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1 **Can permanent grassland soils with elevated organic carbon buffer**  
2 **negative effects of more persistent precipitation regimes on forage**  
3 **grass performance?**

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17 Keywords: Soil organic carbon, land-use history, weather persistence, soil degradation,  
18 intensive grassland, forage quality

19 **Abstract**

20 Agricultural practices enhancing soil organic carbon (SOC) show potential to buffer negative effects of  
21 climate change on forage grass performance. We tested this by subjecting five forage grass varieties  
22 differing in fodder quality and drought/flooding resistance to increased persistence in summer  
23 precipitation regimes (PR) across sandy and sandy-loam soils from either permanent (high SOC) or  
24 temporary grasslands (low SOC) in adjacent parcels. Over the course of two consecutive summers,  
25 monoculture mesocosms were subjected to rainy/dry weather alternation either every 3 days or every  
26 30 days, whilst keeping total precipitation equal. Increased PR persistence induced species-specific  
27 drought damage and productivity declines. Soils from permanent grasslands with elevated SOC  
28 buffered plant quality, but buffering effects of SOC on drought damage, nutrient availability and yield  
29 differed between texture classes. In the more persistent PR, *Festuca arundinacea* FERMINA was the  
30 most productive species but had the lowest quality under both ample water supply and mild soil  
31 drought, while under the most intense soil droughts, *Festulolium* FESTILO maintained the highest  
32 yields. The hybrid *Lolium x boucheanum* kunth. MELCOMBI had intermediate productivity and both  
33 *Lolium perenne* varieties showed the lowest yields under soil drought, but the highest forage quality  
34 (especially the tetraploid variety MELFORCE). Performance varied with plant maturity stage and across  
35 seasons/years and was driven by altered water and nutrient availability and related nitrogen nutrition  
36 among species during drought and upon rewetting. Moreover, while permanent grassland soils  
37 showed the most consistent positive effects on plant performance, their available water capacity also  
38 declined under increased PR persistence. We conclude that permanent grassland soils with historically  
39 elevated SOC likely buffer negative effects of increasing summer weather persistence on forage grass  
40 performance, but may also be more sensitive to degradation under climate change.

## 41        **1. Introduction**

42        Climate change is threatening grassland ecosystem service provisioning (e.g., hay production, carbon  
43        storage) by rapidly altering rainfall patterns, atmospheric temperatures and dryness (De Boeck *et al.*,  
44        2010; Allan *et al.*, 2020) and the related soil water and nutrient regimes within these generally shallow-  
45        rooted ecosystems (Poirier *et al.*, 2012; Klaus *et al.*, 2020; Van Sundert *et al.*, 2020; Van Sundert *et al.*,  
46        2021; Reynaert *et al.*, 2022). For instance, hotter and more intense droughts cause declines in tiller  
47        survival, productivity and quality (Poirier *et al.*, 2012; Fariaszewska *et al.*, 2020), whilst simultaneously  
48        inducing more pronounced transient pulses of microbial activity and nutrient availability upon  
49        rewetting (Borken & Matzner, 2009; Van Sundert *et al.*, 2020) as well as increasing the risk of nutrient  
50        leaching and undesired eutrophication elsewhere (Loecke *et al.*, 2017; Klaus *et al.*, 2020). As such,  
51        altered rainfall variability may not only impact plants directly through altering soil water dynamics,  
52        but also indirectly by altering nutrient cycling, potentially benefiting species which can rapidly extract  
53        available resources during transient periods of abundance (Hofer *et al.*, 2017; Reynaert *et al.*, 2021;  
54        Van Sundert *et al.*, 2021).

55        Much research has focused on how more intense weather with longer, more extreme droughts and  
56        more concentrated rainfall may affect managed grassland ecosystem functioning (Knapp *et al.*, 2008;  
57        Grant *et al.*, 2014; Hofer *et al.*, 2017; Meisser *et al.*, 2019; Fariaszewska *et al.*, 2020; Van Sundert *et*  
58        *al.*, 2020; Van Sundert *et al.*, 2021) but less is known about the effects of recently observed increases  
59        in summer weather persistence in the mid-latitudes, i.e., the lengthening of both dry and wet spells  
60        compared to historic averages (Zolina *et al.*, 2013; Pfleiderer *et al.*, 2019; Felton *et al.*, 2021; Reynaert  
61        *et al.*, 2021; Reynaert *et al.*, 2022; Reynaert *et al.*, 2023a). The few experimental studies performed  
62        on unfertilized systems (Reynaert *et al.*, 2021; Reynaert *et al.*, 2022; Li *et al.*, 2023a; Reynaert *et al.*,  
63        2023b) generally indicate negative species-specific effects on productivity (up to – 50 %; see Li *et al.*  
64        (2023a)) related to differences in plant metabolism under drought (Zi *et al.*, 2023a), and both positive  
65        and negative effects on forage quality (Reynaert *et al.*, 2023b; Zi *et al.*, 2023b). However, plant growth

66 is disturbed more regularly and temporary changes in nutrient availability are more pronounced in  
67 fertilized grasslands under cutting management (Fariaszewska *et al.*, 2020), which could amplify  
68 adverse effects on plant performance due to a lack of built-up reserves (Reynaert *et al.*, 2022) and  
69 reduced drought resistance (Van Sundert *et al.*, 2021).

70 Changing climate regimes guide the selection of forage grass varieties which guarantee future  
71 sustainable ecosystem provisioning in oceanic Europe (Dumont *et al.*, 2015; Lüscher *et al.*, 2022).  
72 Common forage species respond differently to altered water availability (Volaire *et al.*, 2009; Bahrani  
73 *et al.*, 2010; Poirier *et al.*, 2012; Dumont *et al.*, 2015; Fariaszewska *et al.*, 2020; Lüscher *et al.*, 2022).  
74 Whereas *Lolium perenne* variants are highly productive and of high quality under ample water supply,  
75 *Festuca arundinacea* variants are anisohydric and show reduced plant quality but comparably higher  
76 productivity under drought and/or during recovery because of a more developed rooting system,  
77 greater nutrient-use efficiency and the ability to lower their water potential more, though with  
78 considerable variation between cultivars (Durand *et al.*, 1997; Van Eekeren *et al.*, 2010; Cougnon *et al.*  
79 *et al.*, 2017; Fariaszewska *et al.*, 2017; Becker *et al.*, 2020; Fariaszewska *et al.*, 2020). The use of polyploid  
80 species (Bothe *et al.*, 2018) or hybrids such as the increasingly common *Festulolium* variants aims at  
81 providing more stable yields by combining the positive traits of both species (Durand *et al.*, 1997;  
82 Fariaszewska *et al.*, 2017) and some also show improved flooding resistance (Macleod *et al.*, 2013). In  
83 addition, given the adverse environmental impacts of intensive fertilizer use (Stoate *et al.*, 2009) and  
84 the increased risk of N leaching after drought (Klaus *et al.*, 2020), species with high N yield are  
85 increasingly selected since they generally decrease N leaching from the system (Moir *et al.*, 2013).  
86 *Festuca arundinacea* shows overall greater N yield compared to *Lolium perenne* (Cougnon *et al.*, 2014),  
87 indicating higher efficiency of converting available nutrients into biomass but the mechanisms behind  
88 this are not fully understood (Moir *et al.*, 2013; Cougnon *et al.*, 2017). Repeated dry-wet cycles affect  
89 plant-available soil N and thus plant quality and nitrogen status (Borken & Matzner, 2009; Grant *et al.*,  
90 2014; Hofer *et al.*, 2017; Meisser *et al.*, 2019), complicating how increasing weather persistence could  
91 affect N yield, plant quality and leaching.

92 Soil texture and land-use history determine forage grass and soil responses to drought and flooding  
93 (Dodd & Lauenroth, 1997; Alaoui *et al.*, 2018; Nguyen *et al.*, 2019; Klaus *et al.*, 2020; Van Sundert *et*  
94 *al.*, 2020; Chang *et al.*, 2021; Patel *et al.*, 2021; Radujković *et al.*, 2021; Van Sundert *et al.*, 2021). In  
95 particular, practices that increase soil organic carbon (SOC) such as permanent grassland management  
96 (Crème *et al.*, 2020; Guillaume *et al.*, 2021) have been associated with increased crop yields (D'Hose  
97 *et al.*, 2014; Schjønning *et al.*, 2018; Buttler *et al.*, 2019; Oldfield *et al.*, 2019; Sun *et al.*, 2020), greater  
98 potential to buffer climate change impacts by improving soil water availability during drought  
99 (Hudson, 1994; Minasny & McBratney, 2018) and reduced nutrient leaching during periods of intense  
100 rainfall (Kanthle *et al.*, 2016; Xu *et al.*, 2016; Ahmed *et al.*, 2019).

101 We investigated whether previous permanent grassland management enhancing SOC could  
102 ameliorate the response of forage grass varieties to experimentally imposed increased weather  
103 persistence with longer dry and wet spells. Differences in the persistence of precipitation regimes  
104 were imposed by subjecting monoculture mesocosms to either repeated 3-day (historically normal)  
105 or 30-day (extreme) wet/dry cycles whilst keeping total precipitation equal. We tested this across five  
106 recently developed forage grass varieties differing in drought resistance and forage quality and across  
107 both sandy and sandy-loam soils originating from either permanent (high SOC) or temporary (low SOC)  
108 grasslands in adjacent parcels. Our main goals were to (1) evaluate how the quality and aboveground  
109 productivity of distinct cultivars is affected by increased weather persistence across different soil  
110 texture classes and historical grassland management scenarios and (2) identify underlying drivers in  
111 terms of soil water and nutrient dynamics, and their relationships with plant quality and productivity.

112 Since previous studies in unfertilized grasslands indicate that increasing summer weather persistence  
113 generally affects plants by limiting water availability (Reynaert *et al.*, 2021; Reynaert *et al.*, 2023b), we  
114 hypothesized (i) that plant growth would be more constrained under longer dry and wet spells,  
115 accentuating differences between soils and species. Furthermore, regarding the five forage grass  
116 varieties we expected (ii) a gradient of more to less drought resistance (and lower to higher quality)

117 from *Festuca arundinacea* FERMINA to *Festulolium* FESTILO to *Lolium x*  
 118 *boucheanum* kunth. MELCOMBI to *Lolium perenne* (tetraploid) MELFORCE to *Lolium perenne* (diploid)  
 119 MELSPRING, respectively. Finally, we expected (iii) soils from permanent grasslands with a high SOC  
 120 content to be able to maintain plant productivity and quality better under extreme precipitation  
 121 regimes with more pronounced effects on sandy soils, due to the positive effects of finer particles and  
 122 elevated SOC on soil water and nutrient availability.

