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Great and blue tit laying dates vary with fine-scale variation in local tree composition but not tree budburst

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1	Great and Blue Tit laying dates vary with fine-scale
2	variation in local tree composition but not tree budburst
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4	Running head: linking bird phenology with individual trees
5	
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19	Author contributions: EM and FA conceived the study. FA provided the long-
20	term bird data. FA, PVK and KV provided tree data. EM performed all analyses
21	and wrote the manuscript. All authors provided input on the interpretation of the
22	results and on the draft manuscript.
23	

25 **Abstract**

26 Many studies have investigated how spring temperature affects laying dates and how this in 27 turn affects the synchrony between nestling food demands and the insect food peak that 28 follows tree budburst. While there is strong evidence that temperature itself acts as a cue for 29 this plasticity in annual timing, the exact nature of the cue and response remain to be 30 elucidated. Here we use long-term data on Great and Blue Tits and an unprecedented dataset on the location and phenology of 1396 trees to investigate whether small-scale variation in 31 32 laying date can be explained by local tree phenology, and/or by tree species composition around nestboxes. Individual trees maintained their relative timing of budburst between years, 33 34 while differences among tree species were more variable between springs. Contrary to expectation, we found no relation between first-egg dates and average budburst date at 35 36 different distances around the nestbox. This can at least partly be explained by the very low degree of spatial autocorrelation in tree budburst. We did find an effect of local tree 37 38 composition whereby Blue Tits, but not Great Tits, laid earlier in nestboxes surrounded by 39 more oaks and fewer beeches. Although Blue Tit nest failure rate was higher in territories 40 with more beech trees, independently of laying date, we found no evidence for preferential occupation of oak-dominated territories. Thus although we found some evidence for fine-41 42 scale effects of tree species on timing of breeding, the underlying mechanism remains unclear. Key words: phenology, synchrony, breeding success, Parus, Cyanistes 43

45 Introduction

46 Over the past few decades, evidence has accumulated for changes in the timing of bird 47 breeding seasons in response to global warming, and this has become one of the best-studied 48 ecological systems documenting climate change effects in natural populations (Crick et al. 49 1997; Parmesan 2007; Charmantier and Gienapp 2014; Radchuk et al. 2019). In particular, 50 effects of increasing spring temperature have been documented to lead to advances in 51 budburst of trees, emergence of caterpillars, and timing of breeding in forest-dwelling 52 insectivores that largely depend on these caterpillars as staple food for their nestlings. The 53 close link between spring temperature and the timing of these different trophic levels had 54 already been well documented before the impact of climate warming became visible 55 (Slagsvold 1976; Van Noordwijk et al. 1995).

56 In the past few decades, many studies have investigated how bird species and populations 57 may or may not be able to maintain synchrony between the timing of egg-laying and 58 subsequent hatching of nestlings, and the increasingly early insect food peak (Torti and Dunn 59 2005; Visser 2008), which in turn may affect population trends and even population persistence (Both et al. 2006; Møller et al. 2008; Miller-Rushing et al. 2010). Mean laving 60 61 dates in temperate environments generally covary with spring temperatures (Dunn 2004; 62 Parmesan 2007), in interaction with photoperiod (Lambrechts and Perret 2000; Gienapp et al. 2010). Indoor experiments on Great Tits have suggested that temperature itself, rather than 63 64 exposure to leafing trees, acts as a direct cue triggering egg-laying (Visser et al. 2009; 65 Schaper et al. 2011). Other studies have shown that temperature Nevertheless, our 66 understanding of how exactly temperature affects timing of egg-laying in the wild remains 67 incomplete (Schaper et al. 2012; Caro et al. 2013). Given that neither temperature nor tree budburst are amenable to large-scale experimentation in the field, and given the high degree 68 69 of temporal correlation between the two processes, the study of temporal variation has 70 obvious limitations in further elucidating how birds finetune their timing of breeding in 71 relation to lower trophic level phenologies (Caro et al. 2013). Moreover, remarkable 72 differences in response to spring temperatures have been shown among populations of the 73 same species in similar environments (Visser et al. 2003; Matthysen et al. 2011; Samplonius 74 et al. 2018), and even among individuals within a population (Nussey et al. 2005). Thus, it 75 seems likely that birds may use additional cues besides temperature to modulate their decision 76 to start laying (Bison et al. 2020).

As an alternative to temporal variation, the study of spatial variation in synchrony between

timing of breeding and tree budburst can provide valuable complementary information.

