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2
3 **Title**

4 Bacterial community and filamentous population of industrial wastewater treatment plants in Belgium

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12
13 **Abstract**

14 The discharge of industrial water requires the removal of its pollutants, where biological wastewater treatment
15 plants (WWTPs) are the most used systems. Biological WWTPs make use of activated sludge (AS), where bacteria
16 are responsible for the removal of pollutants. However, our knowledge of the microbial communities of industrial
17 plants is limited. Understanding the microbial population is essential to provide solutions to industrial problems
18 such as bulking. The aim of this study was to identify at a high taxonomic resolution the bacterial population of
19 29 industrial WWTPs using 16S rRNA Amplicon Sequencing. Our results revealed that the main functional groups
20 were dominated by *Thauera* and *Zoogloea* within denitrifiers, *Dechloromonas* in phosphate-accumulating and
21 *Defluvicoccus* in glycogen-accumulating organisms. The activated sludge characterization indicated that 59% of
22 the industrial plants suffered from bulking sludge, with DSVI values up to 448 mL/g. From the bulking cases, 72%
23 corresponded to filamentous bulking with *Thiothrix* as most abundant filament, meanwhile the other 28%
24 corresponded to viscous bulking sludge in which *Zoogloea* was the most abundant genus. Furthermore, the
25 bacterial population did not share a core of taxa across all industrial plants. However, 20 genera were present in at
26 least 50% of the plants comprising the general core, including *Thauera*, *Ca. Competibacter* and several
27 undescribed microorganisms. Moreover, statistical analysis revealed that wastewater salinity strongly affected the
28 microbial richness of the industrial plants. The bacterial population across industrial plants differed considerably
29 from each other, resulting in unique microbial communities that are attributed to the specificity of their
30 wastewaters.

31 **Keywords:** Industrial activated sludge WWTP, filamentous bacteria, viscous bulking.

32 **Key points**

- 33 • The general core taxa of industrial plants was mostly made up of undescribed bacterial genera.
34 • Filamentous bacteria constituted on average 4.1% read abundance of the industrial WWTPs.
35 • Viscous bulking remains a significant type of bulking within industrial WWTPs.

45 Introduction

46 Wastewater treatment plants (WWTPs) ensure that the effluents that are discharged in the local surface waters
47 are properly treated. The activated sludge (AS) process is the most commonly used biological treatment system
48 for industrial and domestic WWTPs. The industry is one of the largest consumers of water worldwide, accounting
49 for 57% of total water consumption in Europe and 19% globally (FAO 2015). The treatment of industrial
50 wastewaters can be challenging due to the large variety of pollutants that it can contain and the seasonal variation
51 in wastewater quality. Furthermore, some wastewaters lack nutrients such as nitrogen and phosphorus, or have
52 high a salinity and contain toxic compounds. These conditions, frequently observed in industrial wastewaters, can
53 inhibit the biological process of wastewater treatment.

54 Continuous flow activated sludge (CAS) systems are amongst the most commonly used AS systems for
55 industrial wastewater. These systems are susceptible to excessive growth of filamentous bacteria (FB) (Caluwé et
56 al. 2017). This type of treatment system is generally characterized by low substrate concentrations, or diffusion
57 dominated conditions that will result in FB extending from the flocs (Martins et al. 2003). In a recent study of 90
58 industrial WWTPs in Belgium, it was found that 37% suffered from bulking problems, and 46% of the total number
59 of plants had a CAS system (Cornelissen et al. 2018). Bulking sludge has been defined as sludge that settles slowly
60 and compacts poorly (Martins et al. 2004), it can be classified as filamentous bulking sludge (FBS) and non-
61 filamentous or viscous bulking sludge (VBS). FBS has been shown to occur in different industrial WWTPs
62 worldwide, including different AS systems such as nutrient removal and enhanced biological phosphorus removal
63 (EBPR) systems. The filaments found to cause bulking in industrial plants are Morphotype 021N in the potato
64 processing industry (Levantesi et al. 2004), *Thiothrix* in the brewery and dairy sector, *Sphaerotilus* in pulp and
65 paper making industry (Oppong et al. 2003) and *Kouleothrix* in the tank truck cleaning sector (Tsertou et al. 2022).
66 Additionally, it has also been reported that VBS can cause poor settling in industrial plants including the bottling
67 and brewing processing (Peng et al. 2003), chemical (Lajoie et al. 2000) and winery (McIlroy et al. 2011). VBS
68 sludge has received little attention as in most bulking cases filaments are the major cause, but so far no clear causes
69 of VBS are known.

70 There is a limited number of studies related to bulking sludge in industrial WWTPs, that have reliably identified
71 filaments and other microorganisms that could be responsible for bulking. In terms of separation problems caused
72 by filaments, there are many contradictions related to the quantity of filaments. Kaetzke et al., 2005 reported that
73 bulking sludge might occur due to a small fraction of some filaments in the biovolume, that is the case of *Ca. M.*
74 *Parvicella* with relative abundance between 0.54% and 2.47% causing bulking sludge (Zhang et al. 2019). Other
75 studies have shown that sludge only bulks when filaments extend from the surface of the flocs (Vervaeeren et al.
76 2005). Nevertheless, the link between the number of filaments and incidents related to them is still not well
77 understood and continues to be a serious problem for AS plants worldwide (Seviour 2010). Furthermore, compared
78 to the extensive literature that exists for FBS, only limited research has addressed VBS, the focus of such studies
79 has been on EPS content (Shao et al. 2019) or operational parameters leading to VBS (Peng et al. 2003), and little
80 attention has been paid to the microbial population involved in VBS in industrial wastewater treatment plants.
81 Additionally, little is known about the microbial communities of industrial AS plants at higher taxonomic
82 resolution such as species level. There is little information on the groups of functional bacteria found in industrial
83 plants, responsible for the removal of nutrients such as nitrogen and phosphorus.

84 The objective of this study is to identify the bacterial community of 31 industrial plants from 7 different
85 industrial sectors at high taxonomic resolution using 16S rRNA Amplicon Sequencing. We aimed to identify the
86 core community and the most abundant genera-species within functional groups including nitrifiers, denitrifiers,
87 phosphate-accumulating and glycogen accumulating organisms. Furthermore, we elucidated the microorganisms
88 involved in bulking, including FBS and VBS, and determined the causes of bulking. Additionally, a summary of
89 the composition of industrial wastewaters is presented as well as their impact on the microbial communities.
90 Finally, the impact of the most encountered FB and their relation with settling properties along the different
91 industrial plants are critically presented.

92

93

94 **Materials and Methods**

95 **Biomass sampling and plant information**

96 Samples including wastewater, activated sludge (AS) and treated effluent were collected from 31 full-scale
97 industrial plants (Table 1, for more details see Table S1 and Table S2). Activated sludge samples (5L) were
98 collected from the homogenized aerobic tanks, and were immediately transported to the laboratory. Samples were
99 taken from each WWTP twice, once during autumn (November-December) and once during summer (June-
100 August). All investigated WWTPs had been in operation for several years under similar conditions. The exact
101 operation time of each plant is however not known.

102 **Sludge characterization**

103 The samples were collected with a maximum of 24 hours before the experiments were carried out, and were
104 kept at 4°C. Biomass concentration was estimated using the mixed liquor suspended solids (MLSS) and mixed
105 liquor volatile suspended solids (MLVSS) (APHA 1998). The morphology of AS samples was characterized using
106 light microscopy, where filaments were identified using the Eikelboom classification based on morphological
107 characteristics observed from phase contrast, Gram and Neisser stain (Eikelboom 2000). Furthermore, the
108 abundance of FB was determined subjectively using a filamentous index (FI, Richard et al., 2003) ranging from 0
109 (no filaments observed) to 6 (excessive filaments observed). Sludge settleability was determined based on sludge
110 volume index SVI (van Loosdrecht et al. 2016). In order to be able to compare the settleability of the different
111 plants, a diluted SVI (DSVI) was performed, the AS was diluted with effluent until the settled volume was between
112 150-250 ml L⁻¹ after 30 min as described in (van Loosdrecht et al. 2016).

113 Most of the information related to the wastewater composition including pH, EC, COD, TN, TP was obtained
114 from the industrial plants. Additionally S²⁻ and VFA were measured using Hach test kit (LCK653 and LCK365),
115 while SO₄²⁻ was measured using Macherey Nagel test kit.

116 **DNA extraction and PCR amplification**

117 DNA was extracted from 2 samples (volume 500 µL) of each WWTPs, using the FastDNA® SPIN kit for soil
118 following the protocol described in (van Loosdrecht et al. 2016). DNA concentration of each extraction was
119 measured using Qubit dsDNA BR Assay kit following the manufacturer's protocol. All samples were then diluted
120 to a final DNA concentration of 7.5 ng/µL. Samples with high DNA concentration (above 100 ng/µL) were diluted
121 twice to minimize dilution error. The 16S rRNA gene amplification was performed targeting the V1-3 region of
122 the 16S rRNA gene, as recommended for AS samples by (Albertsen et al. 2015). Library preparation and
123 sequencing were conducted as described in (Caluwé et al. 2022) using an Illumina MiSeq instrument. The raw
124 sequencing data have been deposited in the NCBI SRA database BioProject ID: PRJNA938135. Amplicon
125 sequencing variants (ASV) were clustered as described in (Dueholm et al. 2020). Taxonomy was assigned using
126 the Microbial Database for Activated Sludge (MiDAS 4) (Dueholm et al. 2022), the obtained dataset was analysed
127 and visualised using R v. 3.2.3 (R Core Team, 2015) with the ampvis2 v. 1.24.0 (Andersen et al. 2018), ggplot2
128 (Gómez-Rubio 2017) and vegan v.2.5-7 packages. Samples from 2 plants were lost during DNA extraction and
129 therefore no amplification was obtained, resulting in sequencing information of 29 plants.

130 **Data analysis**

131 Data analysis was carried out on 2 samples per plant, one representing winter and one representing summer
132 period. The amplicon data were analyzed using the ampvis2 package (Andersen et al. 2018), samples with less
133 than 10000 reads were discarded, resulting in 58 samples (from 29 WWTPs) and 12301 ASVs with a minimum of
134 42756 and maximum of 154543 qualified reads per sample. ASVs read counts were normalized to 100 (i.e. percent)
135 per sample using the function amp_subset_samples() from the ampvis2 R package (Andersen et al. 2018). The
136 effective bacterial sequences were taxonomically assigned from phylum to species level. It was possible to classify
137 ASVs as far as species thanks to the high coverage of the specific database for WWTP bacteria. The non-parametric
138 Kruskal-Wallis test was used to assess the statistically significant differences between read abundance of bacterial
139 genera and process type. Alpha diversity indices, including Shannon and Chao1 indices, were calculated using the
140 “amp_alphadiv” function from the ampvis2 R package. Statistical comparison of alpha diversity indices and
141 reactor configuration were performed using a mixed linear model. Spearman’s rank correlation coefficient was

142 calculated to test the correlation between alpha diversity index (Chao1 and Shannon) and operational conditions
143 (pH, EC, COD, TN, TP).