## 123 2. Material and methods

### 124 2.1. Field site and experimental setup

125 An open-air field experiment was set up during spring 2021 at the Drie Eiken Campus in Wilrijk,  
 126 Belgium (51°09'41" N, 04°24'9" E). A total of 200 mesocosms (30 cm diameter, 50 cm depth) were  
 127 equally distributed across five plots and filled with sandy or sandy-loam soil from neighboring parcels  
 128 differing in organic C content generated by differences in grassland management (Table 1). Mesocosm  
 129 pots were filled halfway with top soil (0-25 cm) and halfway with deeper soil (25-50 cm) from parcels  
 130 in Zele, Belgium (51°04'25" N, 04°00'49" E; sandy soil) and Melle, Belgium (50°59'02" N 03°47'01" E;  
 131 sandy-loam soil), after which all soils were equally compacted.

132 **Table 1.** Soil characteristics at the start of the experiment.

Grassland									
Soil texture	management	Depth	OC	pH	N <sub>tot</sub>	K	Mg	Ca	P
Sand	Permanent	0-25 cm	1.83	5.9	0.17	7	27	104	19
		25-50 cm	0.16	5.4	0.02	4	13	69	4
	Temporary	0-25 cm	1.48	5.7	0.13	20	19	108	50
		25-50 cm	0.21	5.8	0.02	14	19.	99	28
Sandy loam	Permanent	0-25 cm	1.39	4.8	0.14	10	18	61	31
		25-50 cm	0.38	4.8	0.04	6	15.	42	9

	0-25 cm	0.87	6.0	0.09	27	16	81	23
Temporary	25-50 cm	0.33	5.5	0.04	10	10	54	10

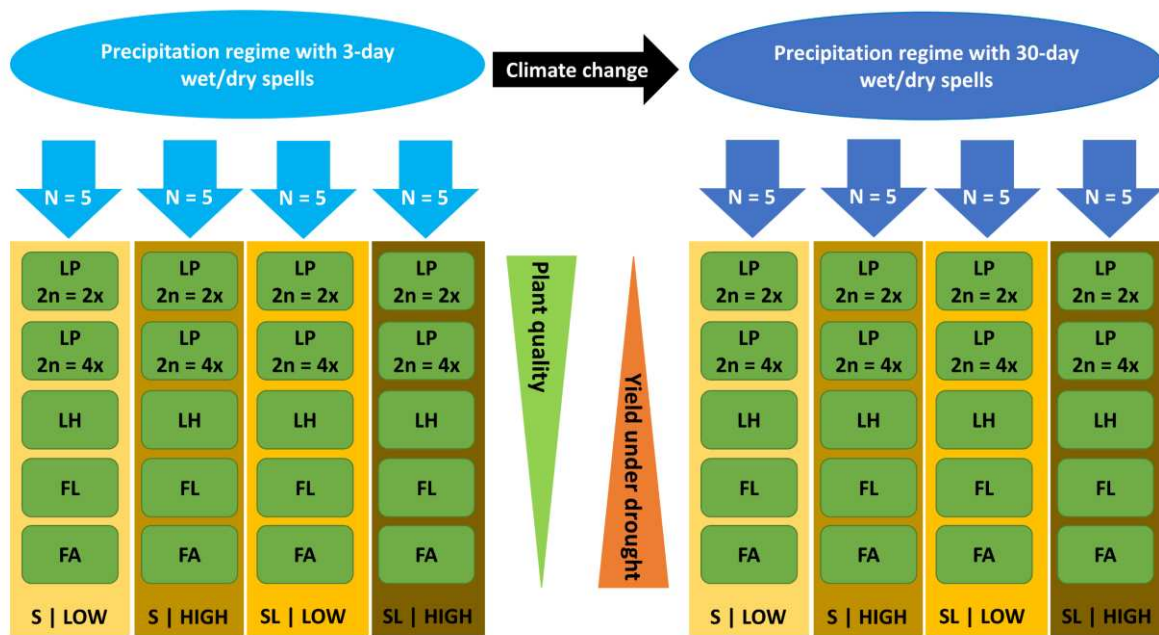
133 Organic carbon (OC) and total nitrogen ( $N_{\text{tot}}$ ) content are expressed in % dry matter (DM). K, Mg, Ca  
 134 and P concentrations are expressed in mg 100 g<sup>-1</sup> DM.

135 Five distinct cultivars of commonly used C3 forage species were selected to create a gradient in trade-  
 136 offs between drought/flooding resistance and plant quality. From generally more to less drought  
 137 resistant (and generally less to more quality; Fariaszewska *et al.* (2020)) these included *Festuca*  
 138 *arundinacea* FERMINA, *Festulolium* FESTILO, *Lolium x boucheanum* kunth. MELCOMBI - referred to as  
 139 *Lolium hybridum* from here-on -, *Lolium perenne* (tetraploid) MELFORCE and *Lolium perenne* (diploid)  
 140 MELSPRING. Seeds were sown in seedling containers under greenhouses late February and regularly  
 141 watered over the course of the following two months to promote germination and seedling  
 142 development. At the end of April, 37 healthy seedlings per species were transplanted into each of 40  
 143 mesocosms in a predetermined grid, ensuring equal planting densities (interspace of 4.5 cm). Two  
 144 replicate mesocosms of every unique combination of soil type and species were assigned to each of  
 145 the five plots, and their locations were randomized to account for potential edge effects. Seedlings  
 146 that did not survive transplanting were replaced until the end of May. Manual weeding was done  
 147 throughout the experiment. All mesocosms received equal amounts of water until mid-July 2021, to  
 148 maintain the soils close to field capacity before the start of the experiment (Elst *et al.*, 2017; Reynaert  
 149 *et al.*, 2021).

150 From 14/07/21 until 10/11/21 (DOY 165 – DOY 314; 120 days total in 2021), half of the mesocosms  
 151 were subjected to a historically close to normal precipitation regime (PR) (sequences of 3 days wet /  
 152 dry) and the other half to an extreme PR with highly persistent weather (sequences of 30 days  
 153 wet/dry) (Reynaert *et al.*, 2021). These regimes were simulated accurately utilizing automated dripper  
 154 irrigation below plot-level screens that close automatically upon rain-detection to block-out ambient  
 155 precipitation (Reynaert *et al.*, 2021). Water volumes applied were equal in total between treatments  
 156 (60 days wet and 60 days dry) and per irrigation event (6.87 L m<sup>-2</sup> which is 1.5 times the daily Belgian



157 average to compensate for artificially increased evapotranspiration in the mesocosms; Reynaert *et al.*  
 158 (2021)). This resulted in a full factorial design containing a total of five replicates (1 per plot) per unique  
 159 factor combination of the two soil textures, two SOC levels, two PR treatments and five species (Fig.  
 160 1).



161

162 **Figure 1.** Schematic of the experimental setup including all treatments, with a total of 200 mesocosms.  
 163 *Lolium perenne*, *Lolium hybridum*, *Festulolium* and *Festuca arundinacea* are abbreviated by LP, LH, FL  
 164 and FA, respectively. Diploid, tetraploid, sandy soil, sandy-loam soil and soil organic carbon level are  
 165 abbreviated by 2n = 2x, 2n = 4x, S, SL, LOW and HIGH, respectively. The plant quality and yield under  
 166 drought trends show hypothesized differences based on previous studies (Fariaszewska *et al.*, 2017;  
 167 Fariaszewska *et al.*, 2020).

168 Since our experiment focused on simulation of increasing summer weather persistence, mesocosms  
 169 were subjected to ambient precipitation over winter. From 02/03/22 – 27/10/22 (DOY 61 - DOY 300;  
 170 240 days total in 2022), all mesocosms were again subjected to their respective PR treatment.  
 171 Fertilization was equal for all mesocosms and occurred in line with common agricultural practices  
 172 (Table S1), resulting in the application of 153 g N ha<sup>-1</sup>, 38 kg P ha<sup>-1</sup> and 165 kg K ha<sup>-1</sup> in the installation

173 year (2021) and 302 g N ha<sup>-1</sup>, 38 kg P ha<sup>-1</sup> and 325 kg K ha<sup>-1</sup> in the first production year (2022). (Table  
174 S2). A dissolved fertilizer was used and to mimic realistic grassland management, fertilization was  
175 fractionated and applied after every cut (except the last per year) when the soil was wet (Table S2).

## 176 **2.2. Measurements and calculations**

177 Across the full experimental period, soil water content averaged over 30 cm depth and soil surface  
178 temperatures were logged automatically every half-hour utilizing CS650-DS Reflectometers  
179 (Campbell® Scientific INC.) in a total of 40 mesocosms (1 per unique full-factorial treatment  
180 combination). Photosynthetically active radiation (PAR) was measured every 10 min in the vicinity of  
181 the plots, utilizing a SKP215 Quantum Sensor (Campbell® Scientific INC., Logan, Utah, USA). Hourly  
182 values of air temperature (T) and relative humidity (RH) were collected at 1.5 m height from a nearby  
183 weather station in Woensdrecht, The Netherlands (50°56'34" N, 5°34'44" E). Utilizing these values,  
184 we calculated average daily vapor pressure deficit (VPD) during sunshine hours. We also collected  
185 green cover estimates from all mesocosms by visually estimating the fraction of plant cover that was  
186 green to the closest 5% on a weekly basis starting from the first harvest every year and with a break  
187 in winter.

188 Aboveground biomass was harvested four times in 2021 and five times in 2022 in all mesocosms by  
189 clipping all standing biomass above 4.5 cm (Table S2). Biomass was then air-dried, stored in paper bags  
190 and oven dried for > 72 h, followed by immediately weighing to the closest 0.01 g. For two cuts in 2022  
191 (June and August – to capture variation across summer), we also determined crude protein content  
192 (CPC), *in vitro* organic matter digestibility (IVOMD) and nitrogen (N) content of the harvested plant  
193 material. Samples were first ground (mesh size = 1 mm) and then scanned for near infrared reflectance  
194 spectroscopy (NIRS) spectra utilizing a Foss NIRSystems 5000 (FOSS NIRSystems, Silver Springs, MD,  
195 USA) and the ISIScan 2.85.1 software (Infrasoft international, Port Matilda, PA, USA). The NIRS  
196 equations for CPC and IVOMD were based on chemical calibration of 40 of the most spectrally distant

197 samples, following the Kjeldahl method and Tilley and Terry (1963), respectively (Cougnon *et al.*,  
198 2014). Per sample, N content was then estimated as % CPC / 6.25 (Jones, 1931).

199 Based on plant N content and biomass estimations we then calculated the nitrogen nutrition index  
200 (NNI) for June and August of 2022 to assess the nitrogen limitation following Lemaire *et al.* (1997) as  
201  $NNI = N_{\text{measured}} / N_{\text{critical}}$ . Here,  $N_{\text{measured}}$  is the N concentration (%) in the aboveground biomass and  $N_{\text{critical}}$   
202 the critical N concentration needed in the grass for unlimited growth in case of unlimited other  
203 resources (i.e., light, water, etc.) depending on dry matter yield (DMY). For  $C_3$  grasses  $N_{\text{critical}} = 4.8 \times$   
204  $DMY$  (in  $t\ ha^{-1}$ )<sup>-0.32</sup>. When  $DMY < 1$ , we considered  $N_{\text{critical}}$  to be 4.8 %, assuming absence of competition  
205 among individuals with low aboveground productivity (Poirier *et al.*, 2012).