79 Several studies have examined spatial variation in bird phenology across geographic gradients

80 (Slagsvold 1976; Visser et al. 2003; Both et al. 2004; Møller et al. 2008; Grimm et al. 2015),

81 but comparatively few studies have examined spatial variation in phenology at smaller scales

82 within regions experiencing similar climates (Nilsson and Kallander 2006; Møller 2008;

83 Bourgault et al. 2010; Dekeukeleire et al. 2019; Shutt et al. 2019). Even fewer studies have

84 examined phenological variation within heterogeneous study sites and linked this to variation

85 in microclimate and/or tree composition (Nager and Van Noordwijk 1995; Cole et al. 2015;

86 Germain et al. 2015; Hinks et al. 2015; Amininasab et al. 2016). Understanding such fine-

87 scale variation in reproductive timing is not only valuable to further our understanding of the

88 phenological cues used by birds as well as timing constraints, but also to properly assess other

89 sources of variation, notably estimates of additive genetic variation (Germain et al. 2016).

90 Several studies have already demonstrated that within the same forest, individual trees show

91 consistent year-to-year variation in their relative timing of budburst (Crawley and

92 Akhteruzzaman 1988; Van Dongen et al. 1997; Wesolowski and Rowinski 2006; Cole and

93 Sheldon 2017; Delpierre et al. 2017) even though this may be modulated by variation in tree

94 age (Augspurger and Bartlett 2003) and consistency may vary among species (Wesolowski

95 and Rowinski 2006). Given this consistency and the longevity of individual trees, this implies

96 that past variation in tree phenology can be reconstructed from present observations, allowing

97 to link repeated observations of bird phenology over long time periods to tree phenology

98 (Hinks et al. 2015).

99 Thus, given the importance of synchronization for breeding success (Hinks et al. 2015), we

100 may expect birds to finetune their timing of breeding to local variation in tree phenology.

101 Even though onset of egg laying often precedes actual budburst, making it unlikely that this is

102 used as a direct cue (Nilsson and Kallander 2006; Schaper et al. 2011; Wesolowski and

103 Rowinski 2014) multiple cues may be available for birds to predict local variation in

104 budburst. For example, bud swelling may precede actual budburst by one to two weeks (Sass-

105 Klaassen et al. 2011; Puchalka et al. 2017). Timing of individual trees can be predicted by

106 phenology in the previous autumn (leaf senescence) (Marchand et al. 2020) when breeders

107 may already be present on their territories. Birds might even detect volatiles emitted by

108 developing buds (Piskorski and Dorn 2010) as they are able to do for caterpillar-infested trees

109 (Amo et al. 2013) or birds may cue on the development of herbivores synchronized to

individual trees (Van Dongen et al. 1997). Finally, birds may simply use their own experience
from the past breeding season (Grieco et al. 2002)or use social information from neighbours

112 (Helm et al. 2006).

113 So far, very few studies have examined the link between bird and tree phenology at small 114 spatial scales. The only study we are aware of is the one by Hinks et al. (2015) who found a 115 strong correlation between Great Tit (Parus major) laying dates and budburst of nearby pedunculate oak trees (Quercus robur) in Marley Wood, UK. This area is characterized by 116 117 strong spatial autocorrelation in budburst creating clusters of early and late trees. Since these clusters are of similar size as or even larger than the birds' territories, this system is highly 118 119 suitable to detect synchronization of birds and trees at the local scale (Hinks et al. 2015). In 120 the same study area, bird phenology also correlated with remote-sensing derived vegetation 121 phenology (including multiple tree species) at a larger scale (240m resolution compared to 122 20m)(Cole et al. 2015). However, such spatial clustering of early and late trees is not 123 necessarily characteristic for forests in general (Wesolowski and Rowinski 2006), and the 124 mechanisms behind individual variation in tree phenology often remain little understood 125 (Polgar and Primack 2011; Hinks et al. 2015). Thus, more studies are required to assess the 126 general validity of spatial synchronization between timing of bird breeding and tree budburst. 127 In this paper we present results from a long-term dataset of timing of clutch initiation in Great 128 and Blue Tits (*Cyanistes caeruleus*) in a study plot where data are available for all mature

and Blue 1its (*Cyanistes caeruleus*) in a study plot where data are available for all mature

129 trees on tree location (in relation to nestboxes), tree species, size and timing of budburst. We

130 first show how tree budburst varies among tree species and show that budburst is repeatable

between years at individual tree level. Secondly, we investigate whether variation in first-egg

132 dates is explained by tree species composition (Cole et al. 2015; Szulkin et al. 2015;

133 Amininasab et al. 2016) and/or by the timing of budburst of surrounding trees at multiple

134 spatial scales. For this we use budburst data collected in a single year, assuming that this

135 reflects relative tree phenology throughout the study period. Thirdly, we explore whether such

136 effects of tree composition and/or budburst variation are reflected in habitat preference (age of

137 birds, occupation rate, and between-year shifts in nestbox use) and ultimately in variation in

138 nest success rates and numbers of fledged young.