144 Beta diversity analysis was used to quantify the similarity between the bacterial communities of the different
145 industrial sectors. Principal coordinate analysis (PCoA) based on Bray–Curtis dissimilarity was used to visualize
146 the beta diversity among the samples. On the basis of beta diversity, canonical correspondence analysis (CCA)
147 was used to evaluate the impact of environmental factors on microbial community distribution. Where CCA was
148 constrained to a given variable including EC, pH, COD, TN, TP, SO_4^{2-} , S^{2-} , VFA, MLSS and MLVSS. Prior to the
149 analysis, ASVs that were not present in more than 0.01% relative abundance were removed. The data was first
150 transformed by applying the Hellinger transformation (Legendre and Gallagher 2001). PCoA and CCA plots were
151 generated using the ampvis2 package.

152 **Results**

153 **Industrial WWTPs and their bacterial community**

154 The 31 surveyed plants were designed for different process types, of which 9 were carbon removal plants only
155 (C), 19 had biological nitrogen removal (nitrification-denitrification; N, DN) and 3 reported biological phosphorus
156 removal activity (P). The WWTPs had wastewater flows ranging from 100 to 32000 $\text{m}^3\cdot\text{day}^{-1}$. The majority of the
157 AS plants were configured as continuous flow systems including Unitank and continuously stirred tank reactors
158 (CSTR), 8 sequencing batch reactors (SBR) and 6 membrane bioreactors (MBR). 35% of the plants reported the
159 use of chemical coagulants to improve the AS settling properties and combat filamentous bulking, of which
160 iron(III)chloride was the most used, only 1 plant from the chemical sector reported ozone addition and 1 plant
161 from the bioindustrial sector dosed polyaluminum chloride (PAC) as control method. Table 2 provides an overview
162 of the wastewater composition range among the industrial sectors.

163 In this study, the microbial community composition of 29 industrial WWTPs from 7 different industrial sectors
164 was investigated. The clustering classification resulted in 12301 ASVs, that were taxonomically classified from
165 phylum until species level. A total of 45 known bacterial phyla were identified in 58 sludge samples. There were
166 between 40-44 bacterial phyla with relative abundances of over 1%, accounting for 90.63-94.67% of the total
167 bacteria. The dominant bacterial phyla were *Proteobacteria* (33.7-55.5%), *Bacteroidota* (19.4-38.9%),
168 *Chloroflexota* (1.5-8.1%), *Myxococcota* (1.5-7.4%) and *Acidobacteriota* (1.9-5.4%). The relative abundance of
169 the dominant phyla varied across the different industrial plants and sectors (Fig. 1). Taxonomic classification at
170 the family level revealed that the most abundant families were *Rhodocyclaceae* (*Proteobacteria*) and
171 *Saprospiraceae* (*Bacteroidota*), see Figure S1. The high abundance of *Rhodocyclaceae* was mainly due to the
172 significant presence of the genera *Thauera* and *Zoogloea*, and the undescribed *Denitromonas*. The second most
173 observed bacterial family was *Saprospiraceae*, of which the genera *midas_g_6*, *Ca. Epiflobacter* and
174 *midas_g_1328*, were the most frequently found.

175 At a higher taxonomic resolution, 81.9% of the generated ASVs were classified at genus and 61.3% at species
176 level (Figure S2). Based on mean of relative abundance, it can be observed that the dominant genera are *Thauera*,
177 *Denitromonas* and *Zoogloea*. However, results obtained from the median of relative abundance among industrial
178 sectors suggest that the main organisms found were *Ca. Competibacter*, *Thauera* and *Elin6067*. Since the median
179 results are much lower than that of the mean, it suggests that the distribution of genus relative abundance is
180 negatively skewed within the samples and that the presence of micro-organisms does not depend on the industrial
181 sector to which it belongs, but rather other factors. Among the most commonly encountered bacterial genera are
182 also *Thiothrix* and *Ca. Competibacter*, both belonging to the phylum *Proteobacteria*. *Thiothrix* is a well-studied
183 bulking filamentous bacteria that was encountered mainly in the vegetable, dairy and bioindustrial sectors. *Ca.*
184 *Competibacter* is a glycogen accumulating organism (GAO) that was abundant in the tank-truck cleaning, meat
185 and brewery sectors with mean of 2.4, 1.6 and 1.4%RA and median of 1.5, 0.7 and 0.8%RA respectively.

186 **Core community**

187 Cumulative analysis was used to determine whether abundant ASVs made up a large proportion of the reads.
188 Fig. 2a presents the cumulative read abundance at genus-level ASVs, across samples from winter and summer (58
189 samples). In each sample, the 10 most abundant ASVs made up 7.8% of the total read abundance, and the 100
190 most abundant genus-ASVs made up only 28.7%. The core community was determined as described by (Dueholm

191 et al. 2022), with a minimum relative abundance threshold of 0.1% of the genus-ASVs to be considered part of the
192 core community. The occurrence cutoff was 80% corresponding to strict core, while 50% as general core and 20%
193 loose core across the 56 samples. A large number of ASVs was observed in one or few samples (Fig. 2b), 105
194 genus-level ASVs made up the loose core and only 20 genus-level ASVs comprised the general core community
195 while no ASVs were found among strict core community.

196 The 20 genera that made up the general core community are presented in Fig. 2c. Only few of the general core
197 micro-organisms have been studied in AS systems, for which their in situ physiology is known. These include the
198 genera *Thauera*, *Ca. Epiflobacter*, *Ca. Competibacter* and *Nitrospira*. However, a large number of microorganisms
199 that have not been described in the literature comprise the general core taxa, including *Ferruginibacter*,
200 *Terrimonas*, *Nannosystis* and *Haliangium*.

201 **Functional guilds**

202 The functional groups of organisms recognized as involved in nutrient removal were identified across the
203 different process types (Fig. 3). It was observed that the complexity of the treatment process which included
204 removal of carbon (C, 9 plants), carbon and nitrogen (C, N, DN, 17 plants), carbon, nitrogen and phosphorus (C,
205 N, DN, P, 1 plant), and carbon and phosphorus (C, P, 2 plants) played a key role in the abundance of functional
206 guilds. The microorganisms responsible for carbon and nitrogen removal differ in abundance from those found in
207 municipal plants. This was also the case for plants with biological P activity, where the abundance of well
208 described PAO was lower than those reported in the literature.

209 The main genera within nitrifiers were the ammonia oxidizing bacteria (AOB) *Nitrosomonas* and the nitrite
210 oxidizing bacteria (NOB) *Nitrospira*. Only one undescribed specie (midas_s_11707) accounted for all
211 *Nitrosomonas* abundance, while 16 species corresponded to *Nitrospira* from which *N. nitrosa* and *N. defluvii* were
212 the most abundant (see Figure S3). Furthermore, *N. nitrosa* has been described as a commamox (complete
213 ammonia oxidizer) like organism (Daims et al. 2015). However, based only on 16S sequences, it cannot be stated
214 that the specie *N.nitrosa* represents a true commamox organism. Further research in the role of *N.nitrosa* in
215 industrial plants is needed to determine whether complete ammonia oxidation occurs. Denitrifying bacteria
216 *Zoogloea* were significantly abundant within carbon removal plants (C, 9) however, none of these plants presented
217 biological nitrogen removal, suggesting that their presence was not due to their denitrifying activity. Furthermore,
218 *Thauera* was the dominant denitrifier in the C and N removal systems (C, N, DN, 17 plants). 15 species were
219 found within the genus *Thauera*, of which 2 undescribed species were the most abundant of this genus
220 (midas_s_1356 and midas_s_256).

221 In the three plants where biological phosphorus removal was reported, the most abundant phosphate-
222 accumulating organisms (PAOs) were *Dechloromonas*, followed by *Tetrasphaera* (recently renamed to
223 *Phosphoribacter*, Singleton et al. 2022) and *Ca. Accumulibacter*. From which, *Ca. Accumulibacter* has been
224 extensively studied and is often considered as the most important PAO genus (Petriglieri et al. 2022). Additionally,
225 we investigated the presence of known putative PAOs (Rubio-Rincón et al. 2017; Stokholm-Bjerregaard et al.
226 2017) among these plants. Our results revealed that the most encountered putative PAOs in the industrial samples
227 are *Thiothrix* (10.6%), *Gemmatimonas* (0.5%) and *Ca. Obscuribacter* (0.1%) (see Figure S3). Moreover, these
228 plants treat sulfide rich wastewaters (2-16 mgS.L⁻¹) which could contribute to the mixotrophic metabolism of
229 *Thiothrix* that combines sulfur oxidation with biological P removal (Rubio-Rincón et al. 2017; de Graaff et al.
230 2020). The polyphosphate kinase gene (ppk2) has also been encountered in several species within the genus
231 *Thiothrix*, including *T. Caldifontis* (Chernousova et al. 2009; Matsuura et al. 2021). However, it cannot be
232 concluded that the abundance of *Thiothrix* in the industrial plants is due to its mixotrophic metabolism of S and P,
233 for this, further research is needed on the metabolic potential of *Thiothrix*.

234 Glycogen-accumulating organisms (GAOs) were mainly encountered within C removal (9 plants) from which
235 *Defluvicoccus* and *Ca. Competibacter* were the most abundant genera. The maximum abundance within the
236 industrial plants was 14.3% (sample from meat sector) for *Ca. Competibacter* and 10.4% (sample from tank truck
237 cleaning) for *Defluvicoccus*.

238 **Activated sludge settling properties**

239 The AS characterization revealed that the surveyed industrial plants had DSVI values ranging from 34 to 448
240 mL.g⁻¹. 41% of the studied samples presented good settling properties, with DSVI values below 150 mL.g⁻¹, with
241 mean of 78 mL.g⁻¹ and median of 76 mL.g⁻¹. Few filaments were observed microscopically in the well-settling
242 samples, where some filaments were commonly observed, but were not present in all of the flocs. On average, the
243 filamentous index was 2. However, results from Amplicon sequencing revealed total filaments abundance ranging
244 from 0.003 to 11.3% read abundance, for these well-settling samples. The high abundance of total filaments found
245 in well-settling samples is mainly due to the presence of *Ca. Villigracilis* and *Leptothrix*. *Ca. Villigracilis* was
246 found in 50% of the samples with good settleability with mean of 0.6% and max read abundance of 4.2%. Similarly
247 *Leptothrix* was present in 67% of the well settling AS, with mean of 0.9% and maximum of 7.2% read abundance.

