635 followed by solidification of floating organic drop. *Food Control* 62: 171-177.
- 636
- 637 42. Viñas P., Campillo N., Pastor-Belda P., Oller A., Hernández-Córdoba M. (2015).
638 Determination of phthalate esters in cleaning and personal care products by dispersive liquid-
639 liquid microextraction and liquid chromatography-tandem. *J Chromatogr A* 1376: 18-25.
- 640
- 641 43. Sun J-N., Shi Y-P., Chen J. (2013) Simultaneous determination of plasticizer di(2-
642 ethylhexyl)phthalate and its metabolite in human urine by temperature controlled ionic liquid
643 dispersive liquid–liquid microextraction combined with high performance liquid
644 chromatography. *Anal Methods* 5: 1427-1434.
- 645
- 646 44. Taguchi G, Yokoyama Y (1994). Taguchi Methods: Design of Experiments. American
647 Supplier Institute, Dearborn MI,in conjunction with the Japanese Standards Association, Tokyo,
648 Japan.
- 649
- 650 45. Munegumi T (2013). Where is the Border Line between Strong Acids and Weak Acids?.
651 *World Journal of Chemical Education* 1:12-16.
- 652
- 653