140 Materials and Methods

141 Study area and methods

142 Study area and tree composition

143 Data were collected in a study plot in northern Belgium used for a long-term population study 144 of Blue and Great Tits since 1979. The 12-ha plot (in previous studies referred to as plot B or 145 'Boswachter'; 51°16'N, 4°29'E) is part of a larger forested area of ca. 300 ha in size. It is a 146 mixed mature broadleaved stand dominated by pedunculate oak (henceforth 'oak') and beech 147 (Fagus sylvatica), with smaller numbers of other broadleaved trees and a few conifers. A 148 sparse subcanopy is present, mostly consisting of black alder (Alnus glutinosa), rowan 149 (Sorbus aucuparia), hazel (Coryllus avellana) and black cherry (Prunus serotina). The herbal layer is patchy and primarily consists of bracken (Pteridium aquilinum) and bramble (Rubus 150 151 spp.). In 2010, all large trees (diameter > 20 cm) in the plot were labelled, their diameter at 152 breast height (DBH) measured and tree species and tree condition noted. In 2017, a full resurvey of all labelled trees in the plot, with detailed tree positioning was performed using 153 154 Field-MapTM technology (www.fieldmap.cz, Kovacsova and Antalova 2010). All tree 155 diameters were also remeasured. The final dataset included 1558 individual standing trees 156 (alive or dead) in 2017. Of these 1558 trees the majority were oak (73%) and beech (17%), 157 with small numbers of northern red oak Quercus rubra (2.7%), silver birch Betula pendula 158 (2.0%), black alder (1.3%), hornbeam Carpinus betulus (0.8%), Norway spruce Picea abies 159 (0.7%) and larch Larix decidua (0.7%), none of the remaining species accounting for more 160 than 0.3% (Fig. 1, Online Resource Table S1).

161 Budburst data

Budburst data were collected for all deciduous trees in 2010, and for a selection of trees in 162 163 2009 and 2017. In 2010, all trees were visited 5-6 times between 23 March and 19 May, with 164 most frequent observations (three visits per tree) between 22 April and 6 May. Leaving out 165 dead trees and conifers, the dataset comprised 1396 individual trees. One overall score was 166 given per tree, on a scale from 0 to 6 (Online Resource Fig. S2). After inspection of the data, 167 we chose to define budburst date as the first observation day when the tree reached score 4. 168 For trees where score 4 was missing we interpolated between the dates when the nearest 169 scores were recorded (e.g. when scores 3 and 5 were recorded but not 4, we used the middle

- 170 date between them). For the earliest trees where the first observed score was 5, we arbitrarily
- 171 assigned budburst date two days earlier; for very late trees that had 3.5 as the maximum score,
- two days later. Six trees that reached a maximum score of 1 or 2 or showed no budburst at all,
- 173 were given a missing value; four of these trees were probably already in very poor condition
- as they turned out to be dead in 2017. In 2009 we recorded budburst on a single date and only
- 175 for a subset of trees holding nestboxes (20 april, 117 trees). On 21 april 2017 budburst was
- 176 recorded again for most of these trees (N = 100).
- Since results indicated that the relative timing of oak and beech trees differed between years, we consulted an external dataset for information on the relative timing of oak and beech over a larger span of years (Demolder et al. 2017). This information was used to test whether annual variation in the relative timing of oak versus beech could affect the relative timing of bird phenology in territories with more oak or beech trees. The data were collected since 2003 for ca. 40 trees of each species in two large forested areas (Meerdaalbos and Zoniënwoud) both at ca. 50km from the study area. From these data we derived for each year a value for the
- relative onset of leaf development of oak versus beech, which varied between 14 days in 2006
- and minus 7 days in 2012 (a negative value indicates that oaks were earlier than beech).