206 To investigate effects of soil texture, organic carbon content, PR treatment and species on nutrient  
207 supply rates in the rooting zone upon rewetting after drought in the 30-day PR, we installed eight  
208 Plant root simulator (PRS<sup>®</sup>) probes (Western Ag Innovations, Saskatoon, Canada) at 3-9 cm depth in  
209 each of 40 mesocosms (2 cation and 2 anion pairs per unique treatment combination) in 2022 between  
210 30/06 and 5/7 (DOY 181-186; following harvesting). After five days, probes were collected, cleaned  
211 with distilled water and returned to the manufacturer for lab analysis of total N,  $NO_3^-$ , P, K, Ca, Mg,  
212 Fe, B, Al, Zn and Mn contents.

213 Finally, after the last biomass harvest, we collected a total of 20 soil cores in the 5-10 cm soil layer to  
214 determine top soil bulk density (BD) and the water retention curves of the four respective soils. To  
215 capture variation induced by the various treatments, we collected 5 cores per unique soil type (texture  
216 x SOC), 4 cores per species and 10 cores per PR. We then calculated separate mean values of FC and  
217 PWP per unique soil texture x grassland management combination (Table 2), which were further  
218 utilized to determine the plant available water per soil type and to calculate a soil drought index.

219

220

221 **Table 2.** Soil hydrological properties per soil texture class at the end of the experiment.

Soil texture	Grassland management	Bulk density (g / cm <sup>3</sup> )	Average volumetric moisture content (cm <sup>3</sup> / cm <sup>3</sup> )				AWC (FC - PWP)
			pF 4.2 (PWP)	pF 2.7	pF 2.0 (FC)	pF 0	
Sand	Permanent	1.098 ±	0.042 ±	0.160 ±	0.264 ±	0.526 ±	0.222 ±
		0.032 <sup>a</sup>	0.001 <sup>a</sup>	0.011 <sup>a,b</sup>	0.012 <sup>a</sup>	0.009 <sup>a</sup>	0.011 <sup>a</sup>
	Temporary	1.152 ±	0.039 ±	0.150 ±	0.258 ±	0.488 ±	0.219 ±
		0.008 <sup>a</sup>	0.001 <sup>a</sup>	0.016 <sup>a,b</sup>	0.016 <sup>a</sup>	0.014 <sup>a,b</sup>	0.015 <sup>a,b</sup>
Sandy loam	Permanent	1.206 ±	0.038 ±	0.170 ±	0.302 ±	0.470 ±	0.264 ±
		0.029 <sup>b</sup>	0.001 <sup>a</sup>	0.019 <sup>b</sup>	0.010 <sup>b</sup>	0.003 <sup>b</sup>	0.008 <sup>b</sup>
	Temporary	1.196 ±	0.035 ±	0.141 ±	0.266 ±	0.478 ±	0.231 ±
		0.020 <sup>b</sup>	0.001 <sup>a</sup>	0.009 <sup>a</sup>	0.009 <sup>a</sup>	0.007 <sup>b</sup>	0.009 <sup>a</sup>

222 Average soil hydrological properties ( $\pm 1$  SE) per unique soil texture class x grassland management  
 223 combination (n=5) taken from the 5-10 cm soil layer at the end of the experiment. Permanent wilting  
 224 point, field capacity and plant available water (difference between PWP and FC) are abbreviated as  
 225 PWP, FC and AWC, respectively. <sup>a,b</sup>Different letters indicate significant differences ( $P < 0.05$ ) of Tukey  
 226 tests comparing treatments.

227 The drought index utilized was drought intensity ( $I_s$ )(Vicca *et al.*, 2012), which integrates both the  
 228 duration and extent to which daily average SWC declines below a critical soil water threshold (at  
 229 relative extractable water  $< 0.4$ ). First, total extractable water (TEW) was calculated over the 30 cm  
 230 sensor depth:

231  $TEW_{max} (\%) = SWC_{FC} (\%) - SWC_{PWP} (\%)$

232  $TEW_{day} (\%) = SWC_{day} (\%) - SWC_{PWP} (\%)$

233 We then determined relative extractable water (REW) as  $REW_{day} = TEW_{day} / TEW_{max}$ , finally resulting in

234  $I_s$  at daily and over a period basis:

235  $I_{s, day} = 0.4 - REW_{day}$  if  $REW_{day} < 0.4$  and  $I_{s, day} = 0$  if  $REW_{day} > 0.4$

236  $I_{s, \text{period}} = \sum I_{s, \text{first day of period}} + \dots + I_{s, \text{last day of period}}$

237 We calculated this index for each mesocosm with SWC sensor (n = 40) both per harvest and per year  
238 to assess how precipitation regime, soil texture, grassland management and species identity  
239 determined ecological soil drought extremity over the course of the two year experiment.

### 240 **2.3. Statistical analysis**

241 All statistics were performed in R (version 4.0.4). We assumed significance of effects for *P*-values <  
242 0.05. Where possible, we always started from a full model including all hypothesized interactions and  
243 selected final models by removing insignificant higher-order interactions based on F-test model  
244 comparison. Model assumptions were verified utilizing residual plots and tests. Post-hoc Tukey tests  
245 were applied to further explore 2-by-2 differences utilizing the package emmeans (Lenth *et al.*, 2018).  
246 Where necessary because of residual heteroscedasticity, we log transformed variables and/or allowed  
247 variance to vary with factor levels, utilizing the *varIdent()* or *dispformula()* functions in nlme (Pinheiro  
248 *et al.*, 2017) and glmmTMB (Magnusson *et al.*, 2017), respectively.

249 To assess general drivers and indicators of plant stress, we first performed ANOVA on two generalized  
250 linear mixed models (GLMMs) (family = gaussian, link = identity) testing for the effects of soil texture,  
251 year, grassland management, species identity, PR and their interactions on average soil moisture and  
252 drought intensity. These models included mesocosm nested in plot as random intercept. Next, we  
253 conducted ANOVA on a GLMM (family = beta, link = logit) including green cover as response variable,  
254 with soil texture, DOY, grassland management, species identity, PR and their interactions as fixed  
255 effects and mesocosm nested in plot as random intercept. Cover proportions were transformed in  
256 order to avoid extreme values of 0 and 1 which is necessary for beta regression (Damgaard & Irvine,  
257 2019; Clavel *et al.*, 2021). To further assess plant responses, we then used ANOVA to test for the effects  
258 of soil texture, grassland management, species identity, PR and their interactions on aboveground  
259 productivity (both cumulative per year, and per harvest) and plant quality (IVOMD, CPC and N-

260 content; per harvest) in five GLMMs (family = gaussian, link = identity), again including mesocosm  
261 nested in plot as random intercept.

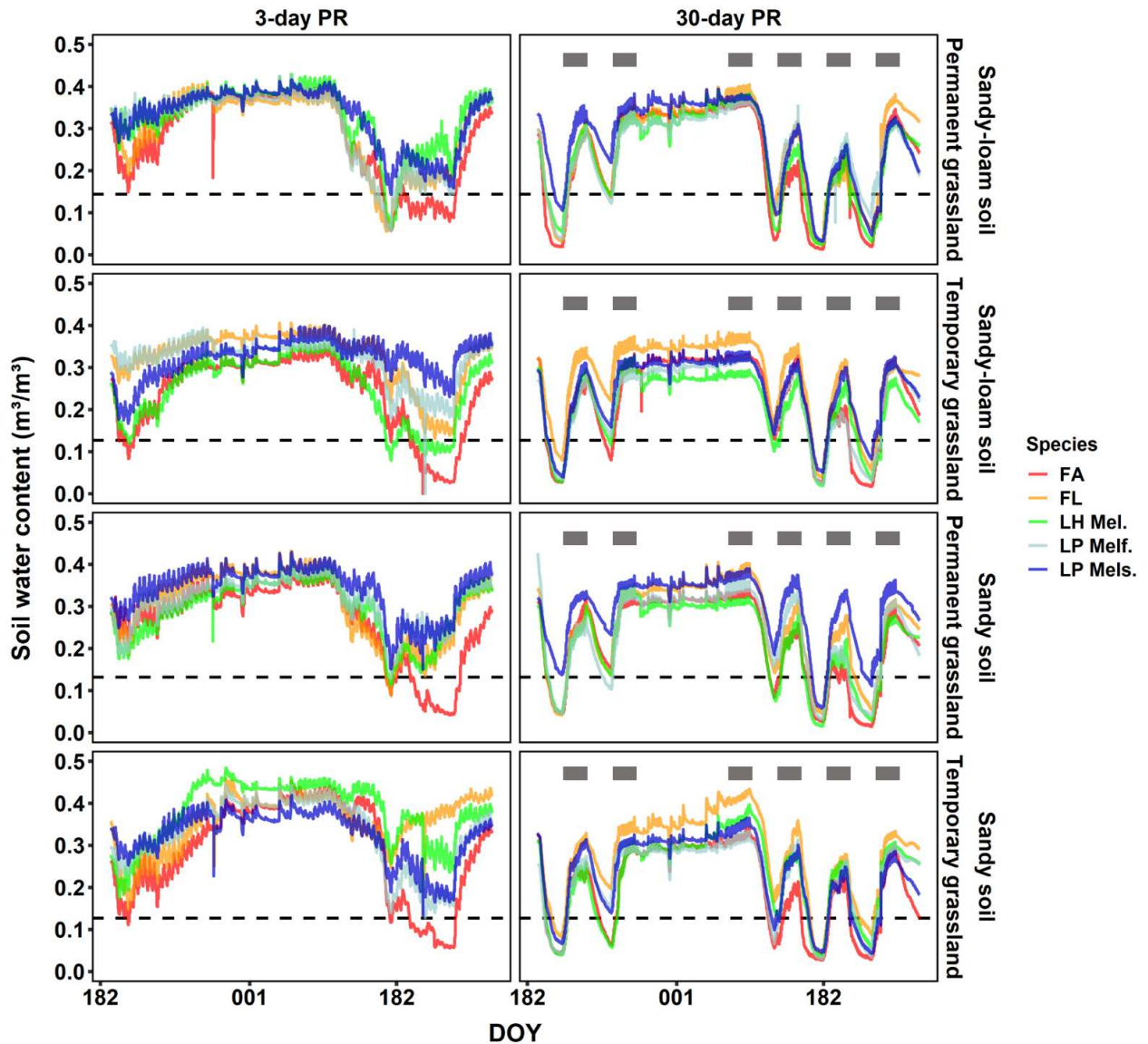
262 Underlying mechanisms for the observed patterns related to soil water availability were then explored  
263 by performing ANOVA on a GLMM (family = gaussian, link = identity) including bulk density as response  
264 variable, with soil texture, PR and grassland management and their interactions as fixed effects and  
265 species identity as random intercept (due to limited observations, species samples were not taken  
266 representatively across all treatment levels). Next, we constructed a GLMM (family = gaussian, link =  
267 identity) with moisture content as response variable and pF value, soil texture, grassland  
268 management, PR and their interactions as fixed effects and mesocosm nested in species identity as  
269 random intercept (due to limited observations, species samples were not taken representatively  
270 across all treatment levels). To unravel the role of soil nutrient dynamics, we also performed ANOVA  
271 on 11 GLMMs (family = gaussian, link = identity) for each of the measured nutrient supply rates during  
272 rewetting in July of 2022 (total N, NO<sub>3</sub><sup>-</sup>, P, K, Ca, Mg, Fe, B, Al, Zn and Mn), always including PR, species,  
273 soil texture and grassland management as well as their two-way interactions as fixed effects, and  
274 mesocosm as random intercept. To asses effects on the nutrient status of the plants, we performed  
275 ANOVA to test for the effects of soil texture, grassland management, species identity, PR and their  
276 interactions on NNI (family = gaussian, link = identity), again with mesocosm nested in plot as random  
277 intercept. Finally, we constructed a set of five models (family = gaussian, link = identity) testing the  
278 interactive effects of water (drought intensity) and nutrient (NNI) limitation on plant biomass and  
279 quality across species and harvests, and integrated these findings in nonmetric multidimensional  
280 scaling (NMDS) ordination on observations from mesocosms with soil water sensors (N = 40) in June  
281 (early 30-day PR drought) and August (post 30-day PR rewetting) of 2022.

