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Does biotransformation of aryl phosphate flame retardants in blood cast a new perspective on their debated biomarkers?

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21 **Abstract**

22 Aryl phosphate flame retardants (aryl-PFRs), such as triphenyl phosphate (TPHP) and 2-
23 ethylhexyl diphenyl phosphate (EHDPHP), are emerging contaminants that can exhibit toxic
24 properties, including severe aquatic toxicity and endocrine disruptive effects. Monitoring
25 exposure to aryl-PFRs through specific biomarkers is necessary to assess the health risk
26 associated with chronic exposure. Hydrolytic serum enzymes could play an important role in
27 the formation of the hydrolysis product diphenyl phosphate (DHP), the seemingly most
28 abundant *in vivo* biomarker of TPHP in urine. Here, we assess whether serum enzymes have
29 an impact on the toxicokinetics of TPHP and EHDPHP and on the contribution of both aryl-
30 PFRs to *in vivo* DHP levels. TPHP and EHDPHP were incubated separately with pooled
31 human serum to measure the formation of hydrolysis products DHP and 2-ethylhexyl phenyl
32 phosphate (EHPHP) by liquid chromatography-tandem mass spectrometry. Clearance of
33 TPHP and EHDPHP was 70 and 8.6 mL/min/L serum (as measured by formation of DHP
34 and EHPHP, respectively). No discernible amount of DHP was produced from EHDPHP by
35 serum hydrolases. Our results suggest that serum hydrolases can significantly contribute to the
36 *in vivo* levels of DHP formed from TPHP and can play an important role in the
37 toxicokinetics, toxicity, and selection of biomarkers for aryl-PFRs.

38

39 Introduction

40 Aryl phosphorus flame retardants (aryl-PFRs) and plasticizers are used in a large variety of
41 consumer products (plastics, textile, paints, etc.) and their use is increasing as replacements
42 for the recently banned polybrominated diphenyl ethers (PBDEs).¹ Despite their well-intended
43 use to delay the spreading of fire and/or to improve the physicochemical properties of
44 polymers, aryl-PFRs are chemicals of concern since they easily leach into the environment
45 and can exert toxic properties. Recent evidence has appeared regarding endocrine disruptive
46 effects of triphenyl phosphate (TPHP);²⁻⁴ one of the most commonly used aryl-PFR (see
47 Figure 1), raising suspicion against its use in consumer goods, such as electronics, curtains,
48 and building materials.¹ Exposure to this chemical may also cause hepatotoxicity and reduced
49 fertility.⁵⁻⁷ Another ubiquitous aryl-PFR is 2-ethylhexyl diphenyl phosphate (EHDPHP),
50 which, differently from TPHP, has one aliphatic side chain attached to the phosphate group
51 (Figure 1). EHDPHP is used similarly to TPHP as an additive in vinyl polymers and also as
52 plasticizer in food packaging.⁸⁻⁹ Chronic exposure to EHDPHP could be also worrisome due
53 to its potential for developmental toxicity.⁹ For this reason, the Californian government has
54 listed EHDPHP as a priority chemical for biomonitoring.¹⁰

55 TPHP is abundantly present in the indoor environment, including dust and air.¹ Considering
56 the likely daily exposure via inhalation, dermal uptake, and inadvertent ingestion, more
57 information regarding the presence and behavior of TPHP in the human body is needed.
58 Previous research suggested that TPHP was mostly transformed to diphenyl phosphate
59 (DPHP) in the liver by hydrolytic and oxidative enzymes,¹¹⁻¹² and thus this metabolite was
60 considered its major biomarker *in vivo*. Efforts to monitor human exposure to TPHP have
61 revealed urine levels of more than 100 ng/mL of DPHP.¹³⁻¹⁵ In contrast, the biomonitoring of
62 EHDPHP exposure has not received much attention yet, despite the ubiquity and potential
63 toxicity of this PFR. So far, the correlation between the exposure to TPHP via dust and the

64 presence of DPHP in human urine was significant in only one study.¹⁴ Since several other
65 aryl-PFRs share a common structure with TPHP and have shown to produce DPHP as a
66 metabolite or degradation product,¹⁶⁻¹⁷ it remains unclear which aryl-PFR exposure is
67 reflected in the biological levels of DPHP. Additionally, the possible existence of unexplored
68 biotransformation pathways of TPHP and other aryl-PFRs leading also to DPHP could be a
69 confounding factor in the interpretation of urine levels of DPHP. For instance, when only its
70 formation from TPHP by liver enzymes is considered, the estimated exposure to TPHP would
71 be much higher than when the sum of several substrates and/or pathways are taken into
72 account.

73 More toxicokinetic research is needed to provide a better exposure assessment to aryl-PFRs
74 and to improve the understanding of the formation of DPHP as suitable biomarker by
75 exploring the possibility of other sites of PFR metabolism in the human body, and more
76 specifically blood. Serum enzymes, such as phosphate triester hydrolases (including
77 paraoxonases, E.C.3.1.8.1), may also be involved in the hydrolysis of aryl phosphates
78 considering the structural similarity between these PFRs and organophosphate pesticides, e.g.
79 paraoxon.¹⁸ Paraoxonases were demonstrated to be the major factor of paraoxon clearance in
80 blood compared to irreversible binding to circulating cholinesterases¹⁹ and may therefore also
81 play an important role in the toxicokinetics of aryl-PFRs in the human body. Furthermore,
82 aryl esterases in serum have also shown hydrolytic activity towards phosphate esters such as
83 paraoxon.²⁰

84 In the present study we aim to evaluate the extent of DPHP formation from aryl-PFRs by
85 serum hydrolases by investigating TPHP and EHDPHP as representative compounds.
86 Formation rates of the identified hydrolysis products of these two PFRs are presented in this
87 paper. Special attention is given to the formation of DPHP to elucidate new pathways for the

88 generation of this possibly major aryl-PFR biomarker and to clarify the potential role of this
89 additional biotransformation route in the toxicity and biomonitoring of TPHP.