186 Bird breeding data

187 From 1997 to 2017, a constant set of 118 nestboxes were operational in the study plot. A third 188 of these had small (26mm) entrances (evenly spaced across the plot) allowing access to Blue 189 Tits only (Dhondt and Adriaensen 1999) (Fig. 1). Due to irregular spacing of the boxes, the 190 minimum distance between neighbouring boxes varied between 5 and 44m. Nestboxes were 191 visited approximately weekly from the onset of nest-building. First-egg dates were estimated 192 from the first observation of a partially laid clutch, assuming one egg was laid per day. Only first clutches were included in the study, discarding second clutches or replacement clutches. 193 All laying dates are expressed as days since the 1st of March. Occupation rates (disregarding 194 195 nestbox type) were rather low with on average 25% for Great Tits and 28% for Blue Tits, 196 based on all first clutches, including unidentified females. Occupation rates of the two types 197 were very similar (51 and 53% for small and large entrances, respectively) reflecting the high 198 proportion of Blue Tits breeding in large-entrance holes. We captured 80 to 90% of all parents 199 on the nest when feeding 8-day old nestlings, but also recorded the identities of many females 200 while incubating. Their age (first breeding season, or older) was derived from lifetime capture 201 histories and/or based on plumage. Nestlings were ringed close to fledging when they were

202 approximately two weeks old and their body mass was measured to the nearest 0.1g (either 203 individually, or total brood mass divided by number of nestlings). The majority of females 204 (75% of individuals used in this study, N = 778) used only a single nestbox for breeding in 205 their lifetime; for others this ranged from two to five. Clutches laid by unknown females were 206 excluded from the study (19.1% and 16.4% of Blue and Great Tits, respectively). Most of these were deserted or otherwise did not hatch any young (64 and 69% respectively); the 207 208 proportion of unknown females among nests with hatched young was therefore much lower 209 (6.9 and 5.1% respectively). There was no significant difference in laying dates between 210 known and unknown females (model with fixed species and random year effect, $t_{1269} = 1.76$, P 211 = 0.08) with unknown females laying on average 0.68 (+ SE 0.39) days later. The final dataset 212 contained 553 first clutches of Blue Tit, and 509 of Great Tit. The total number of first 213 clutches per nestbox over the study period varied between 2 and 17. Basic breeding and 214 population data for each species are reported in Table 1.

215 Tree characteristics around focal nestboxes

216 Since we had no data on actual territory boundaries, we used two different approaches to link 217 breeding data with characteristics of nearby trees (species, size and budburst date), in line 218 with earlier studies (Wilkin et al. 2007; Hinks et al. 2015). In the first approach we linked 219 nestboxes with the surrounding trees regardless of the position of nearby occupied nests, 220 assuming that birds would time their breeding to the characteristics of surrounding trees, and 221 not necessarily trees inside their own territory. We considered three different distances: 15, 25 222 and 35m. We chose this distance range because most nestboxes had only a few trees within 223 10m distance. On the other hand, nests were typically between 25 and 75m from the nearest 224 occupied nest of the same species. This approach resulted, for each radius, in a single set of 225 tree characteristics for each nestbox throughout the study ('nestbox variables'). However, this 226 approach neglects the fact that occupation rates varied strongly throughout the study, and thus 227 both the size and position of territories used by the breeding pair could differ substantially 228 between years. Therefore, in a second approach we assigned trees to putative territories 229 approximated by Thiessen polygons, in line with previous studies (Wilkin et al. 2007; 230 Schlicht et al. 2014; Hinks et al. 2015). Here we assume that timing of breeding is influenced 231 by trees within the pair's estimated territory. Within each year, all trees were allocated to the 232 nearest occupied nestbox. We did this allocation independently for Great Tits and Blue Tits as 233 these do not show interspecific aggression when territories overlap, even though there is some 234 exploitative competition for food affecting reproductive success (Dhondt 2012; Gamelon et

al. 2019). In order to avoid unrealistically large territories in low-density years and/or parts of

the area with low occupation, we set a maximum cut-off distance at either 25 or 35m

237 (Schlicht et al. 2014)). We did not include a lower cut-off distance (e.g. 15m) because at this

short distance there was hardly any difference with tree characteristics within a fixed 15m

radius. This approach resulted in different values for tree composition per nestbox and year

240 ('territory variables').