248 59% of the AS samples presented poor settling properties, with DSVI values ranging from 165 to 448 mL.g⁻¹,
249 considered as bulking sludge (BS). In most of the bulking cases a high FI was determined, however in some cases
250 poor settleability was not related to filaments extending out of the flocs (Fig. 4a). In 72% of the bulking cases a
251 high abundance of filaments was microscopically observed (FI 3-6), therefore this type of bulking is referred to as
252 filamentous bulking sludge (FBS). The mean and median of total filaments among FBS were 6.08% and 1.69%
253 respectively. These results show that the abundance of filaments was not uniformly distributed resulting in few
254 cases (7 cases) with high abundance of filaments (10-27.4%), while the majority (17 cases) of FBS presented
255 abundance lower than 5% of total filaments encountered with amplicon sequencing. The remaining 28% of the
256 bulking cases had FI values below 2 and also low abundance with mean and median of 1.5%, therefore it is
257 considered as VBS. According to microscopy results, FBS and VBS presented a very different floc morphology,
258 in which FBS had filaments extending from the flocs while VBS exhibited open poorly compacted flocs with no
259 filaments. This was also confirmed by Amplicon results, where the filament *Thiothrix* was the dominant genus
260 encountered within FBS. The most abundant genera within VBS (Fig. 4b) samples were *Zoogloea* (mean 4.84%)
261 and *Ferruginibacter* (mean 3.26%). These bacteria are well known floc formers involved in extracellular
262 polymeric substances (EPS) production (Han et al. 2018). VBS was observed in plants designed for C (4) and
263 C,N,DN (3) process types, operating under food to microorganisms in the range of 0.05 to 0.18 g COD/g
264 MLSS.day (mean and median 0.12), and the majority of the affected plants were configured as continuous flow
265 feeding systems. A wide range of bacteria produce different types of EPS, including carbohydrates and proteins,
266 that affects microbial aggregates resulting in VBS. Further research into the type of EPS produced and microbial
267 characterization is needed to address the organisms responsible for VBS in industrial plants. Although FB were
268 responsible for most cases of bulking, VBS remains a significant type of bulking within industrial WWTPs.

269 **Filamentous bacterial population**

270 Filamentous bacteria made up a significant part of the microbial community with read abundance ranging from
271 0.003% to 27.4%, and an average read abundance of 4.1% among all the studied samples (58 samples from summer
272 and winter, 29 plants) (Fig. 5c). The most abundant filaments, based on mean relative abundance were *Thiothrix*,
273 *Leptothrix*, *Kouleothrix*, *Ca. Villigracilis* and *Ca. Sarcinithrix* (Fig. 5a), we evaluated the impact of process type
274 and period of sampling on the abundance of the top 5 most abundant FB (see Figure S4 and Table S3). *Thiothrix*
275 (*Proteobacteria*) was found to be the most abundant genus within the bioindustry, dairy and vegetable sectors,
276 being encountered in 31 samples with mean read abundance of 2.28%, reaching 27% in the potato processing
277 industry. *Thiothrix* presence was positively correlated with bulking (correlation coefficient 0.48, p<0.05) and was
278 microscopically observed in high abundance (FI 4-6) as presented in Fig. 5b. *Thiothrix* abundance was strongly
279 influenced by the process type (see Figure S4 and Table S3), being more abundant in complex systems such as
280 nutrient removal plants. Several genera from the *Chloroflexota* phylum were encountered among the dominant
281 filamentous bacteria, from which *Kouleothrix* was found abundant in the meat (mean 0.6%) sector in plants with
282 C and N removal and in the tank-truck cleaning sector (mean 3% and max 18%) where only C removal was
283 performed, *Ca. Villigracilis* found throughout all industrial sectors with a max read abundance of 4.2% in the meat
284 sector, *Ca. Sarcinithrix* found only in the dairy (mean 0.1%), brewery (mean 0.2%) and in the meat (mean 0.8%RA
285 and max of 14.4%) sectors. Only *Ca. Sarcinithrix* abundance was positively correlated (p<0.05, Figure S4 and
286 Table S3) to process type being abundant in nutrient removal plants. No impact of winter and summer samples
287 was observed in the abundance of the described abundant filaments (see Table S3).

288 Among the most abundant filaments only two genera do not always present filamentous morphology in-situ,
289 these include species belonging to the genera *Leptothrix* and *Trichococcus* (Nierychlo et al. 2020). *Leptothrix*,

290 known to have a variable morphology including straight rods cells and filamentous growth, was encountered in 51
291 samples with the highest relative abundance (>1%) in the brewery, meat, dairy and chemical sectors respectively.
292 *Leptothrix* abundance was on average 0.99%, median of 0.23% and maximum of 7.2% in the brewery sector.
293 Additionally, for 4 samples (2 plants) from the chemical sector no abundance of total filaments was identified
294 based on the 16S RNA gene, however a large number of FB extending from the flocs were observed
295 microscopically.

296 External factors affecting the microbial community

297 Alpha diversity analysis was conducted to estimate the microbial richness and diversity within the samples,
298 where the Chao1 measures the microbial richness and Shannon the microbial diversity. Chao1 richness index
299 ranged from 355 to 2432 and Shannon diversity index ranged from 1.95 to 5.92 with an overall lower richness and
300 diversity found in winter samples compared to summer samples (see Figure S5). A linear mixed model was used
301 as quantitative trait association between microbial diversity or richness and process type, industrial sector. No
302 significant association between industrial sector or process type was encountered with microbial richness and
303 diversity (data not shown) among the industrial plants. Additionally, Spearman rank correlation was used to find
304 association between wastewater composition (pH, EC, COD, TN, TP and VFA) and microbial richness and
305 diversity. The results showed that there is a significant correlation between wastewater salinity EC and microbial
306 richness Chao1 (Spearman correlation coefficient 0.42, $p = 0.002$), indicating that high salinity has a negative
307 effect on microbial richness, but no significant correlation was found between EC and diversity.

308 Beta diversity analysis was used to quantify the similarity between microbial communities, PCoA revealed a
309 wide taxonomic diversity across industrial sectors (Fig. 6a) where samples from the same industrial sector did not
310 cluster strongly. Winter and summer samples from individual plants cluster together, indicating an overall
311 microbial stability within a period of 6 months, which was the time between the two samples. However, the first
312 two principal coordinates represented only 10% of the total variation, indicating a lower degree of microbiome
313 similarity. The higher sample dispersion was encountered within the meat sector presumably due to the wide range
314 of its wastewater composition (Table 1), while the dairy, chemical, vegetable and bioindustry did cluster together.
315 These results suggest an overall greater microbial complexity within industrial AS plants.

316 Constrained correspondence analysis CCA was used to determine the influence of external factors on microbial
317 community composition. CCA of the 58 samples, that corresponded to all the studied AS plants, revealed that
318 environmental factors including EC, SO_4^{2-} , and TP had significant effects on the bacterial population belonging to
319 chemical, brewery and bioindustrial sectors (Fig. 6b). Whereas the lower TP encountered in one brewery plant led
320 to a more specific microbial community. Similarly the higher EC and SO_4^{2-} of the chemical wastewater resulted
321 in a more specific microbial community. However, CCA showed that the investigated environmental variables
322 explained only 22.37% of the variability in the industrial WWTPs, suggesting that other factors not considered
323 may explain the microbial variability of the other industrial plants. This can be attributed to the specificity of the
324 incoming industrial wastewaters that will result in unique microbial ensembles. Additionally, CCA analysis did
325 not reveal significant effect of environmental factors on bulking bacterial population, which in turn can be related
326 to the high microbial specificity that exists between the plants, and does not allow for a common pattern to be
327 established.

328 Discussion

329 In this study we aimed to gain more insight into the microbial community of industrial full-scale WWTPs. Our
330 findings indicate that several yet undescribed taxa were highly abundant and comprised the general core
331 community of the studied industrial plants. The main functional groups were dominated by *Thauera* and *Zoogloea*
332 within denitrifiers, *Dechloromonas* in the PAO group and *Deffluvicoccus* in the GAO group. Furthermore, no
333 significant seasonal variations were observed in the overall microbial community of each plant from winter to
334 summer samples. Additionally, our findings revealed the extent of the problem of bulking sludge for industrial
335 plants with 59% of the samples presenting settleability problems, from which 72% corresponded to FBS and 28%
336 to VBS.

337 The complexity of the industrial wastewater seems to depend to a large extent on its sector of origin, e.g.
338 wastewater originating from breweries are characterized by high COD concentrations but lack nutrients as

339 previously reported (Stes et al. 2018). Similarly, wastewaters originating from the vegetable industry were rich in
340 COD, P and N (Dobbeleers et al. 2017). The wastewaters from the dairy sector contained high COD and N
341 concentrations, but lacked P. The chemical sector was characterized by high salinity wastewaters, with an average
342 electrical conductivity of 16 mS.cm⁻¹, and a maximum up to 50 mS.cm⁻¹. Among the various ions that contributed
343 to the high salinity, chlorides (up to 20 g.L⁻¹) and sulfates (up to 5 g.L⁻¹) stand out. However, there was also a
344 significant difference in the composition of the wastewater coming from the same sector, which may be the result
345 of the different industrial processes that lead to a wide range of pollutants.

346 **Industrial plants revealed unique bacterial communities**

347 Microbial community analysis revealed that the most abundant phylum were *Proteobacteria*, *Bacteroidota*,
348 *Chloroflexota* and *Acidobacteriota*, these findings are in line with previous studies (Ibarbalz et al. 2013;
349 Selvarajan et al. 2018; Wang et al. 2020; Kristensen et al. 2021). However, there are still several abundant phyla
350 that have not yet been described within AS systems, including the *Myxococcota*, *Patescibacteria*,
351 *Gemmatimonadota* and *Latescibacterota*. At the family taxonomic classification, *Rhodocyclaceae*
352 (*Proteobacteria*) and *Saprospiraceae* (*Bacteroidota*) were highly abundant, suggesting an important role in the
353 industrial AS plants. The genus *Thauera* was the most abundant representative of the family *Rhodocyclaceae*
354 reaching up to 51% read abundance in the chemical sector. The high abundance of *Thauera* can be attributed to
355 the versatility of substrates that can be used, including aromatic compounds, amino acids and organic substrates
356 (Thomsen et al. 2007). The abundant representatives found for the family *Saprospiraceae* include *midas_g_6* and
357 *Ca. Epiflobacter*, which have been described to possess a broad metabolic potential, including the degradation of
358 proteins, polysaccharides and complex molecules (Kondrotaitė et al. 2022), which may give them an advantage in
359 industrial plants, resulting in high abundance.