90

91 **Experimental section**

92 Materials

93 Tris(hydroxymethyl) aminomethane (TRIS), CaCl₂, paraoxon, 4-nitrophenol, DPHP, and
94 DPHP-d₁₀ were purchased from Sigma Aldrich (Bornem, Belgium). TPHP was purchased
95 from Chiron AS (>98% purity, Trondheim, Norway). 2-ethylhexyl phenyl phosphate
96 (EHPHP) was synthesized by dr. Vladimir Belov (Max Planck Institute, Göttingen, Germany)
97 as an ammonium salt (98% purity). EHDPHP was bought from Sigma Aldrich at a purity of
98 91%. Serum was collected in the frame of an ongoing study (registered at
99 <http://clinicaltrials.gov/> with number NCT01778868). This study was approved by the Ethical
100 Committee of the Antwerp University Hospital (Belgian Registry number B30020097009)
101 and all participants provided their written informed consent. Serum from 10 lean volunteers
102 was pooled and stored at -20°C until analysis. Acetonitrile (analytical grade) was obtained
103 from Merck KgA Chemicals (Darmstadt, Germany) and ultrapure water (18.2 MΩ) from an
104 Elga LabWater water purification instrument (Saint Maurice, France).

105

106 Incubation methods

107 *Assessing the linearity of serum concentration and incubation time.* Optimization experiments
108 were conducted similarly to our previous study.²¹ Incubation mixtures consisted of 100 mM
109 TRIS buffer (adjusted to pH 8.5 at 37 °C), 2 mM CaCl₂, and serum in a range of 0.5 to 3%
110 concentration (v/v), final volume of 500 μL. Reactions were initiated by adding 20 or 100 μM
111 of TPHP (in 1% acetonitrile, total volume). All samples were prepared in duplicate. Reactions
112 were stopped by adding 250 μL of acetonitrile containing 25 ng of DPHP-d₁₀ as internal

113 standard. A negative control was similarly prepared without the addition of serum. Samples
114 were vortexed and centrifuged, after which supernatants were filtered using a 0.22 μm nylon
115 filter. At both concentration levels and with 10 min incubation time, DPHP formation was
116 linear up to 2% serum (v/v). Different incubation times (3-20 min) were then tested with 20 or
117 100 μM of TPHP. At both concentrations, DPHP formation was linear up to 10 min.
118 Similar experiments were conducted for EHDPHP at 50 μM . A linear increase in EHPHP
119 production was observed up to 1% (v/v) serum. Using this serum concentration, the optimal
120 incubation time was 15 min. DPHP formation was also observed, with similar optimal
121 conditions for incubation time.

122

123 *Substrate concentration dependent kinetics.* TPHP (1-200 μM) was incubated in 100 mM
124 TRIS buffer (adjusted to pH 8.5 at 37 $^{\circ}\text{C}$) and 2 mM CaCl_2 in a final volume of 500 μL at 37
125 $^{\circ}\text{C}$. 10 μL serum (2% v/v) was added to initiate the reaction. For every concentration,
126 replicates were prepared. After 10 min, reactions were stopped by addition of 250 μL
127 acetonitrile containing 25 ng DPHP- d_{10} as internal standard. Samples were vortexed and
128 filtered using a 0.22 μm nylon membrane filter before analysis. The same incubation
129 experiments were carried out for EHDPHP, with slightly different values of serum
130 concentration (1%, v/v) and incubation time (15 min).

131 Negative controls (same preparation procedure, three replicates per substrate level) were
132 prepared by adding acetonitrile to the samples before the reaction was initiated in order to
133 account for chemical hydrolysis or impurities present in the standards. For all incubating
134 conditions, paraoxon was included in separate replicates at concentrations of 1 to 100 μM ,
135 monitoring 4-nitrophenol as the target end product. The samples containing paraoxon were
136 both included for the purpose of positive control, and to have a comparison between kinetics
137 of enzymatic hydrolysis for an organophosphate pesticide and these two aryl-PFRs.

138 Procedural blanks were treated the same as the samples, except for the addition of 5 μL of
139 pure acetonitrile instead of substrate, and were included in every batch, to exclude possible
140 interferences by contamination.

141

142 *Serum hydrolase activity.* The hydrolase activity of the pooled serum employed in this study
143 was tested by assessing the formation of 4-nitrophenol. Paraoxon (1 mM final concentration)
144 was incubated in triplicate in 100 mM TRIS buffer (adjusted to pH 8.5 at 37 $^{\circ}\text{C}$) and 2 mM
145 CaCl_2 in a final volume of 500 μL at 37 $^{\circ}\text{C}$. 5 μL serum (1% v/v) was added to initiate the
146 reaction. Since enzyme reaction rate is dependent on substrate concentration, the same
147 incubation conditions and substrate concentration were used as described by Furlong et al.²⁰,
148 so that specific activity of serum paraoxonase could be compared with literature. Negative
149 controls containing the same mixture were prepared simultaneously, but with the addition of
150 the internal standard DPHP-d₁₀ (25 ng) in 250 μL acetonitrile before incubation (so that the
151 reaction was not initiated). All samples except the negative controls were quenched after 5
152 min by the addition of the same amount of internal standard solution.