241 In both approaches, we characterized tree composition by summing basal area (calculated 242 from tree diameter at breast height) per tree species (including the nestbox tree) and expressed 243 this relative to the total basal area summed across all tree species. The latter was done to 244 allow comparison between the 'territory' and 'nestbox' approaches, and because we were 245 interested in the role of tree species composition and phenology rather than overall tree 246 density. We included all trees recorded in 2010, regardless of their condition in 2017 (dead or 247 alive) since most tree deaths occurred in the last third of the study period (i.e. after 2010). 248 Since relative basal areas of the most common tree species were significantly correlated, we 249 performed a Principal Component Analysis on basal areas for the four most common tree 250 species (oak, beech, red oak and alder) as well as the combined values for birch (Betula spp.) and conifers (several species). This was done at the three scales mentioned above for the 251 252 nestbox approach and for the two cut-off distances for the territory approach. Results were 253 very similar for all five PCA analyses (Online Resource section S3), with a first axis 254 contrasting the abundance of oak versus beech, the second mostly reflecting the abundance of 255 birch and/or conifers versus other tree species, and the third reflecting the abundance of red 256 oak. Since the latter had an eigenvalue only slightly above one and represented a single and 257 not very abundant tree species, we did not consider it further.

258 For budburst data, we similarly calculated the mean budburst date of all trees within the 259 radius (nestbox-level) or within the putative territory (territory-level) as we did for tree 260 species composition. For this we only used the data of 2010, assuming that these reflect the 261 relative timing of trees throughout the study period. We also calculated the mean for oak and 262 beech trees separately. In addition, we calculated the percentage of early trees, to test the 263 hypothesis that laying date would be affected by the earliest trees rather than by average 264 budburst. 'Early trees' were identified based on the frequency distribution of budburst for all 265 trees (Online Resource Fig. S4). This showed that a large number of trees (about one third) all 266 had their budburst within a few days (from 27 to 30 April) while before this date there was

- 267 more variation. Therefore we considered early trees as those having budburst up until 26
- April. These represented 26% of all trees, but only 10% of oaks (Online Resource Fig. S4).
- 269 Overall, correlations between the nestbox-level and territory-level characteristics were high
- 270 (Pearson's r from 0.86 to 0.98, N = 1355). As expected, correlations were somewhat higher
- for the 25m distance comparisons, than for the 35m comparisons (details not shown).

272 Statistical analysis

All analyses were performed in SAS v9.4 (SAS Institute, Cary, NC). Principal Component
Analysis was performed using proc PRINCOMP and spatial autocorrelation of tree budburst
with proc VARIOGRAM.

276 We used generalized linear mixed models (proc MIXED) to model variation in laying date in 277 relation to local tree composition and mean budburst. All linear models were run with normal 278 error structure, and residuals were checked for normality. We ran separate models using either 279 tree variables at nestbox level (one value per nestbox) or at territory level (one value per 280 nestbox per year; see above) (overview in Online Resource Table S5). Species were analysed 281 separately but to verify differences in outcomes we ran additional models on the joint dataset 282 to test for interactions between species and the variable of interest. All models included age 283 of female (first-year or older; cf. Jarvinen 1991; Reed et al. 2013) as fixed effect, and year, 284 nestbox and female ID as random effects. For tree composition we always included the first 285 two axes of the PCA.

286 Inspection of the data showed that the first axis of tree composition was highly correlated 287 with mean budburst date of all trees, for all distances considered (Pearson's r = 0.55 to 0.63, 288 N = 118 nestboxes; see also Online Resource Table S6). Therefore we did not include these 289 variables in the same model. In a first set of models (1-6; Online Resource Table S5) we 290 included all nestboxes and analyzed laying date in relation to either local tree composition or 291 mean budburst date, and repeated this at three different scales (15m, 25m, 35m). In the 292 second set of models (7-9) we focused on territories dominated by oak trees, since these are 293 by far the most common trees in the study area. Here we only included nestboxes if at least 294 50% of the basal area (within the relevant distance class) consisted of oaks. In these models 295 we used mean budburst of oak trees as explanatory variable, as well as the two tree 296 composition variables, since these were not correlated (all r between 0.05 and -0.13). All 297 previously described models were run again at the territory level instead of nestbox level,