## 282       **3. Results**

### 283       **3.1. Patterns in weather, soil water availability and plant green cover**

284       Ambient weather conditions contrasted between the two years of the experiment, with a relatively  
285       cold and humid summer in 2021 followed by a hot and dry summer in 2022, as indicated by differences  
286       in air temperature, vapor pressure deficit and soil surface temperatures (Fig. S1 & S2).

287       Soil moisture trajectories generally reflected these weather differences, with more extreme declines  
288       in available soil water in 30-day PR and during 2022 (Fig. 2). As expected, the 30-day PR generally  
289       reduced the available water in the soil on average (-8 % compared to the 3-day PR), resulting in higher  
290       drought intensity across both years ( $P < 0.001$ ; Fig. 2; Fig. S4; Table S3). Soils from permanent  
291       grasslands contained on average more water compared to temporary grasslands, but only in 2021  
292       (+2.3 %;  $P = 0.004$ ; Table S3). In contrast, differences in soil water availability induced by PR and  
293       between species were more pronounced in 2022 ( $P < 0.001$ ; Table S3) and within the 30-day PR among  
294       species ( $P = 0.007$ ; Table S3), with *Festuca* experiencing more extreme soil moisture declines  
295       compared to all other species, *Lolium hybridum* more compared to *Festulolium* and *Lolium perenne*  
296       MELSPRING, and *Lolium perenne* MELFORCE more compared to *Lolium perenne* MELSPRING (Fig. S3;  
297       Table S3).



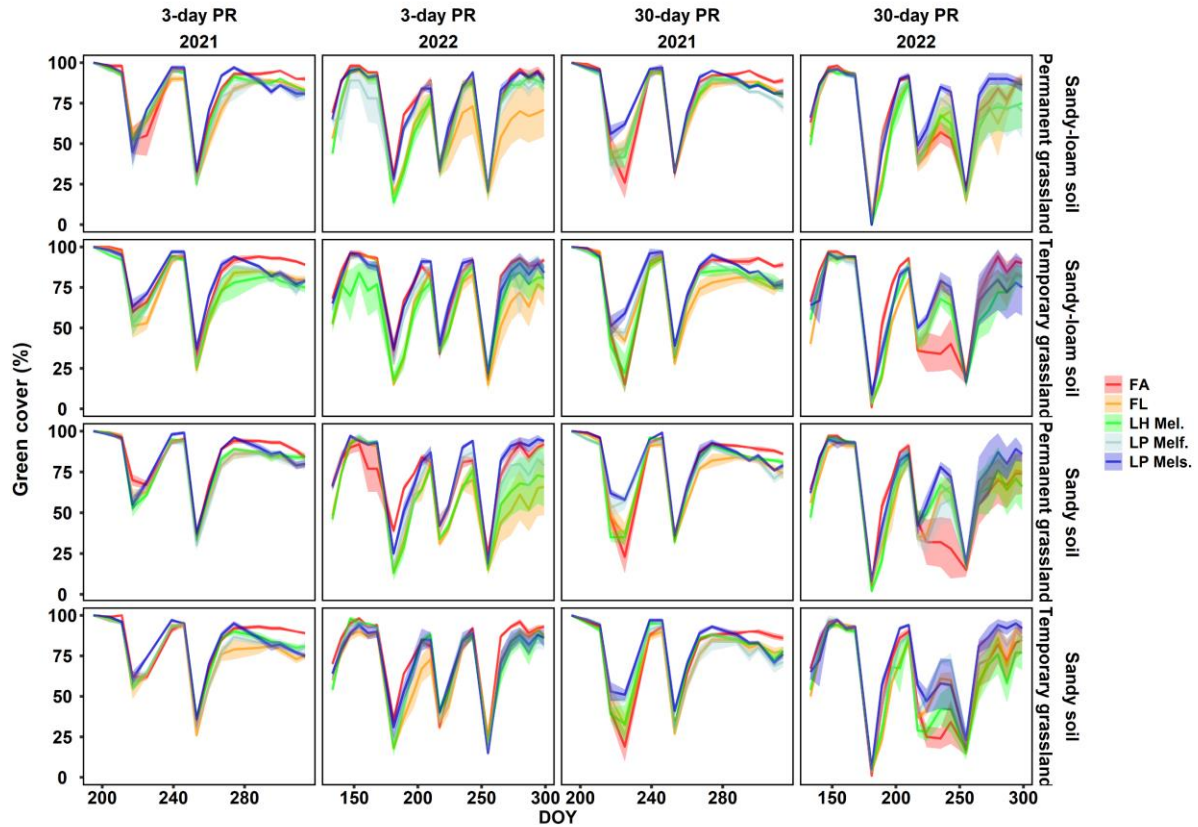
298

299 **Figure 2.** Effects of precipitation regime (PR) on temporal trajectories of soil water content per species,  
 300 texture class (sandy vs sandy-loam), grassland management level (permanent vs temporary  
 301 grassland). Grey boxes highlight the wet periods of the 30-day regimes and the dashed line indicates  
 302 where 40 % of water at FC is still available in the soil. Species names including *Festuca arundinacea*,  
 303 *Festulolium*, *Lolium hybridum*, *Lolium perenne* MELFORCE (tetraploid) and *Lolium perenne* MELSPRING  
 304 (diploid) are abbreviated as FA, FL, LH Mel., LP Melf. and LP Mels., respectively. Day of the year is  
 305 abbreviated as DOY.



306 Plant green cover trajectories generally confirmed these trends. Plants subjected to the 30-day PR  
307 were on average 20 % less green compared to those growing under the 3-day PR. Differences between  
308 PR were most pronounced under hotter and drier weather conditions and towards the end of the  
309 growing season (Fig. 2;  $P < 0.001$ ; Table S4). Plants growing on soils from permanent grasslands were  
310 on average greener across 2021, but this effect disappeared and even reversed by September and  
311 October of 2022 (Fig. 2;  $P < 0.001$ ; Table S4). This grassland management effect differed between soil  
312 textures, with mostly positive effects of permanent grassland history in sandy-loam soils, but neutral  
313 or negative effects in sandy soils (Fig. 2;  $P = 0.022$ ; Table S4). On average, plants were also greener (+  
314 7 %) when growing on sandy-loam soils and these effects were most pronounced in the summer of  
315 2022 (Fig. 2;  $P < 0.001$ ; Table S4).

316 Differences among species were greatest in the recovery phase following harvesting, and during hot  
317 and dry atmospheric conditions in the 30-day PR vs during cold and humid weather in the 3-day PR  
318 (Fig. 2;  $P < 0.001$ ; Table S4). In line with our expectations, *Festuca* was on average greener compared  
319 to *Festulolium*, *Lolium hybridum*, and *Lolium perenne* MELFORCE (Fig. 3;  $P = 0.001$ ; Table S4; Fig. S4).  
320 However, rather unexpected, both *Lolium perenne* cultivars were on average greener than  
321 *Festulolium*, and, *Lolium perenne* MELSPRING was on average greener compared to *Lolium hybridum*  
322 and *Lolium perenne* MELFORCE (Fig. 3;  $P = 0.001$ ; Table S4; Fig. S4). Rapid changes in plant greenness  
323 suggested that *Festuca* had the most plastic response (i.e. recovery ability) to the treatments under  
324 the imposed cutting management, bouncing between the highest highs (e.g., July 2022) and the lowest  
325 lows (e.g., August 2022) out of any species (Fig. 3;  $P < 0.001$ ; Table S4).



326

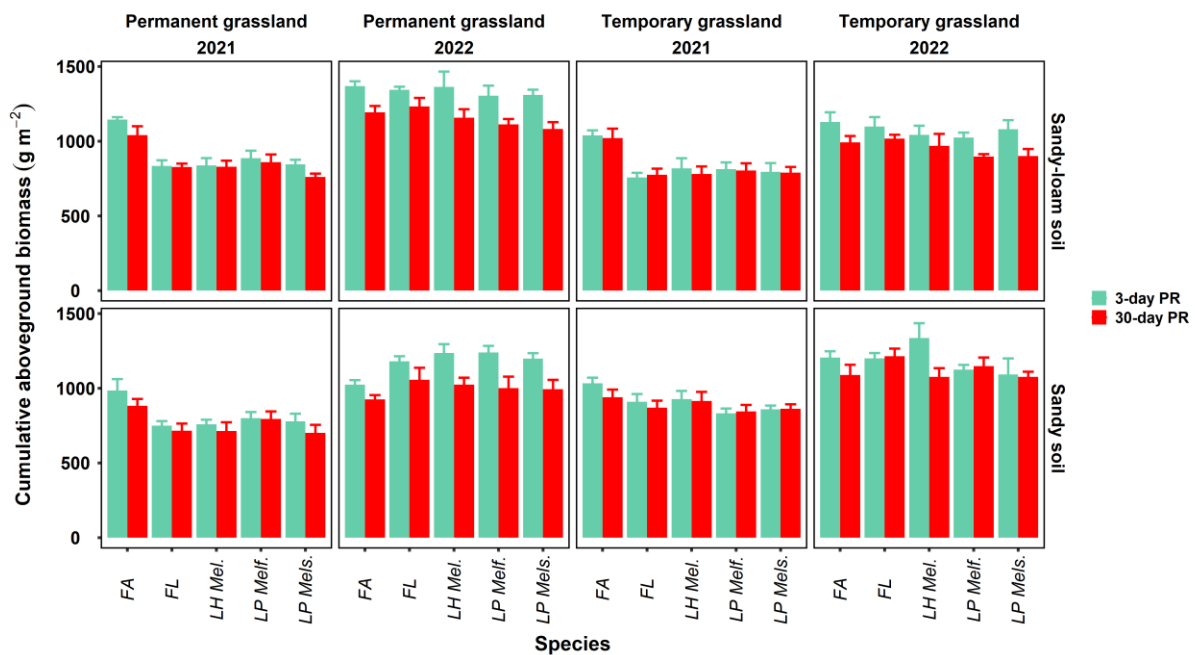
327 **Figure 3.** Effects of precipitation regime (PR) on temporal trajectories of green cover per species,  
 328 texture class (sandy vs sandy-loam) and grassland management level (permanent vs temporary  
 329 grassland). The shaded area indicates +/- 1 SE on the mean. Species names including *Festuca*  
 330 *arundinacea*, *Festulolium*, *Lolium hybridum*, *Lolium perenne* MELFORCE (tetraploid) and *Lolium*  
 331 *perenne* MELSPRING (diploid) are abbreviated as FA, FL, LH Mel., LP Melf. and LP Mels., respectively.  
 332 Day of the year is abbreviated as DOY. The sharp current dips in the curves indicate biomass  
 333 harvesting.