153

154 *Analytical Quality Control.* Matrix-matched calibrations were prepared to account for
155 possible matrix effects on the signal response of EHPHP and 4-nitrophenol (for which
156 isotopically labelled standards were not used, instead DPHP-d₁₀ was employed). Calibration
157 curves were prepared for all analytes in the same mixture of buffer, solvent and serum as the
158 samples and the negative controls and were analyzed under the same conditions.

159

160 Instrumental analysis

161 DPHP, EHPHP, and 4-nitrophenol were measured by liquid chromatography-tandem mass
162 spectrometry (LC-MS/MS) using an Agilent 1290 LC coupled to a 6460 triple quadrupole

163 MS. A Kinetex Biphenyl column (50 x 2.1 mm, 1.7 μ m, Phenomenex) was used for the
164 separation of the extracts. For DPHP and EHPHP analysis, the aqueous mobile phase
165 consisted of 5 mM ammonium acetate in water at pH 5.5 (A), while for 4-nitrophenol the
166 aqueous mobile phase (of 5 mM ammonium acetate) was acidified with acetic acid until pH
167 4.5. Acetonitrile was used as organic solvent (B) for all analyses. The following gradient
168 conditions were applied: 5% B (1.5 min hold), increase to 20% B in 0.5 min, increase to 60%
169 B in 2 min, followed by a sharp increase to 95% B in 0.8 min (4 min hold), and equilibration
170 at starting conditions for 3 min. Flow rate was 0.3 mL/min, temperature 40 $^{\circ}$ C, injection
171 volume 1 μ L. MS parameters for the analysis of serum extracts were: gas temperature 300 $^{\circ}$ C,
172 sheath gas heater 250 $^{\circ}$ C, gas flow 5 L/min, sheath gas flow 11 L/min, nebulizer pressure 45
173 psi, capillary and nozzle voltage 3000 and 500 V, respectively. MRM transitions are provided
174 in Table S1. Cell accelerator voltage was 4 V.

175

176 Data analysis

177 Regression analysis for each metabolite formation rate was applied using GraphPad Prism
178 5.0. The following models were tested for the calculation of metabolite formation kinetics:
179 linear regression (on truncated datasets), Michaelis-Menten equation, and substrate inhibition
180 model. No weights were applied for the regression analysis. An F-test was used to detect if
181 one of the models fit significantly better, and the simplest model was retained if there was no
182 significant difference in fit ($p > 0.05$).

$$183 \quad v = \frac{V_{\max} \times [S]}{K_m + [S]} \quad (\text{Michaelis-Menten model})$$

$$184 \quad v = \frac{V_{\max}}{1 + K_m/[S] + [S]/K_i} \quad (\text{substrate inhibition model})$$

185

186 **Results**

187 Selection of main hydrolysis products

188 4-nitrophenol was selected as target end product of the positive control based on literature,²⁰
189 while DPHP and EHPHP (Figure 1) were selected based on their detection in *in vitro* liver
190 assays^{12,22} and their resemblance to hydrolysis products of organophosphate pesticides. For
191 TPHP and EHDPHP, we rather monitored the phosphate diester as end product, because
192 phenol and 2-ethylhexanol are less specific for use as biomonitoring target in *in vivo* studies.

193

194 Serum hydrolase activity determination

195 The activity of serum hydrolases was determined by the formation of the metabolite 4-
196 nitrophenol from paraoxon. In the procedural blanks, 4-nitrophenol was not detected,
197 although a minor amount of 4-nitrophenol was found in the negative controls. This amount
198 was subtracted from the measured concentration in the positive control samples. In every
199 tested batch of TPHP or EHDPHP hydrolysis experiments, the formation of 4-nitrophenol
200 from paraoxon was detectable and constant.

201 Following the conditions described in the literature by Furlong et al.²⁰ for paraoxonase,
202 namely 5 min and 1% (v/v) serum, the activity of the serum employed in this study was
203 calculated as 221 ± 35 U/L (1 unit = 1 μ mol paraoxon hydrolyzed per min). We also tested
204 optimal incubation conditions for paraoxon as we did for TPHP and EHDPHP, observing that
205 net 4-nitrophenol formation increased linearly with longer incubation times or higher serum
206 concentrations, up to 10 min and 2% (v/v) serum. Because of good repeatability, these
207 conditions were used to calculate the rate of 4-nitrophenol formation from 1 to 100 μ M of
208 paraoxon. In contrast to the hydrolysis products of TPHP and EHDPHP, 4-nitrophenol
209 showed a linear formation pattern using substrate levels of 1 to 100 μ M.

210

211 Comparative hydrolysis rates of the three organophosphates

212 Hydrolysis rates of TPHP, EHDPHP, and paraoxon were compared at low substrate
213 concentrations (in the linear range of 1 to 10 μM), since most realistic exposure scenarios to
214 flame retardants involve chronic, low dose exposure. An F-test revealed that the slopes of the
215 regression line for each dataset were significantly different ($p < 0.001$), with slopes (95%
216 confidence interval) of 70 (58 - 82), 8.6 (7.3 - 9.9), and 485 (468 - 503) $\text{mL}/(\text{min L serum})$ for
217 DPHP, EHPHP, and 4-nitrophenol, respectively. These slopes roughly indicate the extent of
218 *in vivo* clearance in blood. The data of DPHP and EHPHP are shown in Figure S1.