- with cut-off distances of 25m and 35m (see above) (models 10-15). Model results were
- 299 compared with a null model without any data on tree composition or budburst (but containing
- 300 female age as well as the above mentioned random effects).
- 301 We also tested alternative models using data from older birds only (not in Table S5), to test
- 302 the hypothesis that birds would need a learning phase to adjust their timing to local tree
- 303 phenology (Nager and Van Noordwijk 1995; Grieco et al. 2002), and hence the influence of
- tree phenology might be only or more clearly expressed in older birds. We also re-ran all
- 305 models using the percentage of early trees (defined as above) instead of mean budburst date.
- 306 Since neither of these models showed any substantial difference with the outcomes of the
- 307 main models, we will not report any of the details.
- 308 For additional analyses we also used linear models either using PROC MIXED or PROC
- 309 GLIMMIX for non-normal error distributions, as explained in the results.
- 310

311 **Results**

312 Variation in budburst

In 2010, trees reached a budburst score of 4 on average on 29 April (i.e. day 58; SD = 7.7, N

= 1396) with an estimated range from 31 March to 21 May (Online Resource Fig. S4). There

315 were significant differences among tree species ($F_{4,1356} = 163.4$, P < 0.001) with birches being

earliest (16 April on average), followed by red oaks (20 April), alders (24 April), beeches (25

April) and oaks, the latter being considerably later than all others (on average 3 May; all

species comparisons P < 0.05 except red oak – alder and alder - beech) (Online Resource Fig.
S7).

320 Relative timing of budburst of individual trees was repeatable among years, as shown by

321 highly significant correlations between the estimated budburst date in 2010 and the budburst

322 score on a single date in either 2009 or 2017 (Table 1). Fig. 2 illustrates this with the

323 correlation for oak trees between 2009 and 2010. Correlations were substantially stronger for

324 oak trees only than for all tree species combined, reflecting that timing of tree species differed

among years (Table 1). Notably, beech trees were on average earlier than oaks in 2009 and

326 2010, but later in 2017 (details not shown).

327 Spatial autocorrelation in budburst date was weak for all tree species combined (Moran's I =

328 0.05, Geary's c = 0.96; both P < 0.001) and close to non-existent when only considering oaks

329 (Moran's I = 0.017, P < 0.01; Geary's c = 0.986, P > 0.3) (Online Resource Table S8).

330 Variation in laying date

Annual laying dates varied between 2 and 28 April per year and species with an overall mean

of 13 April for Blue Tits, and 16 April for Great Tits (Table 2). Within a year and species,

laying dates typically ranged over a period of two to three weeks. In 2010, the year when

detailed budburst data were collected, mean laying dates were 14 and 15 April, respectively.

335 At that time most trees were still in the initial stages of budburst or had not yet commenced,

as the proportion of trees with a score of at least one (visible bud swelling) was only 20% on

337 8 April and still only 66% on the next scoring visit on 22 April (n = 600 and 469 trees

338 scored).

- 339 While a long-term advancement of laying date has been documented in this population for
- both species since 1979 (Visser et al. 2003; Matthysen et al. 2011), there was no significant
- 341 change over the time period in this study (Matthysen, unpubl. data).
- 342 All models of laying date showed highly significant effects of female age ($P \le 0.001$ in nearly
- 343 all models; see Table 3 for estimates in the best model). As expected from earlier work on the
- 344 same population (Matthysen et al. 2011) older females started laying on average one to one
- and a half day earlier in both species.
- 346

347 Correlations between laying date, tree composition and budburst

348 In the Great Tit, none of the models showed any significant contribution of tree composition 349 or mean budburst date at any of the specified scales, neither at the nestbox level or at the 350 territory level (all P-values > 0.5). Consequently, the best model for Great Tits with the lowest

351 AICc value was the null model with age as the only fixed effect (Online Resource Table S9).

352 For Blue Tits, none of the models showed any significant contribution of mean budburst date 353 (all P > 0.1). Most models showed significant contributions of the first principal component of 354 tree composition (Online Resource Table S9); not surprisingly, these were in particular the 355 models that included all nestboxes (i.e. not restricted to nestboxes surrounded by oaks). A few 356 models showed a weak but non-significant contribution of the second principal component of 357 tree composition (P = 0.06 to 0.07). Regression coefficients for tree composition were 358 negative in all cases, showing overall strong support for earlier laying in nestboxes 359 surrounded by more oak trees (PC1) and much weaker support for earlier laying when 360 surrounded by more birches and conifers (PC2). The best model for Blue Tits included tree 361 composition in a 25m radius around the nestbox (Online Resource Table S9). This model 362 performed considerably better than the null model containing only female age as fixed effect $(\Delta AICc = 10.3)$ and also better than similar models with tree composition within 15 and 35m 363 364 $(\Delta AICc = 3.3 \text{ and } 2.6, \text{ respectively})$. The model was also highly similar to the territory-level 365 model with 25-m level cut-off ($\triangle AICc = 0.6$). This is not surprising since the PC1 values at 366 nestbox and territory level were highly correlated at this distance (r = 0.986). Since the 367 contribution of PC2 was not significant in either model, we removed PC2 and again found 368 slightly better support for the nestbox-level model than the territory-level model ($\Delta AICc =$ 369 1.6). We thus chose the model with tree composition within a fixed 25m radius as the best