360 At a higher rank such as genus and species classification, there were several abundant microorganisms for
361 which no ecophysiological information in AS systems exists (see Figure S2), including *Denitromonas* highly
362 present in the chemical sector, *Ferruginibacter*, *Terrimonas* and Ellin6067 in the tank truck cleaning sector.
363 Ellin6067 is an uncultured genus of the *Nitrosomonadaceae* family, and potential nitrogen-transforming bacteria
364 present in both natural water bodies (Mankiewicz-Boczek and Font-Nájera 2022) and in AS systems (Dueholm et
365 al. 2022). A loose core consisting of 20 microorganisms, most of them undescribed, was also found. These results
366 are in line with those of previous industrial studies (Ibarbalz et al. 2013), suggesting that the bacterial populations
367 across industrial plants differ considerably from each other, resulting in a unique composition of microbial
368 assembles that share only few genera as core community. Additionally, the alpha diversity analysis revealed a
369 much lower richness and diversity (see Figure S5) than those from municipal plants (Wang et al. 2016). This may
370 be because industrial waters contain less variety of contaminants which in turn will select for a more specific
371 microflora as previously reported in the MiDAS project (Dueholm et al. 2022). These results seem to be consistent
372 with the PCoA analysis, that revealed a very specific microbial community even within the same industrial sector,
373 indicating lower degree of microbiome similarity among all industrial plants.

374 **Functional bacteria in industrial plants**

375 The presence of functional guilds among the different treatment processes revealed that *Zoogloea* and *Thauera*
376 were highly abundant denitrifiers in C and C,N removal plants. *Zoogloea* was abundant within C removal plants,
377 where no biological nitrogen removal occurred, suggesting that their presence was not related to their denitrifying
378 activity. *Thauera* was the dominant denitrifier within C,N removal plants. The distribution of denitrifying
379 organisms in our study appears to be distinct from the global MiDAS study that includes mostly municipal plants,
380 where *Rodophtherax* are the main denitrifiers (Dueholm et al. 2022). Within the PAO phylotype, *Dechloromonas*
381 was the most abundant representative of this group. Interestingly, the presence of recognized PAOs does not
382 exceed 2% read abundance in our surveyed Bio-P plants. The low presence of recognized PAOs genera can be
383 explained as a result of the low Bio-P activity in the mentioned industrial plants. The biological P removal activity
384 differs from that of EBPR plants. The mentioned plants (3) reported lower bio-P release (15-20 mgP.L⁻¹) and much
385 higher COD-P ratio (40-60) than conventional EBPR systems. The putative PAO *Thiothrix* (Rubio-Rincón et al.
386 2017; Mardanov et al. 2020) was abundant in the bio-P plants, suggesting that the mixotrophic metabolism of
387 *Thiothrix* including sulfur oxidation and biological P removal may play an important role within industrial plants,
388 possibly contributing to the biological P removal. In addition, only a very small number of industrial plants had

389 biological P removal, since chemical removal was used for the most part. Furthermore, the configuration of the
390 industrial plants, mostly CAS type, was not suitable for the growth of PAOs resulting in a low overall abundance
391 of this group of bacteria among all the industrial plants.

392 GAO representatives were found abundant in one C-removal plant from the tank-truck cleaning sector, and in
393 nutrient removal plants. This can be attributed to the unaerated feeding strategy implemented in the tank-truck
394 cleaning plant (Caluwé et al. 2022), or by the presence of anaerobic/anoxic zones in the latter. *Defluviococcus*
395 (*Alphaproteobacteria*) and *Ca. Competibacter* (*Gammaproteobacteria*) were the most abundant genera.
396 *Defluviococcus* possess a broad substrate affinity including acetate, propionate, pyruvate and glucose, and have
397 been described as potential carbon competitors of PAOs and other GAOs (Burow et al. 2007). However, our results
398 show that these bacteria cohabit the same system, suggesting that there is no such competition for substrate. The
399 well-known GAO, *Ca. Competibacter* (McIlroy et al. 2013) was found among the top 10 most abundant
400 microorganisms (Fig. 1) and in more than 50% of the samples with read abundance higher than 0.1% and also
401 comprising the general core community in the studied industrial plants.

402 The efficiency and performance of a WWTP depends primarily on the composition and activity of its microbial
403 community. The functional community described in this study aligns with established bacterial guilds, including
404 PAOs, GAOs, and denitrifiers responsible for nutrients removal. The bulking community commonly consists of
405 filaments, which not only have a structural role, but also aid in nutrient removal in the AS system. Our findings
406 indicate that the abundance of *Thiothrix* and *Ca. Sarcinithrix* is influenced by process design, with higher
407 abundance found in complex systems designed for N and/or P removal. This aligns with existing literature
408 regarding the *Chloroflexota* phylum filaments (Petriglieri et al. 2023). The presence of certain filaments in nutrient
409 removal plants can be attributed to their substrate storage capacity, giving them similar advantage as for well-
410 known nutrient removal bacteria. Previous studies have reported the storage capacity of filaments including lipids
411 and polyphosphate storage by *Ca. M. Parvicella* (Jon McIlroy et al. 2013), sulfur and polyphosphate by *Thiothrix*
412 (Rubio-Rincón et al. 2017) and glycogen by *Chloroflexota* filaments (Petriglieri et al. 2023). However, the precise
413 contribution of these filaments to nutrient removal in industrial WWTPs is not yet established, and therefore further
414 studies are necessary to determine their functional role.

415 **Settling properties and microbial community**

416 Bulking sludge was observed in 55% of the plants in winter and 62% in the summer. These findings differ
417 from previous studies where bulking is commonly observed in the winter season (V.Tandoi et al. 2006). These
418 differences can be explained as a result of industrial or municipal wastewaters, whereas municipal plants suffer
419 from overgrowth of the filament *Ca. M. Parvicella*, which was not found in the industrial plants studied.
420 Additionally, our study only evaluated samples from winter (November-December) and summer (June-August) in
421 Belgium with a different temperature range than previous studies, more samples should be considered to evaluate
422 in detail the seasonal effect on bulking sludge.

423 The majority of bulking cases corresponded to FBS (72%), our findings suggest that despite the low abundance
424 (<5%) of total filaments, FBS can still occur. The low abundance of total filaments can be attributed to different
425 biases occurring in the amplicon sequencing such as gene copy number, targeting region of the 16S rRNA gene or
426 DNA extraction biases. Furthermore, the samples considered as FBS presented a high abundance of filaments
427 (FI>3) observed microscopically, suggesting that the special distribution of filaments in the flocs rather than
428 abundance affects the sludge settling properties, as previously described (Wágner et al. 2015). The phylum
429 *Chloroflexota* was the third most encountered phylum, and it has recently been reported that almost all of its
430 species have a filamentous morphology (Petriglieri et al. 2023). *Kouleothrix*, *Ca. Villigracilis* and *Ca. Sarcinithrix*
431 were the most abundant filamentous genera from *Chloroflexota*. *Kouleothrix* and *Ca. Sarcinithrix* were observed
432 extended from the flocs, causing bulking and contributing to an increased FI (Fig. 5b). The negative effect of
433 *Kouleothrix* on sludge settleability in municipal nutrient removal plants have been associated to low oxygen
434 concentration in the aeration tank (<1.1 mg O₂.L⁻¹) and the low temperatures being able to grow even at 7°C
435 (Nittami et al. 2020). However, seasonality between winter and summer samples was not a statistically significant
436 factor influencing the abundance of *Kouleothrix*, based on Kruskal-Wallis analysis (see Table S3). *Ca. Sarcinithrix*
437 abundance was significantly correlated with plant configuration, being highly abundant in more complex plant
438 configurations such as nutrient removal, which may be due to their previously reported facultative anaerobic

439 metabolism (Nierychlo et al. 2019). *Ca. Sarcinatrix* is a recurrent filament in municipal Danish plants (Nierychlo
440 et al. 2019) and also found in domestic plant operating under long sludge age (McIlroy et al. 2011).

441 Amplicon sequencing also revealed the presence of described FB in well settling samples, with total filaments
442 accounting for up to 11.3% read abundance, with *Leptothrix*, *Ca. Villigracilis* and *Trichococcus* genera being
443 abundant. The genus *Trichococcus* has been reported to present variable cell morphology in AS plants, growing
444 as single cell or filament (Nierychlo et al. 2020), while *Leptothrix* have also presented single cell morphology in
445 AS systems (Wagner et al. 1994). Our findings suggest that *Leptothrix* did not presented filamentous morphology
446 extended from the flocs and did not contribute to bulking, we suggest that it did not present filamentous
447 morphology at all or it was filamentous but only inside the flocs, not being able to classify using conventional
448 microscopy. The morphology of *Ca. Villigracilis* has been confirmed in situ, found predominantly located within
449 the flocs, suggesting their structural role in well-settling sludge (Nierychlo et al. 2019). In the global MiDAS
450 survey (Dueholm et al. 2022), *Leptothrix* and *Ca. Villigracilis* were found to be the second and fourth most
451 abundant FB worldwide and are also part of the general core (with >0.1% RA in 50% of all plants) of AS systems.
452 However, these bacteria were not part of the general core taxa of the industrial plants studied in this survey. From
453 our results *Ca. Villigracilis* and *Leptothrix* were not microscopically observed but were found to have high read
454 abundance based on amplicon sequencing results. These filaments appear to play a structural role in industrial AS
455 systems, helping to maintain a well-settling sludge. In addition, these results suggest that when filaments are not
456 observed microscopically, it does not mean that their abundance is low.

457 According to our results, unknown filaments were found which were observed microscopically but amplicon
458 sequencing results did not reveal the presence of known filaments. On the one hand these findings can be explained
459 by the non-amplification of the filamentous bacteria by 16S RNA amplicon sequencing. The latter may be due to
460 the targeted region of the 16S rRNA gene (V1-3) which may not be suitable for this organism or due to bias in the
461 DNA extraction. On the other hand, it is possible that filamentous bacteria were amplified and found to be abundant
462 but not classified as filamentous because their morphology has not been reported in the literature. It is therefore
463 important to combine different identification techniques, such as molecular and morphological approaches.
464 Additionally, our results revealed a different abundance of FB than that found in municipal plants, a good example
465 of this is the absence of *Ca. Microthrix* in the studied industrial plants. The excessive growth *Ca. Microthrix* is
466 commonly associated with bulking and foaming problems in municipal treatment systems (Nierychlo et al., 2020,
467 Dueholm et al., 2022). However our results do not reveal a high abundance of *Ca. Microthrix* in the surveyed
468 industrial plants, which was encountered in low abundance (0.002-0.4% read abundance) within the meat and dairy
469 sector. Additionally, *Ca. Microthrix* is one of the easiest filament to identify using conventional light microscopy,
470 and it was seldomly observed within the mentioned samples.