334 **3.2. Patterns in dry matter yield and forage quality**

335 Subjection to the 30-day PR reduced cumulative aboveground productivity by 4 % in 2021 and by 14%  
 336 in 2022 on average compared to 3-day PR ( $P < 0.001$ ; Table S5; Fig. 4) without differentially affecting  
 337 species. Only within the 3-day PR ,growing on soils from permanent grasslands generally improved  
 338 cumulative productivity (+ 5 %).( $P = 0.014$ ; Table S5; Fig. 4). The influence of grassland management

339 on cumulative productivity contrasted between texture classes and years, with permanent grassland  
 340 enhancing productivity compared to temporary grassland on sandy-loam soils (by 23 %, only in 2022),  
 341 but the opposite trend on sandy soils across both years (- 12 % in 2021 and - 6 % in 2022) ( $P = 0.001$ ;  
 342 Table S5; Fig. 4). *Festuca* outperformed all other species on average in 2021 (+ 24 %) but this effect  
 343 disappeared in the following year. In 2022, *Festulolium* performed on average slightly better  
 344 compared to *Lolium perenne* MELFORCE (+ 6 %) and *Lolium perenne* MELSPRING (+ 8 %) ( $P < 0.001$ ;  
 345 Table S5; Fig. 4). These differences in cumulative species performance were generally more  
 346 pronounced in sandy-loam soils compared to sandy soils (species  $\times$  texture interaction;  $P = 0.002$ ;  
 347 Table S5; Fig. 4). Across both years, *Lolium perenne* MELSPRING overall had the lowest cumulative  
 348 productivity ( $P < 0.001$ ; Table S5; Fig. 4).

349



350

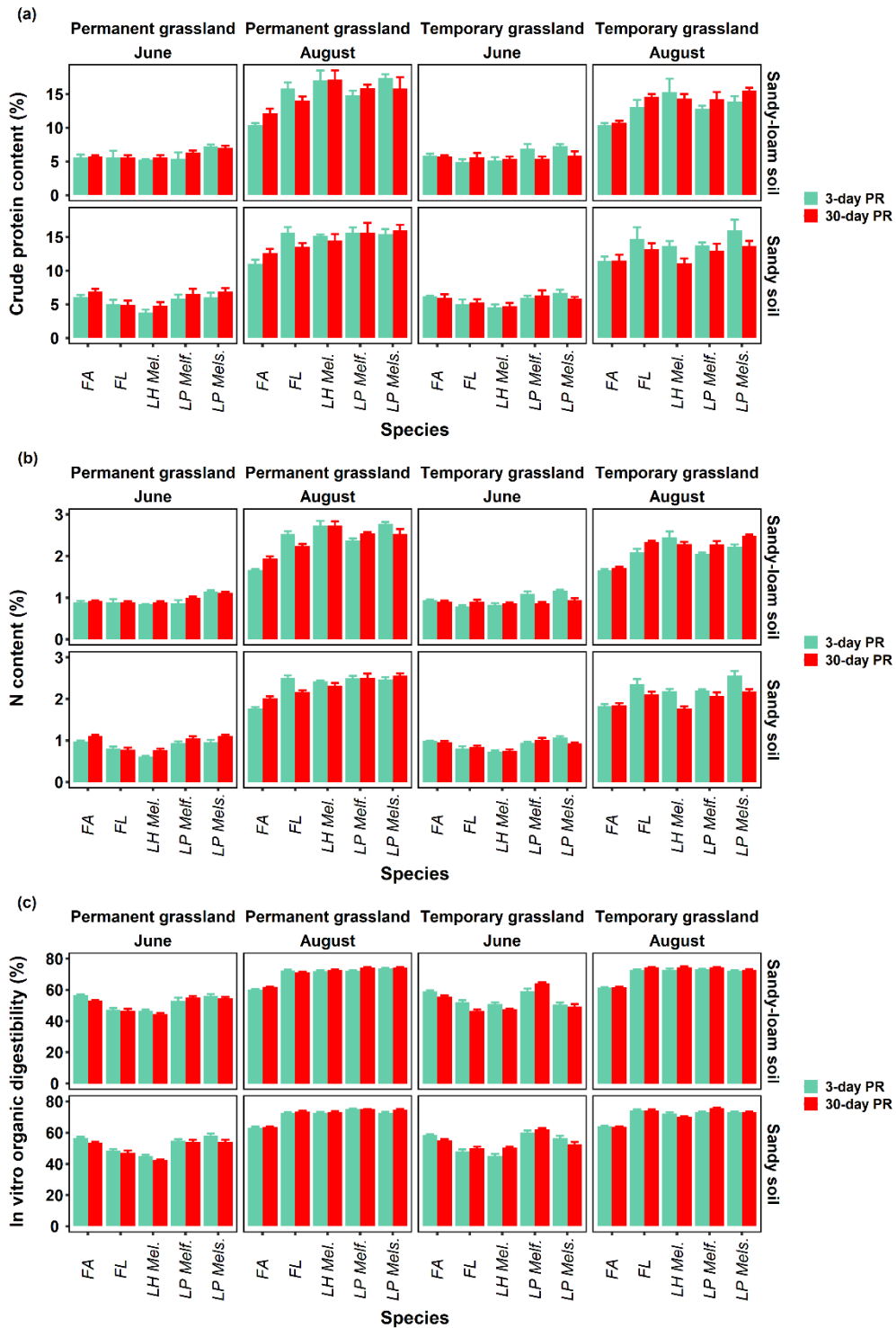
351 **Figure 4.** Effects of precipitation regime (PR) on cumulative aboveground biomass per species and  
 352 year for each soil texture and grassland management level. *Festuca arundinacea*, *Festulolium*, *Lolium*  
 353 *hybridum* MELCOMBI, *Lolium perenne* MELFORCE and *Lolium perenne* MELSPRING are abbreviated by  
 354 FA, FL, LH Mel., LP Melf. and LP Mels., respectively. Error bars indicate +/- 1 SE on the mean.

355 When considering biomass differences on a per harvest basis, similar patterns emerged (Fig. S5). Most  
356 notably, negative effects of PR and modulating influences of grassland management, species identity  
357 and soil texture were strongest during hot and dry periods (i.e. mid-summer: DOY 150 - 250) and in  
358 2022 (Fig. S5;  $P < 0.001$ ; Table S5). Moreover, across 2022, grassland management did not significantly  
359 impact productivity on a per harvest basis within the sandy soils (contrasting with the cumulative  
360 response), whereas a permanent grasslands history consistently positively influenced yields for sandy-  
361 loam soils (Fig. S5;  $P < 0.001$ ; Table S5).

362 The forage quality parameters in summer 2022 were influenced by PR, grassland management, soil  
363 texture, harvesting date and species identity (Fig. 5; Table S6). Under 30-day PR in August, CPC was  
364 increased for soils from permanent grasslands (+ 1.5 %;  $P = 0.032$ ; Table S6; Fig. 5), or with a sandy-  
365 loam texture (+ 1 %;  $P = 0.022$ ; Table S6; Fig. 5). Generally, crude protein content (CPC) was on average  
366 8 % higher in August compared to June ( $P < 0.001$ ; Table S6; Fig. 5). Differences between species also  
367 contrasted between harvests, with *Festuca* having a higher CPC compared to *Festulolium* and *Lolium*  
368 *hybridum* in June, but a lower CPC compared to all others in August ( $P < 0.001$ ; Table S6; Fig. 5).  
369 Furthermore, *Festulolium* and *Lolium hybridum* had lower CPC compared to *Lolium perenne* cultivars,  
370 but only in June ( $P < 0.001$ ; Table S6; Fig. 5) and these species differences were slightly modified by  
371 soil texture ( $P = 0.002$ ; Table S6; Fig. 5)..

372 Given the utilized calculation, trends in N-content were virtually the same as in CPC (Fig. 5; Table S6).  
373 However, in contrast, IVOMD was on average 0.6 % higher in 30-day PR vs 3-day PR in August ( $P =$   
374  $0.012$ ; Table S6; Fig. 5). Following the CPC trends, *Festuca* generally had higher digestibility in June but  
375 lower in August compared to others and both *Lolium perenne* cultivars were most digestible (up to 12  
376 % higher in August compared to *Festuca*;  $P = 0.003$ ; Table S6; Fig. 5).

377



378

379 **Figure 5.** Effects of precipitation regime (PR) on plant quality (a: Crude protein content, b: N content,  
 380 c: In vitro organic matter digestibility) in June and August of 2022 per species, texture class (sandy vs  
 381 sandy-loam) and grassland management (permanent vs temporary grassland) level. The error bars  
 382 indicate +/- 1 SE on the mean. Species names including *Festuca arundinacea*, *Festulolium*, *Lolium*

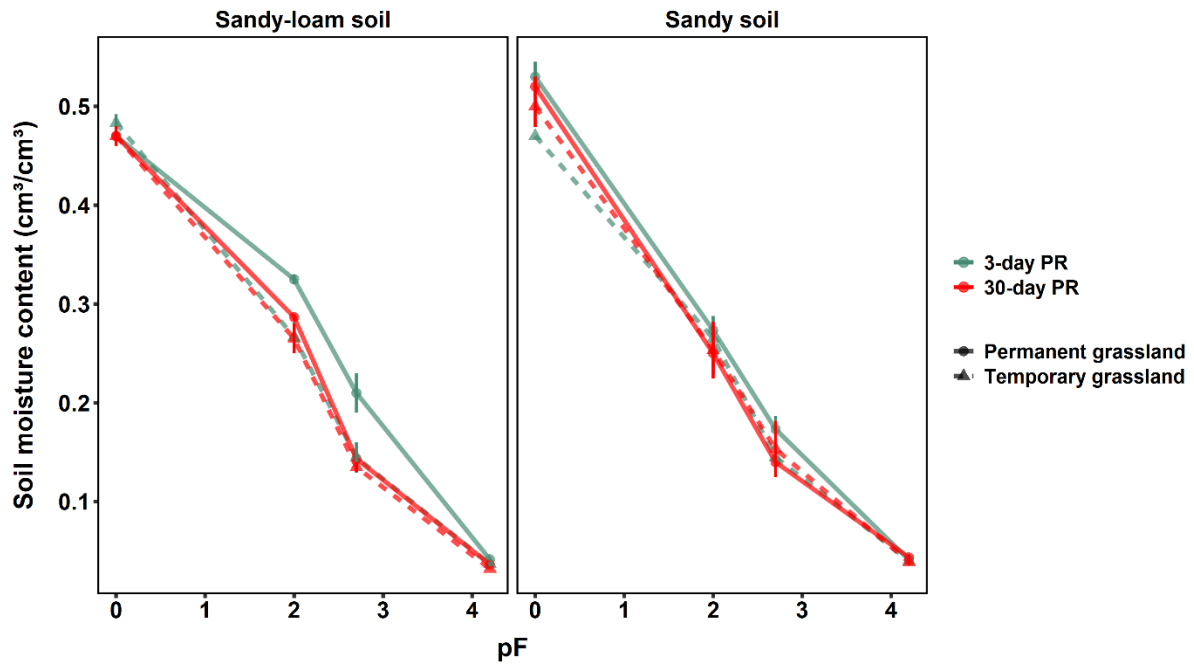
383 *hybridum*, *Lolium perenne* MELFORCE (tetraploid) and *Lolium perenne* MELSPRING (diploid) are  
384 abbreviated as FA, FL, LH Mel., LP Melf. And LP Mels., respectively. Day of the year is abbreviated as  
385 DOY.