- 370 model for Blue Tits (details in Table 3). According to this model, laying dates advanced with
- about 2.5 days over the entire range of tree composition (Fig. 3). The difference between Blue
- and Great Tits was partially confirmed by a joint model which showed a nearly significant
- interaction between species and tree composition ($F_{1,545} = 3.11$, P = 0.078; this model also
- 374 included a random year*species term). In all further analyses, unless otherwise mentioned,
- tree species composition refers to the first principal component in a fixed 25m radius.

376 Further analyses on tree composition

We explored the effect of tree species composition on Blue Tits in more detail by replacing 377 378 tree composition with basal area of either oak or beech within 25m. This showed that either 379 tree species explained significant variation in laying date (both P < 0.001), which is not surprising given the high correlation (r = -0.88). We further tested whether the effect of tree 380 381 composition on laying date was modulated by the relative timing of oak versus beech. 382 Specifically, we predicted that the later laying in beech-dominated territories would be less 383 pronounced or even reversed in years when beech trees were particularly early. We tested this 384 by adding the annual timing difference between beech and oak ('beech earliness') from a 385 different study area (see Methods) to the final Blue Tit model (only for years 2003-2017). 386 This showed that overall, laying dates were later in years when oaks were relatively late 387 compared to beeches ($F_{1;12,9}$ =15.2, P = 0.002). The main effect of tree composition was also 388 maintained ($F_{1;114} = 6.9$, P = 0.01), but in contrast to our prediction, there was no interaction 389 between local tree composition and beech earliness ($F_{1:280} = 0.00$, P > 0.9). Finally, since 390 earlier analyses showed that laying dates advanced progressively throughout the study period, 391 we tested whether the relative timing in oak versus beech territories also changed over the 392 study period; indeed we found a significant advancement of laying date over time (year 393 included as continuous variable: $F_{1:548} = 19.0$, P < 0.001) but no interaction between tree 394 composition and time ($F_{1:548} = 0.9, P > 0.3$).

395 Habitat selection and fitness

For additional analyses on habitat selection and fitness in relation to tree composition, we focused on Blue Tits but will give some results on Great Tits for comparison only. First we checked whether effects of tree composition could be biased by age-dependent occupation of territories. This was not the case, as mean tree composition did not differ between young and older female Blue Tits at the population level (GLMM with tree composition as dependent

- 401 variable, individual age class (first year or older) as fixed effect, and year and bird identity as
- 402 random effects; $F_{1,385} = 0.0$, P > 0.9). We also found no preferential shift by individual
- 403 females towards nestboxes with more oaks, using a paired test on females that changed their
- 404 nestbox location from the first to the second breeding season (N = 63; paired t-test, t=-0.41, P
- 405 > 0.6). Tree composition also did not predict Blue Tit occupation rate per nestbox (proportion
- 406 of years with a first clutch; GLMM with poisson error, $F_{1,115} = 0.2$, P = 0.6).
- 407 Analyses on reproductive success showed that Blue Tit brood loss (i.e. the proportion of nests 408 without any fledged young) was higher in territories with fewer oaks and more beeches within 409 25m (Online Resource Table S10). This was largely explained by the effect on losses in the 410 nestling stage, while for pre-hatching losses there was a weaker and non-significant trend (Online Resource Table S10). Closer examination of the data with respect to actual tree 411 412 composition showed that nests with post-hatching brood loss were surrounded on average by fewer oaks (16.0 + 5.5 versus 17.3 + 5.1; n = 34 and 515) but in particular by more beech 413 414 trees (6.0 + 4.0 versus 3.3 + 3.5). We found no effect of tree composition on number of 415 fledglings or mean fledgling mass, although it did affect clutch size with larger clutches in 416 territories with more oaks (p = 0.03; Online Resource Table S10). For comparison, the same
- 417 analyses in Great Tits showed no relationships at all between breeding success and tree
- 418 composition (all P > 0.7; details not shown).
- 419 Finally, we tested whether the effect of tree composition on Blue Tit brood loss (post-
- 420 hatching) was modulated by nests being early or late, by adding laying date to the model
- 421 (centered within years). Early nests had lower failure rates ($F_{1,545} = 5.09$, P = 0.02) while the
- 422 effect of tree composition remained highly significant (P = 0.002) but there was no interaction
- 423 with tree composition (P > 0.9). Similar results were found for the model on overall nest
- 424 failure that included unknown females as well (interaction term P > 0.3).
- 425
- 426