471 VBS accounted for 28% of the bulking cases, observed in C (4) and C,N,DN (3) process type plants operating
472 under food to microorganisms ratios ranging from 0.05 to 0.18 g COD/g MLSS.day (mean and median 0.12), the
473 majority of the affected plants were configured as continuous flow feeding systems. This resulted in VBS plants
474 enriched in genera belonging to the phyla *Proteobacteria* and *Bacteroidota*. The *Rhodocyclaceae* (*Proteobacteria*)
475 family, including *Zoogloea* and *Thauera* were abundant in VBS samples (Fig. 4b), *Zoogloea* spp. such as *Z.*
476 *ramigera* have been widely associated with VBS due to their excessive production of EPS (Rosselló-Mora et al.
477 1995), similarly to *Thauera* (Allen et al. 2004). Comparative genome analysis revealed that clusters of EPS
478 biosynthesis genes were found in the genomes of members of the *Rhodocyclaceae* family (An et al. 2016),
479 highlighting the importance of members of this family in VBS. Similarly, representatives of the *Bacteroidota*
480 phylum, including *Ferruginibacter* and *Terrimonas* were abundant in VBS samples, however the *in situ*
481 physiology of these and other VBS genera has not been described. The identification of the abundant
482 microorganisms found in VBS samples from industrial AS plants will serve a basis for further study of their role
483 in EPS production and bulking.

484 Several approaches exist to control AS settling properties, with nonspecific control methods such as the
485 addition of chemicals and disinfectants, while specific methods include adjustment of wastewater composition,
486 selectors and feeding strategy. Wastewater composition significantly influences the microbial community of AS
487 systems, however many industrial effluents can be deficient in macro and micronutrients, leading to FBS and VBS,
488 therefore a BOD₅:N:P ratio of 100:5:1 is recommended (Richard et al., 2003). Additionally, in certain scenarios,

489 modifying the composition of wastewater to avoid FBS achievable by employing pretreatment techniques,
490 including Dissolved Air Flotation (DAF) or a fat and grease trap system. These methods aid in eliminating lipids
491 and preventing bulking caused by Actinobacteria filaments. In other cases, however, where the effluent consists
492 of diverse substrates that are difficult to modify, addressing bulking issues proves challenging and requires
493 consideration of additional factors. When it comes to dealing with filaments in wastewater from the petrochemical,
494 tank truck cleaning and brewing industries, switching to an anaerobic feeding has proven to be successful strategy
495 (Caluwé et al. 2017; Stes et al. 2018; Poelmans et al. 2023). Therefore, various operational factors must be taken
496 into account when addressing bulking in industrial WWTPs.

497 This study provides a deeper insight into the microbial community composition of full-scale industrial
498 WWTPs. With focus on functional groups of bacteria responsible for nutrient removal, and microorganisms
499 involved in bulking cases. No common core taxa were found across all the plants, but a general core of mostly
500 undescribed bacteria was found, suggesting that industrial plants have unique microbial communities derived from
501 their specific wastewaters. These results highlight the need for future research to investigate the physiology of
502 important undescribed taxa in full-scale plants. Finally, this study contributes to the overall goal of understanding
503 the ecology of full-scale AS systems.

504 **Supplementary data**

505 Supplementary data associated with this article can be found in the online version of this paper.

506 **Author contribution** TD and JD conceived and designed the research. KS conducted the experiments, data
507 analysis and wrote manuscript. TD and JD revised and edited the manuscript. All authors read and approved the
508 manuscript.

509 **Data availability** The datasets generated during this current study are available from the corresponding author on
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514 **Compliance with Ethical Standards**

515
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521

522 **References**

- 523 Albertsen M, Karst SM, Ziegler AS, Kirkegaard RH, Nielsen PH (2015) Back to basics - The influence of DNA
524 extraction and primer choice on phylogenetic analysis of activated sludge communities. *PLoS One* 10(7).
525 <https://doi.org/10.1371/journal.pone.0132783>
- 526 Allen MS, Welch KT, Prebyl BS, Baker DC, Meyers AJ, Sayler GS (2004) Analysis and glycosyl composition
527 of the exopolysaccharide isolated from the floc-forming wastewater bacterium *Thauera* sp. MZ1T. *Environ*
528 *Microbiol* 6(8):780–790. <https://doi.org/10.1111/J.1462-2920.2004.00615.X>
- 529 An W, Guo F, Song Y, Gao N, Bai S, Dai J, Wei H, Zhang L, Yu D, Xia M, Yu Y, Qi M, Tian C, Chen H, Wu
530 Z, Zhang T, Qiu D (2016) Comparative genomics analyses on EPS biosynthesis genes required for floc
531 formation of *Zoogloea resiniphila* and other activated sludge bacteria. *Water Res* 102:494–504.
532 <https://doi.org/10.1016/J.WATRES.2016.06.058>
- 533 Andersen K, Kirkegaard R, Karst S, Albertsen M (2018) ampvis2: an R package to analyse and visualise 16S
534 rRNA amplicon data. <https://doi.org/10.1101/299537>
- 535 APHA (1998) Standard Methods for the Examination of Water and Wastewater American Public Health
536 Association, 20th editi. Washington, DC, USA
- 537 Burow LC, Kong Y, Nielsen JL, Blackall LL, Nielsen PH (2007) Abundance and ecophysiology of
538 *Defluviicoccus* spp., glycogen-accumulating organisms in full-scale wastewater treatment processes.
539 *Microbiology* 153(1):178–185. <https://doi.org/10.1099/mic.0.2006/001032-0>

540 Caluwé M, Dobbeleers T, Daens D, Blust R, Geuens L, Dries J (2017) The effect of the feeding pattern of
541 complex industrial wastewater on activated sludge characteristics and the chemical and ecotoxicological
542 effluent quality. *Environ Sci Pollut Res* 24(11):10796–10807. <https://doi.org/10.1007/s11356-017-8712-3>

543 Caluwé M, Goossens K, Suazo KS, Tsertou E, Dries J (2022) Granulation strategies applied to industrial
544 wastewater treatment: from lab to full-scale. *Water Sci Technol* 85(9):2761–2771.
545 <https://doi.org/10.2166/wst.2022.129>

546 Chernousova E, Gridneva E, Grabovich M, Dubinina G, Akimov V, Rossetti S, Kuever J (2009) *Thiothrix*
547 *caldifontis* sp. nov. and *Thiothrix lacustris* sp. nov., gammaproteobacteria isolated from sulfide springs. *Int*
548 *J Syst Evol Microbiol* 59(12):3128–3135. <https://doi.org/10.1099/ijs.0.009456-0>

549 Cornelissen R, Van Dyck T, Dries J, Ockier P, Smets I, Van Den Broeck R, Van Hulle S, Feyaerts M (2018)
550 Application of online instrumentation in industrial wastewater treatment plants - A survey in Flanders,
551 Belgium. *Water Sci Technol* 78(4):957–967. <https://doi.org/10.2166/wst.2018.375>

552 Daims H, Lebedeva E V., Pjevac P, Han P, Herbold C, Albertsen M, Jehmlich N, Palatinszky M, Vierheilig J,
553 Bulaev A, Kirkegaard RH, Von Bergen M, Rattei T, Bendinger B, Nielsen PH, Wagner M (2015)
554 Complete nitrification by Nitrospira bacteria. *Nature* 528(7583):504–509.
555 <https://doi.org/10.1038/nature16461>

556 de Graaff DR, van Loosdrecht MCM, Pronk M (2020) Stable granulation of seawater-adapted aerobic granular
557 sludge with filamentous *Thiothrix* bacteria. *Water Res* 175:115683.
558 <https://doi.org/10.1016/j.watres.2020.115683>

559 Dobbeleers T, Daens D, Miele S, D’aes J, Caluwé M, Geuens L, Dries J (2017) Performance of aerobic nitrite
560 granules treating an anaerobic pre-treated wastewater originating from the potato industry. *Bioresour*
561 *Technol* 226:211–219. <https://doi.org/10.1016/j.biortech.2016.11.117>

562 Dueholm MKD, Nierychlo M, Andersen KS, Rudkjøbing V, Knutsson S, Nielsen PH (2022) MiDAS 4: A global
563 catalogue of full-length 16S rRNA gene sequences and taxonomy for studies of bacterial communities in
564 wastewater treatment plants. *Nat Commun* 13(1). <https://doi.org/10.1038/s41467-022-29438-7>

565 Dueholm MS, Andersen KS, McIlroy SJ, Kristensen JM, Yashiro E, Karst SM, Albertsen M, Nielsen PH (2020)
566 Generation of comprehensive ecosystem-specific reference databases with species-level resolution by
567 high-throughput full-length 16s rRNA gene sequencing and automated taxonomy assignment (Autotax).
568 *MBio* 11(5):1–14. <https://doi.org/10.1128/mBio.01557-20>

569 Eikelboom D (2000) *Process Control of Activated Sludge Plants by Microscopic Investigation*

570 FAO (2015) *Food and Agriculture Organization of the United Nations*.
571 <https://www.fao.org/aquastat/en/overview/methodology/water-use>

572 Gómez-Rubio V (2017) ggplot2 - Elegant Graphics for Data Analysis (2nd Edition). *J Stat Softw* 77(Book
573 Review 2). <https://doi.org/10.18637/jss.v077.b02>

574 Han X, Zhou Z, Mei X, Ma Y, Xie Z (2018) Influence of fermentation liquid from waste activated sludge on
575 anoxic/oxic- membrane bioreactor performance: Nitrogen removal, membrane fouling and microbial
576 community. *Bioresour Technol* 250(December 2017):699–707.
577 <https://doi.org/10.1016/j.biortech.2017.11.090>

578 Ibarbalz FM, Figuerola ELM, Erijman L (2013) Industrial activated sludge exhibit unique bacterial community
579 composition at high taxonomic ranks. *Water Res* 47(11):3854–3864.
580 <https://doi.org/10.1016/j.watres.2013.04.010>

581 Jon McIlroy S, Kristiansen R, Albertsen M, Michael Karst S, Rossetti S, Lund Nielsen J, Tandoi V, James
582 Seviour R, Nielsen PH (2013) Metabolic model for the filamentous ‘*Candidatus Microthrix parvicella*’
583 based on genomic and metagenomic analyses. *ISME J* 7(6):1161–1172.
584 <https://doi.org/10.1038/ismej.2013.6>

585 Kaetzke A, Jentsch D, Eschrich K (2005) Quantification of *Microthrix parvicella* in activated sludge bacterial
586 communities by real-time PCR. *Lett Appl Microbiol* 40(3):207–211. <https://doi.org/10.1111/j.1472-765X.2005.01656.x>

587

588 Kondrotaitė Z, Valk LC, Petriglieri F, Singleton C, Nierychlo M, Dueholm MKD, Nielsen PH (2022) Diversity
589 and Ecophysiology of the Genus OLB8 and Other Abundant Uncultured Saprospiraceae Genera in Global
590 Wastewater Treatment Systems. *Front Microbiol* 13(July):1–15.
591 <https://doi.org/10.3389/fmicb.2022.917553>

592 Kristensen JM, Singleton C, Clegg LA, Petriglieri F, Nielsen PH (2021) High Diversity and Functional Potential
593 of Undescribed “Acidobacteriota” in Danish Wastewater Treatment Plants. *Front Microbiol* 12.
594 <https://doi.org/10.3389/fmicb.2021.643950>