219

220 Kinetic profile of TPHP hydrolysis by serum

221 Formation of DPHP, as hydrolysis product of TPHP, increased linearly with increasing time
222 or serum concentration up to 10 min and 2% (v/v) serum. However, DPHP was also observed
223 in the enzyme-deactivated negative controls, which may have been due to chemical
224 hydrolysis.²³ Therefore, DPHP levels in enzyme-deactivated negative controls were
225 subtracted from the levels measured in the samples to obtain net enzymatic hydrolysis rates.
226 The highest concentration (200 μM) was omitted, because TPHP proved to be insoluble in the
227 buffer at this level. Regression analysis indicated Michaelis-Menten kinetics were the most
228 suitable model for DPHP formation (Figure 2), with an extrapolated K_m of $108 \pm 30 \mu\text{M}$ and
229 V_{max} of $8.16 \pm 1.41 \mu\text{mol}/\text{min}/\text{L serum}$. Clearance, calculated as V_{max}/K_m , was $75.6 \text{ mL}/\text{min}/\text{L}$
230 serum, which is comparable to the value of $70 \text{ mL}/\text{min}/\text{L serum}$ mentioned above.

231

232 Kinetic profile of EHDPHP hydrolysis by serum

233 EHPHP was identified as the main hydrolysis product of EHDPHP by serum enzymes, while
234 formation of DPHP was in any case minor (see below). Optimization of the initial rate
235 conditions showed a clear linear EHPHP formation rate up to 15 min at 1% (v/v) serum.

236 When plotting the rate of EHPHP formation versus EHDPHP concentration, a clear decline
237 was observed at the highest tested concentration suggesting substrate inhibition kinetics
238 (Figure 3). When applying non-linear regression to the data using both the substrate inhibition
239 model and the Michaelis-Menten equation, the latter was preferred because more precise
240 estimates of K_m , V_{max} were obtained based on the available data points. V_{max} was 0.273 (SD
241 0.030) $\mu\text{mol}/\text{min}/\text{L}$ serum, while K_m was 13.1 (SD 4.7) μM . However, when the highest level
242 was removed from the dataset, the V_{max} and K_m values increased up to to 0.41 $\mu\text{mol}/\text{min}/\text{L}$
243 serum and 29 μM , respectively. SDs of these non-linear regression parameters increased to
244 0.12 $\mu\text{mol}/\text{min}/\text{L}$ serum and 17 μM , indicating less precise estimates.

245 The formation of DPHP in EHDPHP preparations was corrected for its presence in the
246 negative control samples, with rates ranging from 0.025 $\mu\text{mol}/\text{min}/\text{L}$ serum at 2.4 μM
247 EHDPHP to 0.28 $\mu\text{mol}/\text{min}/\text{L}$ serum at 114 μM EHDPHP. However, since TPHP was present
248 in the EHDPHP standard (3.3% impurity), DPHP could originate both from EHDPHP and
249 from the TPHP impurity. To clarify if the formation of DPHP in the EHDPHP preparations
250 was originating only from the TPHP impurity, we compared the slope of the regression of
251 DPHP formation (32 $\text{mL}/\text{min}/\text{L}$ serum, 95% confidence interval - 0.56 to 70 $\text{mL}/\text{min}/\text{L}$
252 serum) versus the calculated level of TPHP present as impurity in the EHDPHP preparations
253 (2.4-20 μM) with the slope obtained from the TPHP-DPHP regression data explained in
254 section 3.2 (70, 95% confidence interval: 58.3 to 81.7 $\text{mL}/\text{min}/\text{L}$ serum). As such, the slope of
255 DPHP in EHDPHP samples was significantly lower than in TPHP samples (F-test, $p < 0.001$).
256 The position of average DPHP formation in the lower tested levels of EHDPHP (2.4-20 μM)
257 and TPHP (1-2.5 μM) are shown in Figure S2. Furthermore, we compared at the different
258 tested levels of EHDPHP if the measured rate of DPHP was similar to its predicted range
259 (calculated based on the concentration of TPHP as impurity). This was the case for all tested
260 levels of EHDPHP starting from 20 μM on. Only at the lower level of 2.4 μM of EHDPHP, at

261 which the concentration of TPHP as impurity (0.1 μM) could be too low to detect its
262 hydrolysis product, DPHP formation was higher, respectively, than the predicted range (10
263 versus 5.8 to 8.1 $\mu\text{mol}/\text{min}/\text{L}$ serum). These data suggest possible additional production of
264 DPHP from EHDPHP which would occur at a much lower rate than from TPHP. At lower
265 levels than 2.4 μM of EHDPHP, the overall produced amount of DPHP was below the
266 quantification limit (1 ng/mL).

267

268 **Discussion**

269 Serum hydrolase (including paraoxonase) activity

270 The results of our study indicate that hydrolysis by serum enzymes takes place for the two
271 aryl-PFRs. While chemical hydrolysis was reported to occur for TPHP and EHDPHP at
272 alkaline pH,²³ the levels of DPHP and EHPHP in the samples incubated with active serum
273 enzyme exceeded those of the negative controls with deactivated serum enzymes. This means
274 that the enzymatic hydrolysis rate exceeded the chemical hydrolysis rate. Lower clearance
275 was observed for EHDPHP (8.6 mL/min/L serum) and TPHP (70 mL/min/L serum) compared
276 to paraoxon (485 mL/min/L serum). Therefore *in vivo* serum enzymes could degrade aryl-
277 PFRs albeit to different extents. Not only the structural differences between the compounds
278 can play a role here, but also the presence of enzyme polymorphisms. For instance, the assay
279 we used was optimized for measuring the activity of the enzyme paraoxonase, which is
280 known to be dependent on the functional genotype.²⁴ In this study, we used pooled serum
281 because of ethical reasons as a tool to investigate the general role of serum in
282 biotransformation of aryl-PFRs in the human body and DPHP formation. Our data therefore
283 represent the average biotransformation rate in serum from a group of volunteers.