### 386 **3.3. Drivers of patterns in forage grass performance and quality**

#### 387 **3.3.1. Effects of precipitation regime and grassland management on nutrient** 388 **availability and soil hydrological properties**

389 To disentangle nutrient dynamics and soil properties underlying these trends, we characterized  
390 nutrient availability during a 30-day PR rewetting event in July 2022 (see Fig. S6 for detailed soil water  
391 trajectories). Increased PR persistence generally affected nutrient supply on a species-specific basis,  
392 only increasing N and K supply for *Festuca* (+ 102% N;  $P < 0.001$ ; Table S8; Fig. S7a) and *Lolium perenne*  
393 MELFORCE (+ 92% K;  $P = 0.023$ ; Table S7; Fig. S7d), respectively, and doing the opposite regarding K  
394 supply for *Festulolium* (- 76% K;  $P = 0.023$ ; Table S7; Fig. S7d). However, buffering effects of permanent  
395 grassland management on nutrient supply generally mirrored the productivity trends, with permanent  
396 grassland management overall not affecting or improving N,  $\text{NO}_3^-$ , P and Mn supply in sandy-loam  
397 soils, but neutral or opposite effects in sandy soils (Table S7; Fig. S7).

398 In terms of soil characteristics determining available water, bulk density remained unaffected by PR  
399 and grassland management but was generally higher in sandy loam soils compared to sandy soils ( $P =$   
400  $0.02$ ; Table S8; Fig. S9).. However, 30-day PR generally negatively influenced soil water retention  
401 capacity ( $P < 0.001$ ; Table S8; Fig. 6), though the extent of this decline was determined by grassland  
402 management and only significant in permanent grasslands ( $P = 0.02$ ; Table S8; Fig. 6). More water was  
403 retained in soils from permanent grasslands on average across both texture classes, although only  
404 when subjected to the 3-day PR ( $P = 0.02$ ; Table S8; Fig. 6). Furthermore, soil water retention showed  
405 an interaction between grassland management and texture class ( $P < 0.001$ ; Table S7; Fig. 6), indicating  
406 that soils originating from permanent grasslands had on average increased water retention at  
407 intermediate pF but only in the sandy-loam texture class (Fig. 5).



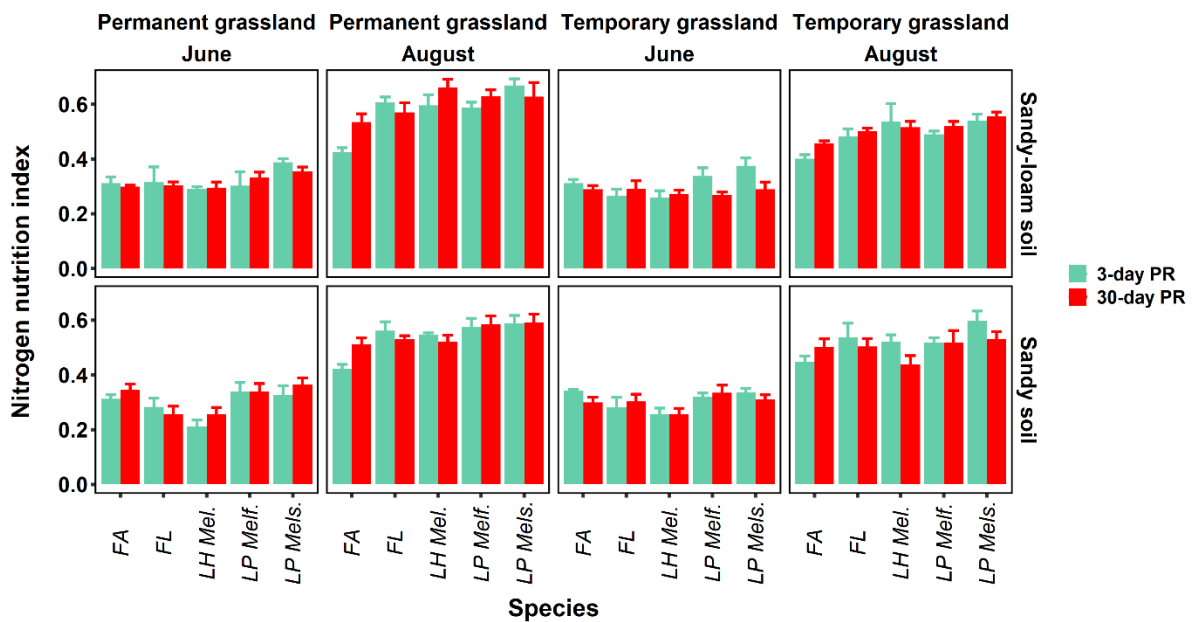
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409 **Figure 6.** Effects of precipitation regime (PR) on water retention curves for the top soil layer (5- 10 cm)  
 410 at the end of the experiment per soil type differing in grassland management and soil texture. The  
 411 error bars indicate +/- 1 SE on the mean.

412 **3.3.2. Effects of precipitation regime and grassland management on forage nitrogen**  
 413 **nutrition**

414 Since increased PR persistence may not only affect nutrient and moisture availability but also uptake,  
 415 we further explored patterns in nitrogen limitation to plant growth (i.e., nitrogen nutrition index (NNI);  
 416 section 3.2). These differed between June (during drought in 30-day PR) and August (after rewetting  
 417 in 30-day PR) of 2022, and were affected by PR, grassland management, species identity and soil  
 418 texture (Fig. 7; Table S9). In August, nitrogen nutrition was higher in 30-day PR vs 3-day PR for *Festuca*  
 419 ( $P = 0.007$ ; Table S9; Fig. 7) and in sandy-loam vs sandy soils within 30-day PR ( $P = 0.015$ ; Table S9; Fig.  
 420 7). Soils from permanent grassland increased NNI only in sandy-loam soils ( $P = 0.001$ ; Table S9; Fig. 7),  
 421 and this effect was strongest in August ( $P < 0.001$ ; Table S9; Fig. 7). Growing on sandy-loam soils  
 422 improved nitrogen nutrition most for *Lolium hybridum* ( $P = 0.002$ ; Table S9; Fig. 7) and differences  
 423 between species were on average greater in sandy-loam soils ( $P = 0.002$ ; Table S9; Fig. 7).

424 Differences in nitrogen nutrition among species were generally also more pronounced in the 3-day PR  
 425 and in June ( $P = 0.007$ ; Table S9; Fig. 7). Within the 3-day PR in June, *Festuca* and both *Lolium perenne*  
 426 variants had increased nitrogen nutrition compared to *Lolium hybridum*, and *Lolium perenne*  
 427 MELSPRING had increased nitrogen nutrition compared to *Festulolium* ( $P = 0.007$ ; Table S9; Fig. 7).  
 428 Under 30-day PR in June, only *Lolium hybridum* had significantly reduced nitrogen nutrition compared  
 429 to both *Lolium perenne* variants ( $P = 0.007$ ; Table S9; Fig. 7). In August, the opposite was seen, with  
 430 *Festuca* having the lowest nitrogen nutrition compared to all other species within 3-day PR and  
 431 reduced NNI compared to *Lolium perenne* MELSPRING in the 30-day PR ( $P = 0.007$ ; Table S9; Fig. 7).



432  
 433 **Figure 7.** Effects of precipitation regime (PR) on the nitrogen nutrition index in June and August 2022  
 434 per species, texture class (sandy vs sandy-loam), grassland management (permanent vs temporary  
 435 grassland) level. The error bars indicate +/- 1 SE on the mean. Species names including *Festuca*  
 436 *arundinacea*, *Festulolium*, *Lolium hybridum*, *Lolium perenne* MELFORCE (tetraploid) and *Lolium*  
 437 *perenne* MELSPRING (diploid) are abbreviated as FA, FL, LH Mel., LP Melf. and LP Mels., respectively.  
 438 Day of the year is abbreviated as DOY.



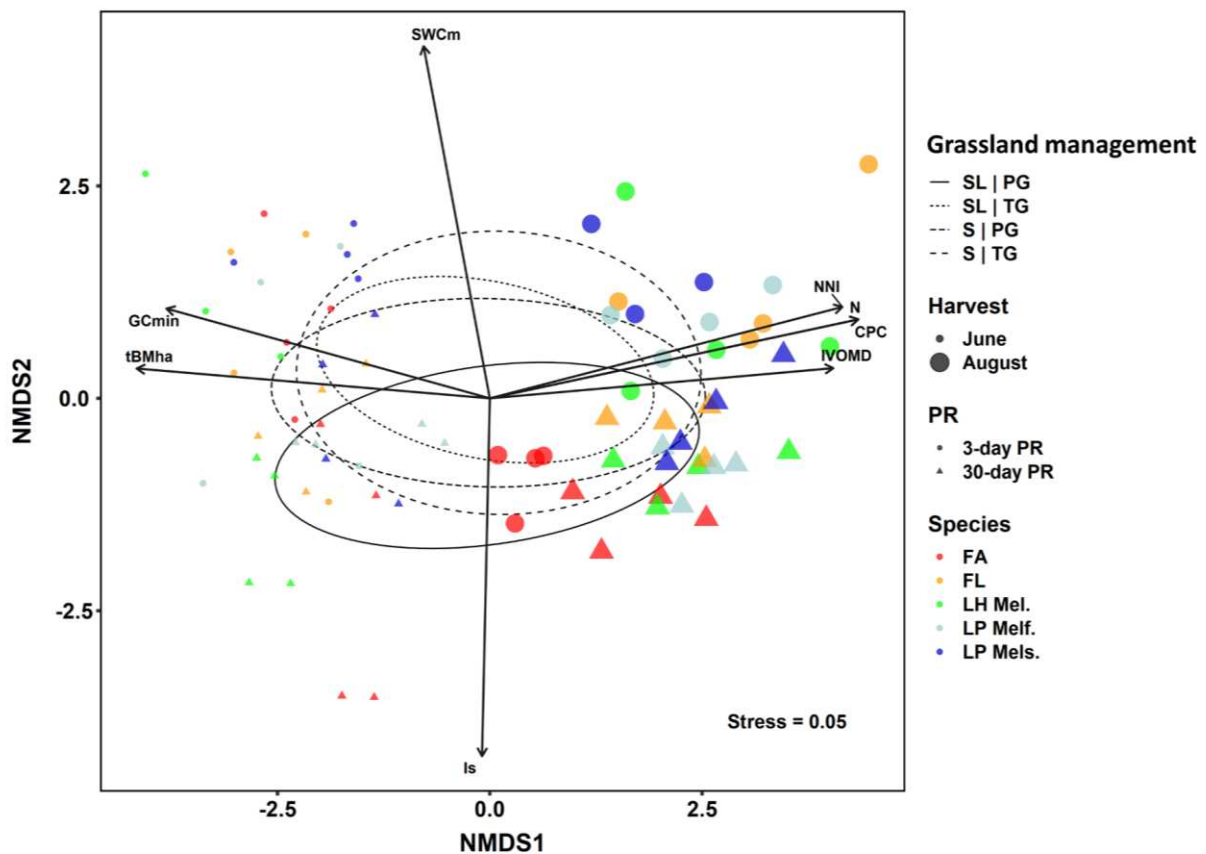
### 439 3.3.3. Interactions between water and nutrient limitations