595 Lajoie CA, Layton AC, Gregory IR, Sayler GS, Taylor DE, Meyers AJ (2000) Zooglear Clusters and Sludge
596 Dewatering Potential in an Industrial Activated-Sludge Wastewater Treatment Plant. *Water Environ Res*
597 72(1):56–64. <https://doi.org/10.2175/106143000x137112>

598 Legendre P, Gallagher ED (2001) Ecologically meaningful transformations for ordination of species data.
599 2001(September 2000):271–280. <https://doi.org/10.1007/s004420100716>

600 Levantesi C, Beimfohr C, Geurkink B, Rossetti S, Thelen K, Krooneman J, Snaird J, Van Der Waarde J, Tandoi
601 V (2004) Filamentous Alphaproteobacteria associated with bulking in industrial wastewater treatment
602 plants. *Syst Appl Microbiol*. <https://doi.org/10.1078/0723202042369974>

603 Mankiewicz-Boczek J, Font-Nájera A (2022) Temporal and functional interrelationships between
604 bacterioplankton communities and the development of a toxigenic *Microcystis* bloom in a lowland
605 European reservoir. *Sci Rep* 12(1):19332. <https://doi.org/10.1038/s41598-022-23671-2>

606 Mardanov A V., Gruzdev E V., Smolyakov DD, Rudenko TS, Beletsky A V., Gureeva M V., Markov ND,
607 Berestovskaya YY, Pimenov N V., Ravin N V., Grabovich MY (2020) Genomic and Metabolic Insights
608 into Two Novel *Thiothrix* Species from Enhanced Biological Phosphorus Removal Systems.
609 *Microorganisms* 8(12):1–13. <https://doi.org/10.3390/MICROORGANISMS8122030>

610 Martins AMP, Heijnen JJ, Van Loosdrecht MCM (2003) Effect of feeding pattern and storage on the sludge
611 settleability under aerobic conditions. *Water Res* 37(11):2555–2570. [https://doi.org/10.1016/S0043-1354\(03\)00070-8](https://doi.org/10.1016/S0043-1354(03)00070-8)

612 Martins AMP, Pagilla K, Heijnen JJ, Van Loosdrecht MCM (2004) Filamentous bulking sludge - A critical
613 review. *Water Res* 38(4):793–817. <https://doi.org/10.1016/j.watres.2003.11.005>

614 Matsuura N, Masakke Y, Karthikeyan S, Kanazawa S, Honda R, Yamamoto-Ikemoto R, Konstantinidis KT
615 (2021) Metagenomic insights into the effect of sulfate on enhanced biological phosphorus removal. *Appl*
616 *Microbiol Biotechnol* 105(5):2181–2193. <https://doi.org/10.1007/s00253-021-11113-4>

617 McIlroy SJ, Albertsen M, Andresen EK, Saunders AM, Kristiansen R, Stokholm-Bjerregaard M, Nielsen KL,
618 Nielsen PH (2013) ‘Candidatus *Competibacter*’-lineage genomes retrieved from metagenomes reveal
619 functional metabolic diversity. *ISME J* 2014 83 8(3):613–624. <https://doi.org/10.1038/ismej.2013.162>

620 McIlroy SJ, Speirs LBM, Tucci J, Seviour RJ (2011) In situ profiling of microbial communities in full-scale
621 aerobic sequencing batch reactors treating winery waste in Australia. *Environ Sci Technol* 45(20):8794–
622 8803. <https://doi.org/10.1021/es2018576>

623 Nierychlo M, McIlroy SJ, Kucheryavskiy S, Jiang C, Ziegler AS, Kondrotaitė Z, Stokholm-Bjerregaard M,
624 Nielsen PH (2020) *Candidatus Amarolinea* and *Candidatus Microthrix* Are Mainly Responsible for
625 Filamentous Bulking in Danish Municipal Wastewater Treatment Plants. *Front Microbiol* 11(June):1–17.
626 <https://doi.org/10.3389/fmicb.2020.01214>

627 Nierychlo M, Miłobędzka A, Petriglieri F, McIlroy B, Nielsen PH, McIlroy SJ (2019) The morphology and
628 metabolic potential of the Chloroflexi in full-scale activated sludge wastewater treatment plants. *FEMS*
629 *Microbiol Ecol* 95(2):1–11. <https://doi.org/10.1093/femsec/fiy228>

630 Nittami T, Kasakura R, Kobayashi T, Suzuki K, Koshiya Y, Fukuda J, Takeda M, Tobino T, Kurisu F, Rice D,
631 Petrovski S, Seviour RJ (2020) Exploring the operating factors controlling *Kouleothrix* (type 1851), the
632 dominant filamentous bacterial population, in a full-scale A2O plant. *Sci Rep* 10(1):1–10.
633 <https://doi.org/10.1038/s41598-020-63534-2>

634 Oppong D, King VM, Bowen JA (2003) Isolation and characterization of filamentous bacteria from paper mill
635 slimes. *Int Biodeterior Biodegrad* 52(2):53–62. [https://doi.org/10.1016/S0964-8305\(02\)00174-9](https://doi.org/10.1016/S0964-8305(02)00174-9)

636 Peng Y, Gao C, Wang S, Ozaki M, Takigawa A (2003) Non-filamentous sludge bulking caused by a deficiency
637 of nitrogen in industrial wastewater treatment. *Water Sci Technol* 47(11):289–295.
638 <https://doi.org/10.2166/wst.2003.0617>

639 Petriglieri F, Kondrotaitė Z, Singleton C, Nierychlo M, Dueholm MKD, Nielsen PH (2023) A comprehensive
640 overview of the Chloroflexota community in wastewater treatment plants worldwide. *bioRxiv*
641 :2023.06.26.546502. <https://doi.org/10.1101/2023.06.26.546502>

642 Petriglieri F, Singleton CM, Kondrotaitė Z, Dueholm MKD, McDaniel EA, McMahon KD, Nielsen PH (2022)
643 Reevaluation of the Phylogenetic Diversity and Global Distribution of the Genus “*Candidatus*
644 *Accumulibacter*”. *mSystems* 7(3). https://doi.org/10.1128/MSYSTEMS.00016-22/SUPPL_FILE/MSYSTEMS.00016-22-S0010.DOCX

645 Poelmans S, Dockx L, Seguel Suazo K, Goettert D, Dries J (2023) Implementation of an anaerobic selector step
646 for the densification of activated sludge treating high-salinity petrochemical wastewater. *Water Sci*
647 *Technol* 87(4):823–833. <https://doi.org/10.2166/WST.2023.033>

648 Richard M, Richard G, Jenkins D (2003) *Manual on the Causes and Control of Activated Sludge Bulking,*
649 *Foaming, and Other Solids Separation Problems*, 3rd Edition. CRC Press

650 Rosselló-Mora RA, Wagner M, Amann R, Schleifer K-H (1995) The abundance of *Zoogloea ramigera* in sewage
651 treatment plants. *Appl Environ Microbiol* 61(2):702–707

652 Rubio-Rincón FJ, Welles L, Lopez-Vazquez CM, Nierychlo M, Abbas B, Geleijnse M, Nielsen PH, van
653 Loosdrecht MCM, Brdjanovic D (2017) Long-term effects of sulphide on the enhanced biological removal
654 of phosphorus: The symbiotic role of *Thiothrix caldifontis*. *Water Res* 116:53–64.
655 <https://doi.org/10.1016/j.watres.2017.03.017>

656 Selvarajan R, Sibanda T, Venkatachalam S, Kamika I, Nel WAJ (2018) Industrial wastewaters harbor a unique
657 diversity of bacterial communities revealed by high-throughput amplicon analysis. *Ann Microbiol*

660 68(7):445–458. <https://doi.org/10.1007/s13213-018-1349-8>

661 Seviour R (2010) *Microbial Ecology of Activated Sludge*

662 Shao Y, Zhang H, Buchanan I, Mohammed A, Liu Y (2019) Comparison of extracellular polymeric substance
663 (EPS) in nitrification and nitritation bioreactors. *Int Biodeterior Biodegradation* 143:104713.
664 <https://doi.org/10.1016/J.IBIOD.2019.06.001>

665 Singleton CM, Petriglieri F, Wasmund K, Nierychlo M, Kondrotaitė Z, Petersen JF, Peces M, Dueholm MS,
666 Wagner M, Nielsen PH (2022) The novel genus, ‘Candidatus Phosphoribacter’, previously identified as
667 Tetrasphaera, is the dominant polyphosphate accumulating lineage in EBPR wastewater treatment plants
668 worldwide. *ISME J* 16(6):1605–1616. <https://doi.org/10.1038/s41396-022-01212-z>

669 Stes H, Aerts S, Caluwé M, Dobbeleers T, Wuyts S, Kiekens F, D’Aes J, De Langhe P, Dries J (2018) Formation
670 of aerobic granular sludge and the influence of the pH on sludge characteristics in a SBR fed with
671 brewery/bottling plant wastewater. *Water Sci Technol* 77(9):2253–2264.
672 <https://doi.org/10.2166/wst.2018.132>

673 Stokholm-Bjerregaard M, McIlroy SJ, Nierychlo M, Karst SM, Albertsen M, Nielsen PH (2017) A critical
674 assessment of the microorganisms proposed to be important to enhanced biological phosphorus removal in
675 full-scale wastewater treatment systems. *Front Microbiol* 8(APR):1–18.
676 <https://doi.org/10.3389/fmicb.2017.00718>

677 Thomsen TR, Kong Y, Nielsen PH (2007) Ecophysiology of abundant denitrifying bacteria in activated sludge.
678 *FEMS Microbiol Ecol* 60(3):370–382. <https://doi.org/10.1111/j.1574-6941.2007.00309.x>

679 Tsertou E, Caluwé M, Goossens K, Dobbeleers T, Dockx L, Poelmans S, Suazo KS, Dries J (2022) Is building
680 up substrate during anaerobic feeding necessary for granulation? *Water Sci Technol* 86(4):763–776.
681 <https://doi.org/10.2166/wst.2022.236>

682 V.Tandoi, David J, Wanner J (2006) *Activated sludge separation problems*. IWA Publishing, London

683 van Loosdrecht MCM, Nielsen PH, Lopez-Vazquez CM, Brdjanovic D (2016) *Experimental Methods in*
684 *Wastewater Treatment*. In: *Water Intelligence Online*. IWA Publishing, pp 9781780404752–
685 9781780404752

686 Vervaeren H, De Wilde K, Matthys J, Boon N, Raskin L, Verstraete W (2005) Quantification of an Eikelboom
687 type 021N bulking event with fluorescence in situ hybridization and real-time PCR. *Appl Microbiol*
688 *Biotechnol* 68(5):695–704. <https://doi.org/10.1007/s00253-005-1963-9>