284

285 Formation of DPHP from TPHP and EHDPHP

286 When considering the two potential hydrolysis products of EHDPHP, the discrepancy
287 between the different tested levels in the likeliness of EHDPHP contributing to DPHP
288 formation may have several reasons. Firstly, at 2.4 μM EHDPHP, the concentration of TPHP
289 ($\sim 0.1 \mu\text{M}$) was well below the lowest tested level ($\sim 1 \mu\text{M}$) in the TPHP test samples.
290 Extrapolation of the regression would thus lead to a larger uncertainty in the predicted
291 formation of DPHP from the TPHP impurity. Secondly, since we observed indications of
292 substrate inhibition effects for EHPHP formation, these effects might also influence DPHP
293 formation from EHDPHP negatively. Taking this into account, it is difficult to estimate any
294 significant contribution in DPHP formation from EHDPHP by serum enzymes. However, we
295 emphasize that at levels of EHDPHP $< 10 \mu\text{M}$, where the amount of the TPHP impurity may
296 be insignificant to contribute to DPHP formation, the observed minor amount of DPHP could
297 come mainly from EHDPHP. This could suggest a considerably lower DPHP formation rate
298 from EHDPHP than from TPHP that is masked by the impurity at higher levels. A similar
299 observation was reported using oxidative liver enzymes.²² Furthermore, the preferred
300 hydrolysis product in serum appears to be EHPHP with phenol as leaving group, which is
301 logical considering that in paraoxon hydrolysis by serum enzymes the preferred leaving group
302 is also aromatic (4-nitrophenol). This is also in agreement with a recent study that investigated
303 hydrolysis of EHDPHP in aqueous solutions at alkaline pH, where they reported EHPHP as
304 degradation product and not DPHP.²³ One issue we did not account for was the possible
305 contribution to EHPHP formation by another impurity, namely bis(2-ethylhexyl) phenyl
306 phosphate, that was reported by Ballesteros-Gomez et al.²² However, it should be noted that
307 this impurity can only contribute by producing 2-ethylhexanol as leaving group, which is a
308 less favorable reaction compared to the production of phenol.

309

310 Implications for the toxicokinetics, toxicity, and biomonitoring of aryl-PFRs

311 Given the observed activity of hydrolytic enzymes in serum towards TPHP and EHDPHP, the
312 stability of these PFRs in the human body circulation may be limited, especially in the case of
313 TPHP. This would have several consequences in the field of toxicokinetics, toxicity, and
314 biomonitoring of TPHP and other aryl-PFRs:

315 1. Less TPHP is available for distribution (and accumulation) in tissues;
316 2. The amount of TPHP that reaches the liver (after uptake from the gastro-intestinal
317 system, lungs, or skin) in its original form may be strongly reduced. For this reason, the *in*
318 *vivo* metabolite profile of TPHP can differ from that observed in *in vitro* studies using liver
319 cells or subcellular fractions. In other words, the metabolite 4-HO-TPHP (with its glucuronide
320 and sulfate conjugates), that was found to be the major or second major TPHP metabolite,^{12,25-}
321 ²⁶ could be much less abundant in biological matrices, such as blood and urine, than DPHP. In
322 addition, there is no information whether 4-HO-TPHP would be more stable towards serum
323 enzymes compared to TPHP. Recently, 4-HO-TPHP-(glucuronide) was evidenced in human
324 urine from Canadians²⁷ at much lower concentrations than DPHP. As the authors suggested,
325 the discrepancy with 4-HO-TPHP being the major metabolite formed by hepatic enzymes
326 could be due to either *in vitro* – *in vivo* differences or to contributions to DPHP formation by
327 other aryl-PFRs, such as EHDPHP (to a minor extent) and resorcinol bis(diphenyl phosphate).
328 However, we recommend that also the role of serum enzymes should not be overlooked in the
329 study of toxicokinetics of aryl-PFRs.

330 3. Our findings are also important for the interpretation of *in vitro* toxicity tests. On the
331 one hand, serum enzymes could play a role in the detoxification of TPHP since interactions
332 that were found between nuclear receptors and TPHP^{1,28} could be occurring only to a very
333 minor extent because of the limited stability of TPHP in the circulation. Furthermore, DPHP
334 is, because of its high polarity, probably cleared faster from the human body and less capable
335 of penetrating tissues by passive diffusion to cause harm. Next, Kojima et al.²⁹ demonstrated

336 that the 4-HO-TPHP metabolite expressed similar toxicity compared to its parent. Therefore,
337 the reduced formation of 4-HO-TPHP in the liver due to earlier hydrolysis of TPHP to DPHP
338 in the blood, suggests that this process could be considered as detoxification. On the other
339 hand, we still do not have full knowledge how TPHP induces toxic effects in various organs.
340 Moreover, Su et al.⁶ discovered that in chicken embryonic hepatocytes, toxic effects were
341 mostly correlated to DPHP. Therefore, bio-activation by serum enzymes cannot be ruled out.

342 4. All of the aforementioned points could also apply to EHDPHP and its hydrolysis
343 product EHPHP. However, for this aryl-PFR the impact of serum enzymes on its toxicokinetic
344 and toxicodynamic properties and biomonitoring targets would be less pronounced since the
345 enzymatic activity was relatively low. We do emphasize however, that the metabolites of
346 EHDPHP were only recently confirmed in human tissues and no assessment has yet been
347 made of the toxicity of the transformation products, including EHPHP.