440 Both nitrogen limitation and water availability together determined aboveground productivity and  
441 forage quality. More intense PR-induced soil drought reduced cumulative biomass on average  
442 significantly in 2022 but not in 2021 ( $P < 0.001$ ; Table S10; Fig. S10a). These effects also differed  
443 between soils and species, with the biomass of *Lolium hybridum* actually increasing with increasing  
444 drought severity in temporary grassland soils in 2021, and biomass of *Lolium perenne* MELFORCE  
445 decreasing most out of all species in permanent grassland soils in 2022 ( $P = 0.03$ ; Table S10; Fig. S10a).  
446 On a per harvest basis, differences among species followed similar trends ( $P < 0.001$ ; Table S10; Fig.  
447 S10b), with generally more positive effects of drought intensity on biomass production during autumn  
448 2021 and spring 2022 (wet and cool conditions) and more negative effects during summer 2022 (dry  
449 and hot conditions). Furthermore, more productive mesocosms were generally more nitrogen limited  
450 ( $P < 0.001$ ; Table S10; Fig. S11). The effects of drought intensity on biomass interacted with nitrogen  
451 limitation, showing a tendency for more negative effects of drought on productivity in mesocosms  
452 with low NNI, or, more positive effects of increased NNI on biomass in mesocosms under more intense  
453 drought ( $P = 0.001$ ; Table S10; Fig. S12a).

454 Regarding forage quality, NNI interacted with drought intensity to determine CPC, leading to a positive  
455 effect of drought intensity on CPC (strongest for *Lolium perenne* MELSPRING), but only under low  
456 nitrogen nutrition ( $P = 0.016$ ; Table S10; Fig. S12c). IVOMD was not influenced by drought intensity ( $P$   
457  $= 0.9$ ; Table S10; Fig. S12b). However, IVOMD generally related positively to NNI ( $P < 0.001$ ; Table S10;  
458 Fig. S12b) and this effect was strongest for *Lolium perenne* MELFORCE ( $P < 0.001$ ; Table S10; Fig.  
459 S12b).

460 Integrating our observations, PERMANOVA analysis revealed that across all factors, harvesting  
461 moment ( $R^2 = 0.73$ ;  $P = 0.001$ ; Table S11), followed by PR ( $R^2 = 0.05$ ;  $P = 0.001$ ; Table S11) explained  
462 most of the variance in observations, with generally higher soil drought intensity under 30-day PR, and  
463 low plant quality but high productivity and vice versa in June versus August 2022 (Fig. 8). Aboveground

464 productivity and minimum green cover during the period preceding harvest were overall positively  
 465 influenced by increased water availability but related negatively to NNI and plant quality, though more  
 466 negative to IVOMD than to CPC and N content (Fig. 7). Additionally, NNI was positively related to plant  
 467 quality and IVOMD declined least under reductions in soil water availability of all plant quality  
 468 parameters (Fig. 8). Separations among species were also visible ( $R^2 = 0.04$ ;  $P = 0.001$ ; Table S11) and  
 469 species identity interacted with harvesting moment ( $R^2 = 0.03$ ;  $P = 0.011$ ; Table S11; Fig. S13a),  
 470 because *Festuca* tended to perform largely equally than other species in June, while it was the most  
 471 productive under the most intense soil drought and of the least quality in August (Fig. 8; Fig. S13a).  
 472 Moreover, in June, *Lolium hybridum* clearly separated from *Lolium perenne* MELSPRING, with the first  
 473 being on average more productive under more intense soil drought and of less quality compared to  
 474 the latter (Fig. 8; Fig. S13a). Finally, species identity also interacted with PR, indicating that *Festuca*  
 475 was greener and more productive in 3-day PR but under most intense soil drought in 30-day PR (closely  
 476 followed by *Lolium hybridum*) compared to other species ( $R^2 = 0.02$ ;  $P = 0.026$ ; Table S11; Fig. S13b).



477

478 **Figure 8.** Nonmetric multidimensional scaling ordinations of mesocosms with soil water content  
479 sensor (N = 40) based on Euclidian dissimilarities of harvests in June (during drought) and August (post-  
480 rewetting) of 2022. Species names including *Festuca arundinacea*, *Festulolium*, *Lolium hybridum*,  
481 *Lolium perenne* MELFORCE (tetraploid) and *Lolium perenne* MELSPRING (diploid) are abbreviated as  
482 FA, FL, LH Mel., LP Melf. and LP Mels., respectively. Sandy-loam soils, sandy soils, permanent  
483 grasslands and temporary grasslands are abbreviated by SL, S, PG and TG, respectively. Based on  
484 permutation tests (n=999), significant vector fits were found for all ordination variables ( $R^2 \geq 0.75$ ;  $P$   
485 = 0.001) including mean soil water content (SWCm), minimum green cover (GCmin) and drought  
486 intensity (Is) in the period preceding harvest, total aboveground biomass per ha (tBMha), N nutrition  
487 index (NNI), as well as, plant N (N), crude protein content (CPC) and in vitro organic matter digestibility  
488 (IVOMD). The estimated vectors of CPC and N content overlap.

#### 489 **4. Discussion**

490 Recent observations indicate that summer weather persistence is increasing in the mid-latitudes  
491 (Zolina *et al.*, 2013; Coumou *et al.*, 2018; Pfleiderer *et al.*, 2019), with profound consequences for the  
492 ecosystem functioning of unfertilized temperate grasslands (Reynaert *et al.*, 2021; Reynaert *et al.*,  
493 2022; Li *et al.*, 2023b; Zi *et al.*, 2023a). We explored if and how such altered weather regimes with  
494 longer dry and wet spells influence the productivity and quality of intensively managed monoculture  
495 forage grasslands across soil textures and whether management aiming at elevated SOC and choice  
496 of species or cultivars buffer those influences. In line with Grant *et al.* (2014) and Fariaszewska *et al.*  
497 (2020), our results indicate that increased persistence in precipitation regimes (PR) leads to more  
498 intense soil drought which reduces the yield of forage grass species. *Festuca arundinacea* and  
499 *Festulolium* remained the most productive species overall, and *Lolium perenne* MELSPRING (diploid)  
500 the least. Forage quality was overall neutrally or positively influenced by increased PR persistence in  
501 30-day PR, with the highest quality found for *Lolium perenne* MELFORCE (tetraploid). Our results also  
502 revealed species-specific trade-offs between plant quality and productivity under increasing drought

503 stress, and differences in nitrogen nutrition during 30-day PR drought and upon 30-day PR rewetting.  
504 Moreover, forage grass responses to increased weather persistence under varying levels of grassland  
505 management related SOC differed between soil textures. Although elevated SOC generally improved  
506 plant quality, its buffering effects on drought damage, productivity and nutrient availability were  
507 found consistently only in sandy-loam soils but not sandy soils.

#### 508 **4.1. Differences in growth strategy determine productivity and quality of forage** 509 **grass under increased weather persistence**

510 In line with our first hypothesis, subjection to precipitation regimes with longer dry and wet spells  
511 reduced productivity and greenness of all cultivars by reducing water availability (Hofer *et al.*, 2017;  
512 Meisser *et al.*, 2019; Reynaert *et al.*, 2022). As expected, either *Festuca* or *Festulolium* were the most  
513 productive species across both years under the more severe PR, confirming their overall greater  
514 productivity under more irregular water availability compared to *Lolium* variants. This was likely  
515 related to differences in root architecture, stomatal regulation and water potential adjustment under  
516 drought (Durand *et al.*, 1997; Ebrahimiyan *et al.*, 2013; Coughon *et al.*, 2017; Fariaszewska *et al.*, 2017;  
517 Becker *et al.*, 2020; Curran *et al.*, 2020; Fariaszewska *et al.*, 2020). Within the *Lolium* varieties, *Lolium*  
518 *hybridum* and *Lolium perenne* MELFORCE extracted more water from the soil, leading to elevated  
519 biomass or improved plant quality during some harvests compared to *Lolium perenne* MELSPRING.