689 Wágner DS, Ramin E, Szabo P, Dechesne A, Plósz BG (2015) *Microthrix parvicella* abundance associates with
690 activated sludge settling velocity and rheology - Quantifying and modelling filamentous bulking. *Water*
691 *Res* 78:121–132. <https://doi.org/10.1016/j.watres.2015.04.003>

692 Wagner M, Erhart R, Manz W, Amann R, Lemmer H, Wedi D, Schleifer K, Munich D-, *Abfallwirtschaft W-*
693 *(1994) Development of an rRNA-Targeted Oligonucleotide Probe Specific for the Genus Acinetobacter*
694 *and Its Application for In Situ Monitoring in Activated Sludge*. (35):792–800

695 Wang P, Yu Z, Qi R, Zhang H (2016) Detailed comparison of bacterial communities during seasonal sludge
696 bulking in a municipal wastewater treatment plant. *Water Res* 105:157–166.
697 <https://doi.org/10.1016/j.watres.2016.08.050>

698 Wang Q, Liang J, Zhang S, Yoza BA, Li QX, Zhan Y, Ye H, Zhao P, Chen C (2020) Characteristics of bacterial
699 populations in an industrial scale petrochemical wastewater treatment plant: Composition, function and
700 their association with environmental factors. *Environ Res* 189.
701 <https://doi.org/10.1016/j.envres.2020.109939>

702 Zhang M, Yao J, Wang X, Hong Y, Chen Y (2019) The microbial community in filamentous bulking sludge
703 with the ultra-low sludge loading and long sludge retention time in oxidation ditch. *Sci Rep* 9(1):1–10.
704 <https://doi.org/10.1038/s41598-019-50086-3>

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716 **Tables and figures**

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718 **Table 1** Distribution of the industrial plants across the different sectors.

Sector	Food				Chemical	Textile	Tank	
	Meat	Vegetable	Brewery	Dairy			truck	BioIndustry*
Number of plants	9	2	3	3	5	2	4	3

719 * Bioindustry: Corresponds to 3 industrial plants processing soya beans, malt and rapeseed.

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726 **Table 2** Range of wastewater composition in different industrial sectors.

Industrial Sector	COD (mgO ₂ .L ⁻¹)	VFA (mg.L ⁻¹)	P total (mg.L ⁻¹)	N total (mg N.L ⁻¹)	S total (mgS.L ⁻¹)	EC (mS.cm ⁻¹)	pH	COD/N	COD/P
Meat	768-5200	578-1162	2-46	15-1090	1-2042	1.1-32	5.2-11.5	5-73	56-636
Vegetable	1992-4682	440-880	6.4-59	38-72	11-22	0.7-3.4	4.3-5.2	43-62	59-418
Brewery	2974-5756	853-1528	5-14	13-47	2-12	2.6-4	6.6-7.5	86-96	227-538
Dairy	894-2252	504-804	6.5-21	75-148	23-38	2.4-3.4	5.7-7.9	7-28	104-249
Bio-industry	702-11304	65-788	8-120	25-126	6-232	2.8-16	5.1-8.7	28-90	25-140
Chemical	561-3504	61-624	4.1-13	6-191	0-1113	1.9-50	6.7-9.5	8-250	176-421
Textile	250-1260	50-290	1-13	11-51	62-142	3.8-5	6,5-8	18-130	114-250
Tankcleaning	864-5120	974-1500	2-10	9-18	18-297	2-3.5	4.2-8.2	75-311	304-557

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737 **Figure captions:**

738 **Fig. 1.** a) Bar chart of 10 most abundant phyla across the 29 plants of 7 industrial sectors b) Heatmap
739 of 10 most abundant genera for each industrial plant, values were calculated as averages of read
740 abundance (%)

741 **Fig. 2.** a) Rank abundance plot presenting the cumulative read abundance across 56 activated sludge
742 samples, including mean and SD of genus-level ASVs b) Core community plot highlighting strict,
743 general and loose core community based on observation frequency c) Heatmap of the general core
744 community at genus-level ASVs across industrial sectors.

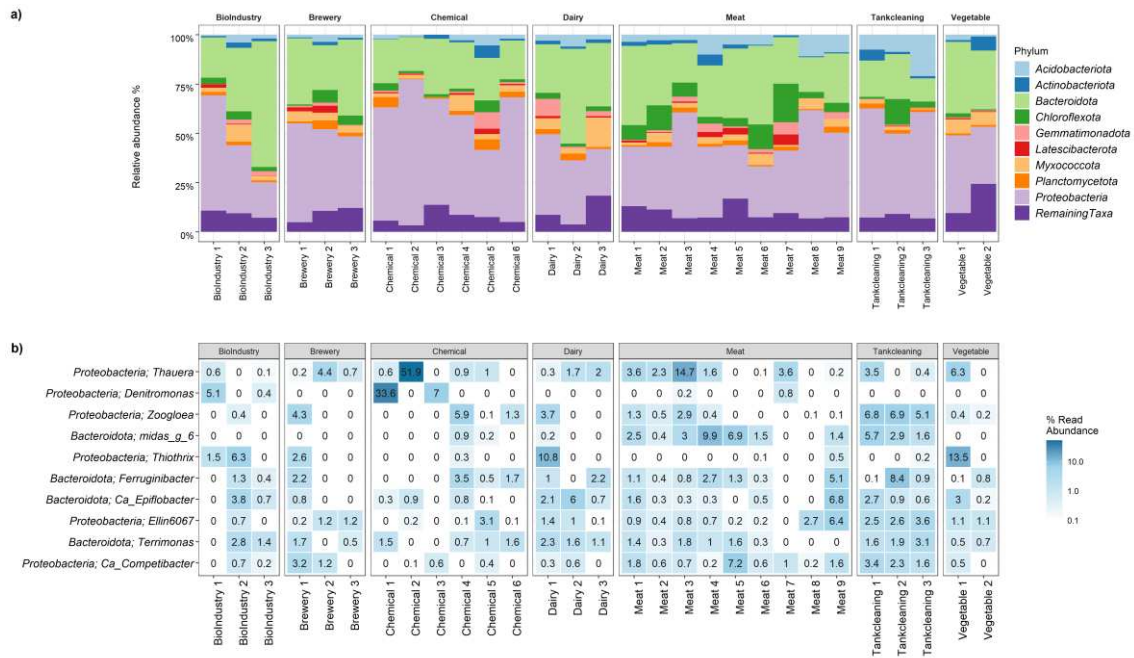
745 **Fig. 3.** Heatmap presenting the functional groups (at genus classification) along the different process
746 type and their respective number of plants. Results are the mean of read abundance.

747 **Fig. 4.** a) High-low plot describing settling properties (DSVI) and abundance of filaments (FI) in 58
748 samples (Range high-low and average are presented); b) Abundant genera in filamentous bulking
749 sludge (FBS) and viscous bulking sludge (VBS) samples

750 **Fig. 5.** a) Heatmap visualising the abundance of the top 10 filaments genera and phyla across the
751 industrial sectors, b) heatmap presenting most abundant FB distributed along FI and c) Boxplot
752 showing the total filaments abundance across industrial sectors

753 **Fig. 6.** a) CCA plot, the arrow length represents the strength of the correlation between the
754 environmental variables and the microbiome. The relative contribution (eigenvalue) of each axis to the
755 total inertia in the data as well as to the constrained, respectively, are indicated in percent at the axis
756 titles. b) PCoA of bacterial communities, the first two principal coordinates are plotted representing
757 10% of the total variation.

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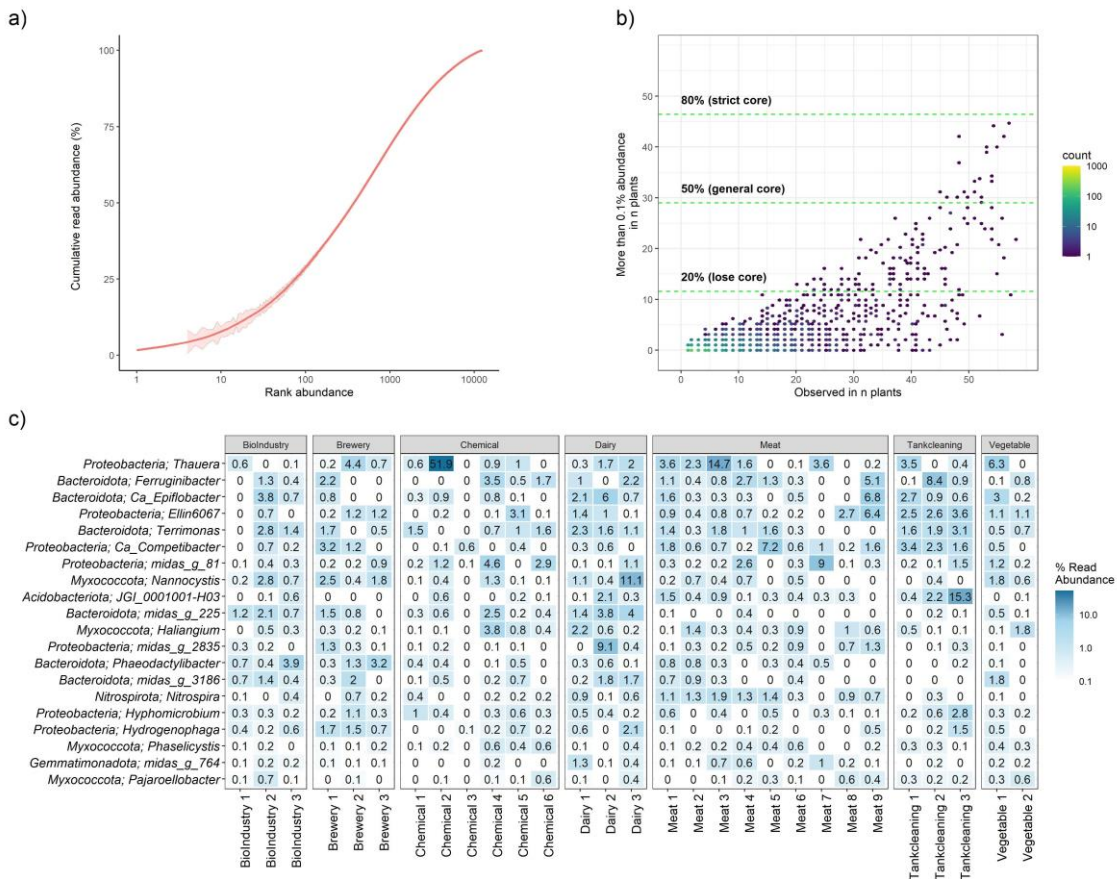