348

349 Implications for selecting biomarkers of exposure

350 DPHP has become heavily debated as biomarker for measuring TPHP exposure in e.g. human
351 urine samples because it likely lacks specificity. Both results of our current and previous
352 study could not evidence a significant contribution from EHDPHP to the presence of DPHP
353 in biological tissues.²² This indication of specificity is a welcome finding for the discussion of
354 which TPHP metabolite should be monitored, since 4-HO-TPHP (sum of the free form and its
355 glucuronide) was not detected in urine¹⁵ or at very low levels.²⁷ Therefore, DPHP may be a
356 useful biomarker of exposure to TPHP alongside with total 4-OH-TPHP so that low exposure
357 can be monitored (using DPHP). In subjects with high urinary DPHP concentration, total 4-
358 OH-TPHP levels might help to confirm high exposure to TPHP. However, since another
359 study¹⁷ demonstrated that DPHP is present as impurity in resorcinol bis(diphenyl)phosphate
360 (RDP or PBDPP) formulations and it can be also formed by the spontaneous hydrolysis of

361 this compound at physiological pH and body temperature, DPHP is recommendable to be
362 used as human biomarker for aryl-PFR exposure rather than for TPHP only. Additionally, our
363 findings can be used as a recommendation for using EHPHP as an alternative biomarker of
364 EHDPHP exposure, in addition to the oxidation products that were reported in hepatic
365 metabolism.²²

366 In conclusion, serum enzymes were shown to be involved in the biotransformation of TPHP
367 and EHDPHP in humans which could affects the metabolite profile, toxicokinetics, and
368 toxicity of both aryl-PFRs. The formation of DPHP from EHDPHP could not be estimated
369 and was minor in any case, which implies that EHDPHP is not likely a confounding factor in
370 the interpretation of DPHP levels in human body fluids and tissues.

371 Further research is needed to estimate inter-individual differences in the clearance of aryl-
372 PFRs in blood and to determine the extent of the impact of serum hydrolases on the kinetics
373 and toxicity of these PFRs and their metabolites in humans. Furthermore, the availability of
374 an EHDPHP standard of higher purity to assess the possible formation of DPHP as a minor
375 hydrolysis product would be highly desirable.

376

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382

383 **Supporting information**

384 Additional information including Table S1 and Figures S1 and S2 can be found in the
385 supporting information. This information is available free of charge via the Internet at
386 <http://pubs.acs.org/>.

387

388 **References**

- 389 1. Wei, G. L.; Li, D. Q.; Zhuo, M. N.; Liao, Y. S.; Xie, Z. Y.; Guo, T. L.; Li, J. J.; Zhang,
390 S. Y.; Liang, Z. Q. Organophosphorus flame retardants and plasticizers: sources, occurrence,
391 toxicity and human exposure. *Environ. Pollut.* **2015**, *196*, 29-46.
- 392 2. Kojima, H.; Takeuchi, S.; Itoh, T.; Iida, M.; Kobayashi, S.; Yoshida, T. In vitro
393 endocrine disruption potential of organophosphate flame retardants via human nuclear
394 receptors. *Toxicology.* **2013**, *314*, 76-83.
- 395 3. Fang, H.; Tong, W.; Branham, W. S.; Moland, C. L.; Dial, S. L.; Hong, H.; Xie, Q.;
396 Perkins, R.; Owens, W.; Sheehan, D. M. Study of 202 natural, synthetic, and environmental
397 chemicals for binding to the androgen receptor. *Chem. Res. Toxicol.* **2003**, *16*, 1338-1358.
- 398 4. Liu, X.; Ji, K.; Choi, K. Endocrine disruption potentials of organophosphate flame
399 retardants and related mechanisms in H295R and MVLN cell lines and in zebrafish. *Aquat.*
400 *Toxicol.* **2012**, *114-115*, 173-181.
- 401 5. Meeker, J. D.; Stapleton, H. M. House dust concentrations of organophosphate flame
402 retardants in relation to hormone levels and semen quality parameters. *Environ. Health*
403 *Perspect.* **2010**, *118*, 318-323.
- 404 6. Su, G.; Crump, D.; Letcher, R. J.; Kennedy, S. W. Rapid in vitro metabolism of the
405 flame retardant triphenyl phosphate and effects on cytotoxicity and mRNA expression in
406 chicken embryonic hepatocytes. *Environ. Sci. Technol.* **2014**, *48*, 13511-13519.
- 407 7. Van den Eede, N.; Cuykx, M.; Rodrigues, R. M.; Laukens, K.; Neels, H.; Covaci, A.;
408 Vanhaecke, T. Metabolomics analysis of the toxicity pathways of triphenyl phosphate in
409 HepaRG cells and comparison to oxidative stress mechanisms caused by acetaminophen.
410 *Toxicol. In Vitro.* **2015**, *29*, 2045-2054.
- 411 8. *2-ethylhexyl diphenyl phosphate: test plan and robust summaries*, U.S. EPA High
412 production volume chemical voluntary testing program, Ferro Corporation: Cleveland, OH,
413 2006.
- 414 9. *Environmental risk evaluation report: 2-ethylhexyl diphenyl phosphate (CAS no.1241-*
415 *94-7)*. UK Environment Agency: Bristol, UK, 2009.
416 [https://www.gov.uk/government/uploads/system/uploads/attachment_data/file/290842/scho08](https://www.gov.uk/government/uploads/system/uploads/attachment_data/file/290842/scho0809bqty-e-e.pdf)
417 *09bqty-e-e.pdf*, (), accessed on 24-01-2016.
- 418 10. Meeting of the Scientific Guidance Panel Potential Priority Chemicals: Non-
419 halogenated Aromatic Phosphates. Biomonitoring California, April 11 2013,
420 http://www.biomonitoring.ca.gov/sites/default/files/downloads/041113NhArP_priority_0.pdf
- 421 11. Sasaki, K.; Susuki, T.; Takeda, M.; Uchiyama, M. Metabolism of phosphoric acid
422 triesters by rat liver homogenate. *Bull. Environ. Contam. Toxicol.* **1984**, *33*, 281-288.
- 423 12. Van den Eede, N.; Maho, W.; Erratico, C.; Neels, H.; Covaci, A. First insights in the
424 metabolism of phosphate flame retardants and plasticizers using human liver fractions.
425 *Toxicol. Lett.* **2013**, *223*, 9-15.