520 However, our second hypothesis was only partially confirmed. In contrast to previous experiments  
521 (Fariaszewska *et al.*, 2017; Fariaszewska *et al.*, 2020), *Festuca* outperformed other species most in the  
522 establishment year. While the improved drought resistance of *Festuca* usually shows only from the  
523 second year after establishment due to slow early sward development (Fariaszewska *et al.*, 2020), the  
524 lack of extreme ecological soil drought in any treatment in 2021 and our method of manually  
525 transplanting individuals may have allowed *Festuca* to effectively utilize its other superior resource  
526 extraction traits already in 2021 when all species still had immature rooting systems. However, rather  
527 unexpectedly, *Festuca* did not outperform others under the most intense droughts in 2022 (June and

528 September), whereas *Festulolium* outperformed or did equally well as *Lolium* variants in those cases.  
529 For *Festuca* this may be attributable to the mesocosm containers simulating relatively shallow soils,  
530 preventing this species to access water in deeper soil layers (Fariaszewska *et al.*, 2017), while  
531 *Festulolium* is able to extract water well in shallow layers (Durand *et al.*, 1997). Nonetheless, *Festuca*  
532 generally showed the most rapid greenness recovery after harvest or drought (excluding August 2022)  
533 and also had greater nutrient supply in the soil and higher nitrogen nutrition in 30-day PR upon  
534 rewetting. Moreover, *Festuca* generally showed the ability to reduce soil water most compared to all  
535 other species across seasons and treatments, without this necessarily negatively affecting its green  
536 cover, productivity or quality (again excluding Aug 2022). In line with previous studies, this indicates  
537 that *Festuca* generally outperforms other cultivars under ample water availability and mild drought  
538 stress (Neal *et al.*, 2009; Cougnon *et al.*, 2014; Becker *et al.*, 2020; Fariaszewska *et al.*, 2020). In  
539 addition to greater water and nutrient use efficiency, these differences may be related to its  
540 crown/root metabolism (Perlikowski *et al.*, 2020; Perlikowski *et al.*, 2023) or greater root biomass and  
541 coarser structure (Cougnon *et al.*, 2017), either improving plant performance directly by facilitating  
542 water and nutrient uptake (Cougnon *et al.*, 2014) or indirectly by increasing SOC (Hayashi *et al.*, 2023),  
543 boosting soil nutrient and water availability in the long term. However, despite the use of mesocosms  
544 limiting rooting depth (Fariaszewska *et al.*, 2017), the low water availability and reduced greenness in  
545 Aug 2022, and the comparatively low yields in June and September 2022 under 30-day PR indicate  
546 that this rapid conversion of available resources into biomass may potentially also reduce yield when  
547 prolonged extreme drought conditions are recurrent within one growing season and plants may not  
548 have enough stored reserves to properly recover (Van Eekeren *et al.*, 2010; Cougnon *et al.*, 2017).  
549 Indeed, despite a lack of direct effect of PR on NNI and thus N limitation to growth under drought in  
550 June 2022 (Hofer *et al.*, 2017; Meisser *et al.*, 2019), we found that increasing drought intensity  
551 generally reduced biomass more in cultivars with lower NNI, indicating that increased soil N availability  
552 or plant N reserves could partially buffer impacts of drought on plant productivity (Hofer *et al.*, 2017).

553 Contrasting with earlier studies and our expectations(Grant *et al.*, 2014; Fariaszewska *et al.*, 2017;  
554 Fariaszewska *et al.*, 2020; Reynaert *et al.*, 2023b), increased PR persistence did not affect plant quality  
555 during drought in June 2022. However, following Reynaert *et al.* (2023b), 30-day PR induced a slight  
556 increase in IVOMD (and nitrogen nutrition for *Festuca*) in August 2022 after rewetting and the  
557 following plant recovery. Nonetheless, it is important to note that declines in leaf non-structural  
558 sugars and increased leaf lignin contents with more persistent PR (Zi *et al.*, 2023b) may nullify positive  
559 effects of increased digestibility on plant quality.

560 Given the strong negative effect of flowering and grass maturity on plant quality (Nelson & Moser,  
561 1994), the high stem to leaf ratio for all species during the relatively late June harvest may have  
562 precluded detection of quality differences in response to increased PR persistency among them. In  
563 line with our second hypothesis, plant quality was generally highest for both *Lolium perenne* variants  
564 (particular *Lolium perenne* MELFORCE) across both PR in June and August of 22. However, rather  
565 surprisingly, nitrogen nutrition and plant quality (mostly digestibility) was significantly greater for  
566 *Festuca* compared to *Festulolium* and *Lolium hybridum* in June. This could be related to its generally  
567 reduced productivity due to comparatively rapid water limitation in May/June resulting in  
568 concentration of nutrients (Grant *et al.*, 2014) and/or its earlier flowering coinciding with the May  
569 harvest (7<sup>th</sup> May on average) in comparison to *Lolium* variants (26<sup>th</sup> of May on average) (Nelson &  
570 Moser, 1994; ILVO, 2022a; ILVO, 2022b).

#### 571 **4.2. Interactions between soil texture and grassland management modify species-** 572 **specific responses to increased weather persistence**

573 Given the contrasting responses of species to elevated SOC from permanent grasslands among texture  
574 classes, the species specific ability to withstand drought and convert more available nutrients and  
575 water from the soil into biomass likely determines whether grassland management or soil texture  
576 positively affects plant productivity and quality under increased PR persistence (D'Hose *et al.*, 2014;  
577 Buttler *et al.*, 2019). Indeed, positive effects of elevated SOC or finer texture on biomass seemed most

578 pronounced for *Festulolium* and *Festuca*, whereas positive effects on nitrogen nutrition and plant  
579 quality showed a tendency to be strongest for the most drought sensitive species (*Lolium perenne* and  
580 *Lolium hybridum*), likely because the more drought resistant other species had already extracted  
581 available nutrients and converted them into biomass prior to harvest (Grant *et al.*, 2014).

582 Overall, positive effects of elevated SOC on nutrient and water availability related to grassland  
583 management (Buttler *et al.*, 2019) were more pronounced in sandy-loam soils compared to sandy  
584 soils, only partially confirming our third hypothesis and contrasting with Minasny and McBratney  
585 (2018). Whereas plant greenness, water and nutrient supply was highest in sandy-loam soils with a  
586 permanent grassland history, nutrient supply and greenness were higher in sandy soil from the  
587 temporary grassland compared to sandy-loam soil from temporary grassland and sandy soil from  
588 permanent grassland. This was possibly due to historically increased nutrient availability as indicated  
589 by higher K and P levels compared to other soils at the start of the experiment, allowing plant  
590 productivity to compensate for expected differences in plant performance related to SOC (Buttler *et*  
591 *al.*, 2019) or texture classes (Dodd & Lauenroth, 1997). In contrast with Mallory and Porter (2007),  
592 these trends indicate that grassland management resulting in increased nutrient availability can  
593 override positive effects of elevated SOC on plant responses to climate change (Minasny & McBratney,  
594 2018), especially when SOC differences are not very pronounced (1.48 % in temporary grassland vs  
595 1.83 % in permanent grassland for sandy soils) or when soils are already relatively SOC saturated  
596 (Georgiou *et al.*, 2022). Nonetheless, positive effects of temporary grassland management in sandy  
597 soils became less pronounced in 2022, whereas elevated SOC consistently positively influenced plant  
598 quality across both texture classes under increased weather persistence, highlighting its importance  
599 for improving plant functioning under altered rainfall variability.

600 In contrast to Van Sundert *et al.* (2020), effects of preceding soil drought intensity on nutrient supply  
601 during post-drought rewetting were only clear for N in *Festuca* across soils. Whereas K supply was  
602 increased under 30-day PR vs 3-day PR for *Lolium perenne* (cfr. Van Sundert *et al.*, 2020), the opposite

603 was found for *Festulolium*. This may indicate that species-specific differences in belowground  
604 investment and resource extraction capacity can strongly influence the magnitude of transient  
605 nutrient pulses upon rewetting under altered rainfall variability (Borken & Matzner, 2009). Moreover,  
606 Birch effects may be generally less pronounced in highly fertilized grasslands (Van Sundert *et al.*, 2020)  
607 and diminish in strength over time under repeated dry-wet cycles, because of exhausted labile organic  
608 C pools (Birch, 1958) or gradual shifts in bacterial composition leading to adaptation (Borken &  
609 Matzner, 2009). Hence, effects of grassland management relating to SOC and historical nutrient  
610 availability may have been more important in determining nutrient supply than the imposed PR  
611 regime for all nutrients except for N and K. As expected, nutrient supply was highest in sandy-loam  
612 soils with high SOC (Borken & Matzner, 2009), but historically elevated P availability in the sandy soil  
613 from temporary grassland was likely able to override positive effects of elevated SOC or texture on  
614 nutrient supply from other soils, translating in comparatively high productivity. Interestingly,  
615 differences in Mn supply reflected general patterns in aboveground productivity related to soil texture  
616 and grassland management most accurately, largely independent of species identity or PR. This result  
617 may further highlight the importance of micronutrients in co-limiting grassland biomass production,  
618 even under high fertilization rates (Radujković *et al.*, 2021).

619 Finally, our study adds to growing evidence that weather regimes with longer and more extreme dry  
620 and wet spells may worsen soil quality and functioning (Borken & Matzner, 2009; Buttler *et al.*, 2019;  
621 Zhang *et al.*, 2019; Klaus *et al.*, 2020; Li *et al.*, 2023b). Despite boosting greenness, biomass and  
622 nutrient supply, sandy-loam soil from permanent grasslands with elevated SOC also showed a  
623 tendency for inducing the highest soil drought intensities (Fig. 8; Fig. S3). The main cause for this was  
624 likely the increased evapotranspiration of generally more productive plant communities with  
625 increased leaf area (Wang *et al.*, 2007), related to higher total water and nutrient availability in SOC  
626 enriched soils (Dodd & Lauenroth, 1997; D'Hose *et al.*, 2014; Buttler *et al.*, 2019). Nonetheless, soil  
627 water retention curves indicated that soils from permanent grasslands (and sandy-loam soil in  
628 particular) were most affected by increased weather persistence in terms of water holding capacity.



629 In contrast to what Buttler *et al.* (2019) found regarding aboveground responses, this may indicate  
630 that soils with high SOC content from permanent grasslands are more sensitive to change in  
631 precipitation regime because they have a greater potential of losing soil structure and aggregation (Lal  
632 & Shukla, 2004). Specifically, whilst such soils may improve growing forage grass yield and quality  
633 under climate change (Buttler *et al.*, 2019), they can simultaneously experience greater degradation  
634 of physical soil properties compared to soils from temporary grasslands which have little to lose in  
635 terms of structure (Lal & Shukla, 2004). Given that SOC rich soils have been associated with high fungal  
636 C (Liu *et al.*, 2023) which may improve soil aggregation and water holding capacity (Wilson *et al.*, 2009;  
637 Querejeta, 2017; Bowles *et al.*, 2018), it is also possible that a reduction of fungal diversity and  
638 abundance under increased PR persistence (Li *et al.*, 2023b) contributed to worsening soil hydrological  
639 properties (Querejeta, 2017).

## 640 **5. Conclusions**

641 Management strategies improving SOC show potential to sustainably improve forage grass  
642 performance under climate change and guarantee future ecosystem service provisioning (Buttler *et al.*,  
643 2019). Our results indicate that forage grass varieties differing in drought strategy and quality  
644 respond differently to increasing summer weather persistence due to differences in resource  
645 extraction and resource use efficiency, and that grassland practices resulting in elevated SOC may  
646 ameliorate these responses. In particular, *Festulolium* and *Festuca* overall retained the highest  
647 productivity under drought, whereas *Lolium perenne* MELFORCE (tetraploid) had the highest quality.  
648 While soils from permanent grasslands with elevated SOC buffered plant quality across texture  
649 classes, they also showed the steepest decline in soil water retention under increased PR persistence,  
650 and the productivity of sandy soils from temporary grasslands was unexpectedly high. Therefore,  
651 future studies should also consider other grassland management factors such as (micro)nutrient  
652 availability besides SOC when assessing forage grass and soil responses to climate change. In  
653 conclusion, permanent grassland soils with historically elevated SOC likely buffer negative effects of

654 increased summer weather persistence on forage grass performance, but may also be more sensitive  
655 to degradation under climate change.

## 656 **6. Declaration of Competing Interest**

657 The authors declare that they have no known competing financial interests or personal relationships  
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