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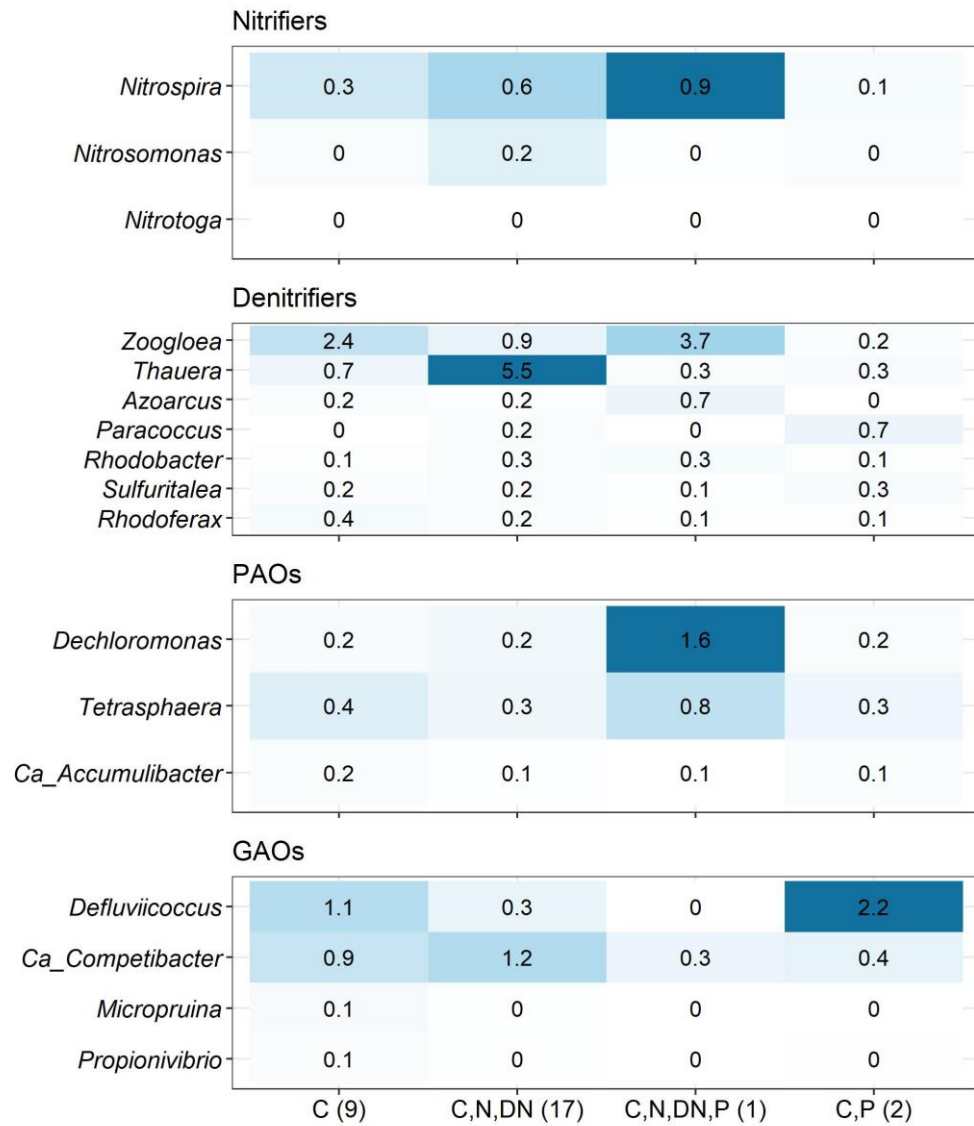
765 including mean and SD of genus-level ASVs b) Core community plot highlighting strict, general and loose core

766 community based on observation frequency c) Heatmap of the general core community at genus-level ASVs

767 across industrial sectors.

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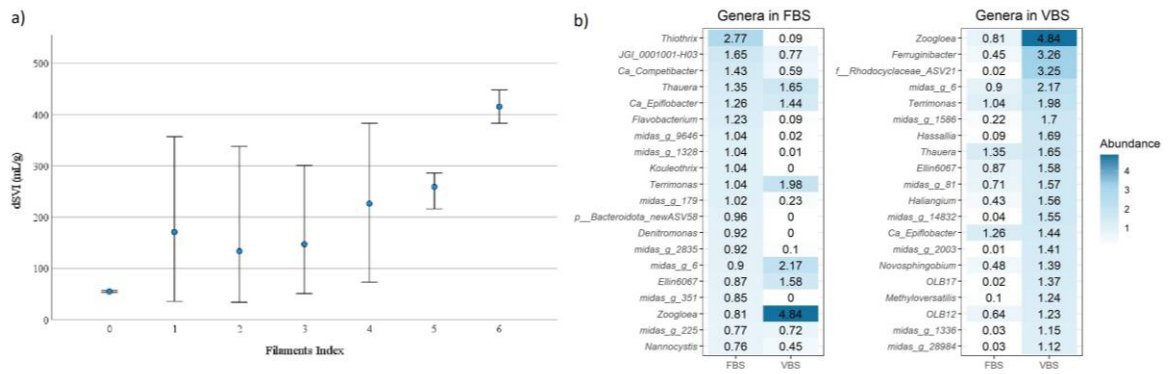


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772 their respective number of plants. Results are the mean of read abundance.

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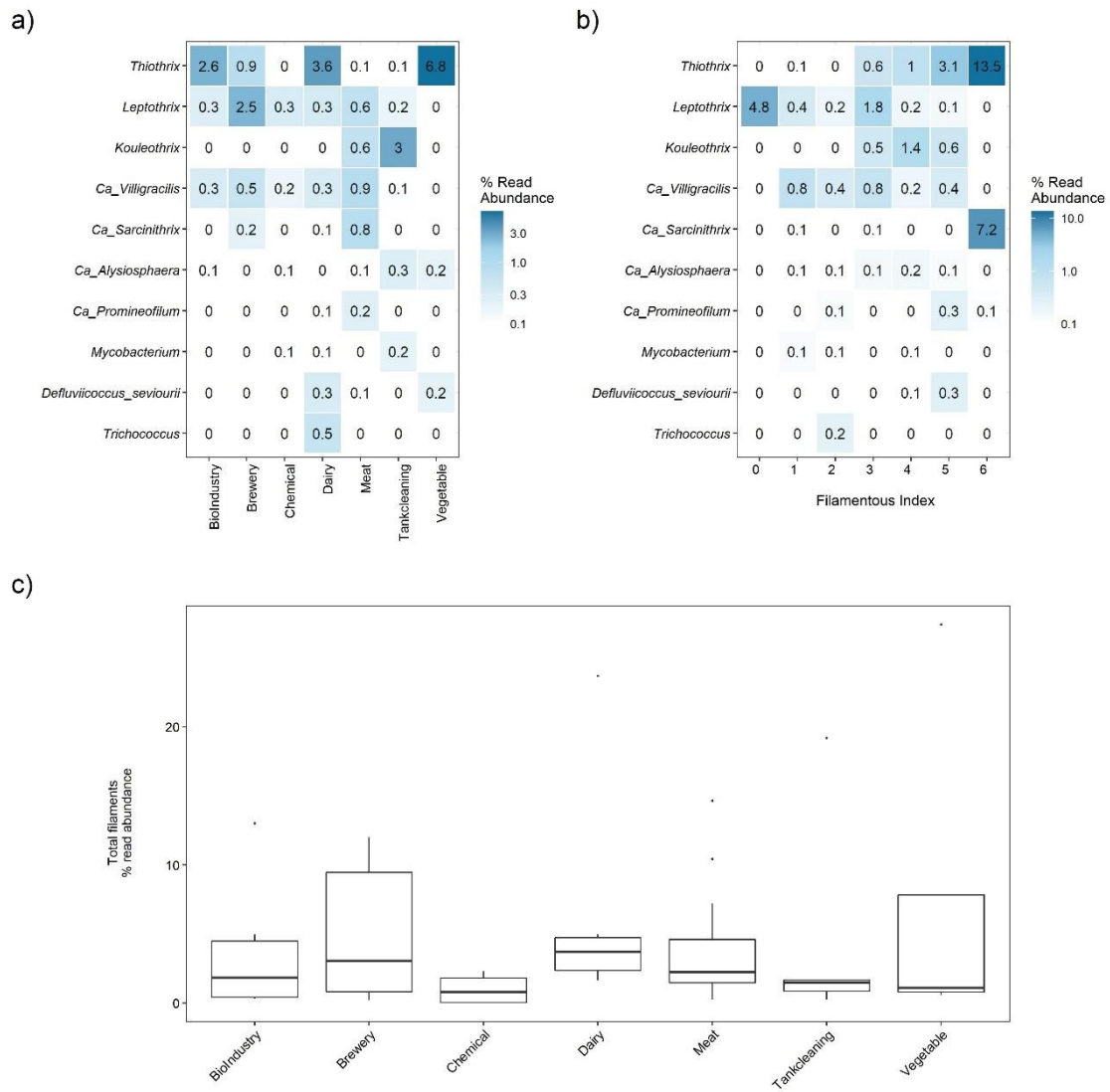


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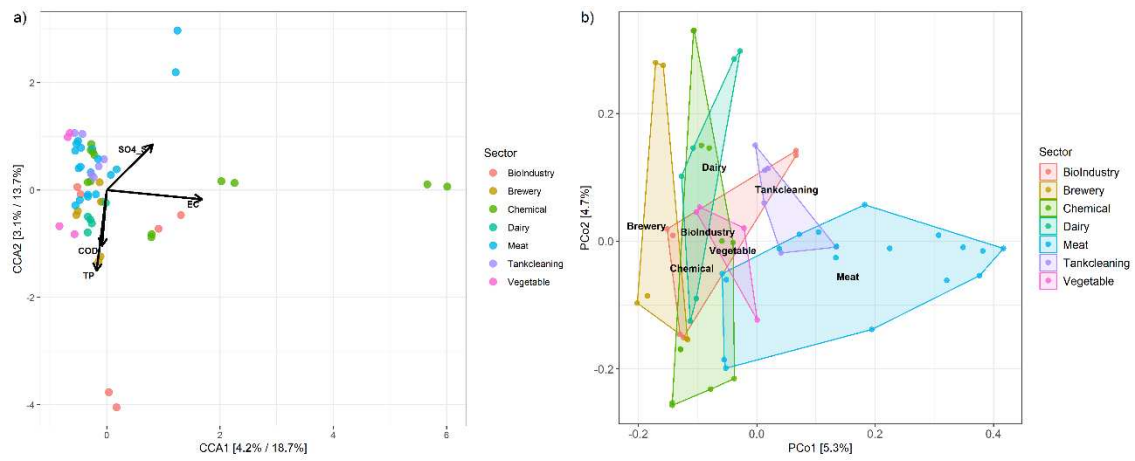
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 788 variables and the microbiome. The relative contribution (eigenvalue) of each axis to the total inertia in the data
 789 as well as to the constrained, respectively, are indicated in percent at the axis titles. b) PCoA of bacterial
 790 communities, the first two principal coordinates are plotted representing 10% of the total variation.

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