- 426 13. Butt, C. M.; Congleton, J.; Hoffman, K.; Fang, M.; Stapleton, H. M. Metabolites of
427 organophosphate flame retardants and 2-ethylhexyl tetrabromobenzoate in urine from paired
428 mothers and toddlers. *Environ. Sci. Technol.* **2014**, *48*, 10432-10438.
- 429 14. Cequier, E.; Sakhi, A. K.; Marce, R. M.; Becher, G.; Thomsen, C. Human exposure
430 pathways to organophosphate triesters - a biomonitoring study of mother-child pairs. *Environ.*
431 *Int.* **2015**, *75*, 159-165.
- 432 15. Van den Eede, N.; Heffernan, A. L.; Aylward, L. L.; Hobson, P.; Neels, H.; Mueller, J.
433 F.; Covaci, A. Age as a determinant of phosphate flame retardant exposure of the Australian
434 population and identification of novel urinary PFR metabolites. *Environ. Int.* **2015**, *74*, 1-8.
- 435 16. Dodson, R. E.; Van den Eede, N.; Covaci, A.; Perovich, L. J.; Brody, J. G.; Rudel, R.
436 A. Urinary biomonitoring of phosphate flame retardants: levels in California adults and
437 recommendations for future studies. *Environ. Sci. Technol.* **2014**, *48*, 13625-13633.
- 438 17. Ballesteros-Gomez, A. M.; Van den Eede, N.; Covaci, A. In vitro human metabolism
439 of the flame retardant resorcinol bis-(diphenylphosphate) (RDP). *Environ. Sci. Technol.* **2015**,
440 *49*, 3897-3904.
- 441 18. Testa, B.; Krämer, S. D., *The biochemistry of drug metabolism: principles, redox*
442 *reactions, hydrolyses*. Verlag Helvetica Chimica Acta: Zürich, Switzerland, 2010; Vol. 1, p
443 278.
- 444 19. Costa, L. G.; Richter, R. J.; Li, W. F.; Cole, T.; Guizzetti, M.; Furlong, C. E.
445 Paraoxonase (PON 1) as a biomarker of susceptibility for organophosphate toxicity.
446 *Biomarkers.* **2003**, *8*, 1-12.
- 447 20. Furlong, C. E.; Richter, R. J.; Seidel, S. L.; Costa, L. G.; Motulsky, A. G.
448 Spectrophotometric assays for the enzymatic hydrolysis of the active metabolites of
449 chlorpyrifos and parathion by plasma paraoxonase/arylesterase. *Anal. Biochem.* **1989**, *180*,
450 242-247.
- 451 21. Van den Eede, N.; Erratico, C.A.; Exarchou, V.; Neels, H.; Covaci, A. In vitro
452 biotransformation of tris(2-butoxyethyl) phosphate (TBOEP) in human liver and serum.
453 *Toxicol. Appl. Pharmacol.* **2015**, *284*, 246-253.
- 454 22. Ballesteros-Gomez, A.; Erratico, C. A.; Van den Eede, N.; Ionas, A. C.; Leonards, P.
455 E.; Covaci, A. In vitro metabolism of 2-ethylhexyldiphenyl phosphate (EHDPHP) by human
456 liver microsomes. *Toxicol. Lett.* **2014**, *232*, 203-212.
- 457 23. Su, G.; Letcher, R.J.; Yu, H. Organophosphate Flame Retardants and Plasticizers in
458 Aqueous Solution: pH-Dependent Hydrolysis, Kinetics, and Pathways. *Environ. Sci. Technol.*
459 **2016**, *50*, 8103-8111.
- 460 24. Furlong, C. E.; Cole, T. B.; Jarvik, G. P.; Pettan-Brewer, C.; Geiss, G. K.; Richter, R.
461 J.; Shih, D. M.; Tward, A. D.; Lulis, A. J.; Costa, L. G. Role of paraoxonase (PON1) status in
462 pesticide sensitivity: genetic and temporal determinants. *Neurotoxicology.* **2005**, *26*, 651-659.
- 463 25. Su, G.; Letcher, R. J.; Crump, D.; Gooden, D. M.; Stapleton, H. M. In Vitro
464 Metabolism of the Flame Retardant Triphenyl Phosphate in Chicken Embryonic Hepatocytes
465 and the Importance of the Hydroxylation Pathway. *Environ. Sci. Technol Lett.* **2015**, *2*, 100-
466 104.
- 467 26. Van den Eede, N.; De Meester, I.; Maho, W.; Neels, H.; Covaci, A. Biotransformation
468 of three phosphate flame retardants and plasticizers in primary human hepatocytes: untargeted
469 metabolite screening and quantitative assessment. *J. Appl. Toxicol.* **2016**, published online:
470 *DOI 10.1002/jat.3293*.
- 471 27. Su, G.; Letcher, R. J.; Yu, H.; Gooden, D. M.; Stapleton, H. M. Determination of
472 glucuronide conjugates of hydroxyl triphenyl phosphate (OH-TPHP) metabolites in human
473 urine and its use as a biomarker of TPHP exposure. *Chemosphere* **2016**, *149*, 314-319.
- 474 28. Pillai, H. K.; Fang, M.; Beglov, D.; Kozakov, D.; Vajda, S.; Stapleton, H. M.;
475 Webster, T. F.; Schlezinger, J. J. Ligand binding and activation of PPARgamma by

476 Firemaster(R) 550: effects on adipogenesis and osteogenesis in vitro. *Environ. Health*
477 *Perspect.* **2014**, *122*, 1225-1232.
478 29. Kojima, H.; Takeuchi, S.; Van den Eede, N.; Covaci, A. Effects of primary
479 metabolites of organophosphate flame retardants on transcriptional activity via human nuclear
480 receptors. *Toxicol. Lett.* **2016**, *245*, 31-39.
481

Figure legends

Figure 1. Structures of 2-ethylhexyl diphenyl phosphate (EHDPHP), triphenyl phosphate (TPHP) and their hydrolysis products, ethylhexyl phenyl phosphate (EHPHP) and diphenyl phosphate (DPHP)

Figure 2. Kinetics of triphenyl phosphate (TPHP) hydrolysis to its diester structure (diphenyl phosphate, DPHP) by serum enzymes.

Figure 3. Kinetic formation of 2-ethylhexyl phenyl phosphate (EHPHP) by serum enzymes in relation to concentration of 2-ethylhexyl diphenyl phosphate (EHDPHP).

Figures

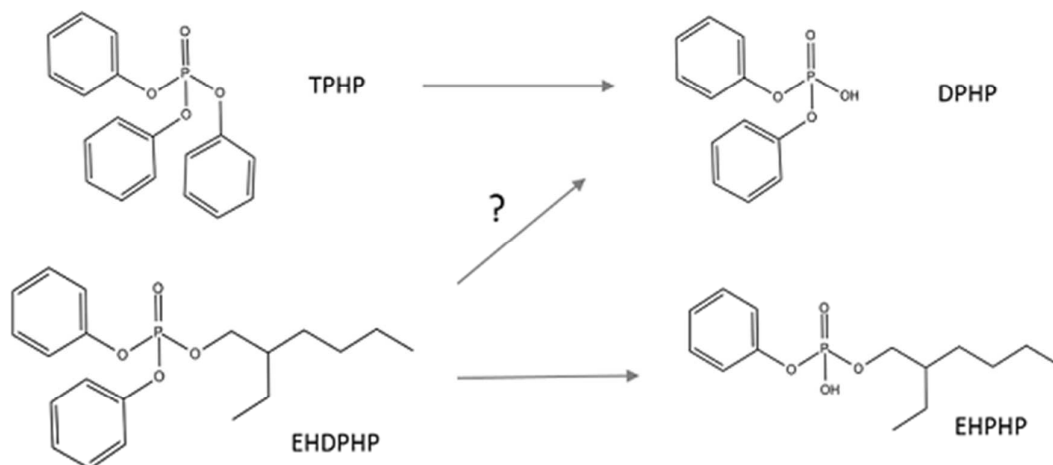


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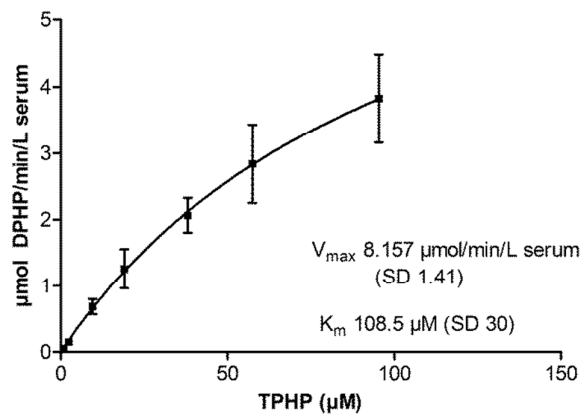


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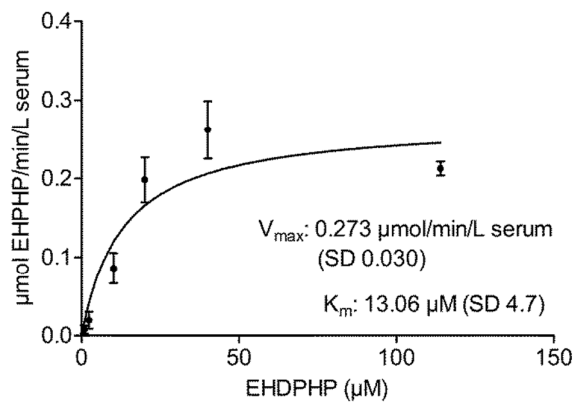
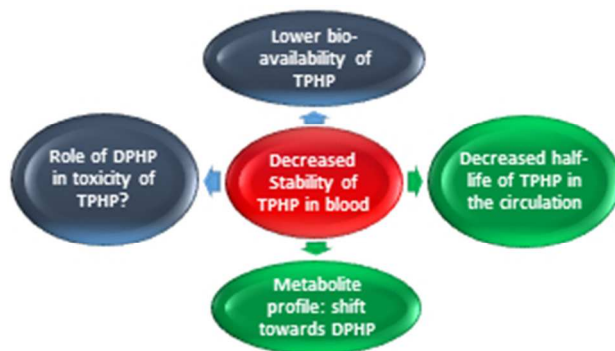


Figure 3. Kinetic formation of 2-ethylhexyl phenyl phosphate (EHPHP) by serum enzymes in relation to concentration of 2-ethylhexyl diphenyl phosphate (EHDPHP).



TOC Art

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