427 **Discussion**

428 In this study we show that fine-scale variation in timing of budburst of individual trees does

429 not explain variation in laying date in either Great Tits or Blue Tits. However, variation in

- 430 local tree composition does explain variation in Blue Tit laying dates, with later first-egg
- 431 dates in territories containing fewer oaks and more beech trees.

432 Tree phenology

433 While multiple studies have investigated phenology of individual trees, we are aware of few 434 studies that have mapped budburst timing of individual trees at the level of entire stands. We 435 found high between-year consistency in the relative timing of budburst of individual trees, 436 confirming previous studies on pedunculate oak (Crawley and Akhteruzzaman 1988; Hinks et 437 al. 2015; Delpierre et al. 2017) as well as other tree species (Wesolowski and Rowinski 2006; 438 Cole and Sheldon 2017). This individual consistency was most pronounced when only oak 439 trees were considered. The lower repeatability for all tree species combined can be explained 440 by shifts in relative timing among tree species between years; indeed, beech trees also showed 441 significant repeatability in budburst between years, but the relative timing of oak and beech 442 trees varied among years. These findings are in agreement with Delpierre et al. (2017) who 443 also showed comparable repeatability values for pedunculate oak, red oak and beech, while 444 Wesolowski and Rowinski (2006) found higher repeatability in oaks than other broadleaved trees. The fact that we found high repeatabilities of budburst (at least within species) and over 445 446 multiple years strongly supports our assumption that relative timing of trees within one study 447 year may be extrapolated over longer time periods, and allows us to use long-term population 448 data to examine bird-tree synchrony.

449 In contrast to Hinks et al. (2015), however, we did not find any spatial autocorrelation in 450 timing of neighbouring trees, thus no clustering of trees with synchronous phenology. This can be explained by both abiotic and biotic differences between our study area and the Marley 451 452 Wood site studied by Hinks et al. (2015). Our study area is comparatively small with a flat 453 topography and homogeneous soil conditions, while the Wytham Woods of which Marley 454 Wood is a part, is a larger forest with a complex geology and topography, showing more than 455 100m difference in elevation between high and low parts (Wilkin et al. 2007). Moreover, oak 456 tree density in Marley Wood is much lower with few but very large oaks interspersed in other 457 vegetation; according to Hinks et al. (2015) only 118 trees were monitored in a 28-ha forest,

- 458 compared to over 1000 trees in our 12-ha study area. Topography and tree density in our
- 459 study area are more in line with the study area in Białowieża forest analysed by Wesolowski
- 460 et al. (2006) where there was no spatial concordance in the timing of different tree species.
- 461 The latter study does not provide details, however, on local patterns of phenology of
- 462 neighbouring trees of the same species, particularly oak trees.

Finally, there are detailed historic sources confirming that the oak and beech trees at the forest site in study were originally planted. For the oaks, written sources dating back as far as the beginning of the 15th century mention both local breeding and purchase of oak saplings; for beech this is the case from the 18th century onwards (Adriaenssens and Verheyen 2013). This may have led to a mixture of trees with different regional provenances and different phenology.

469 Variation in laying date

470 We did not find any evidence for spatial synchrony between laying dates and budburst of 471 trees surrounding individual nestboxes. This lack of synchrony contrasts with several other 472 studies (Møller 2008; Bourgault et al. 2010; Cole et al. 2015; Hinks et al. 2015). However, all 473 of these studies were performed in substantially different conditions from ours. Møller et al. 474 (2008) examined the relation between barn swallows (Hirundo rustica) and local tree and 475 flower phenology by comparing colonies across farms in a large study area (45 km²). 476 Bougault et al. (2010) found a strong correlation between oak phenology and Blue Tit laying 477 dates by comparing forest stands at different elevations in Corsica. The only studies we know 478 of that have looked at bird-tree synchrony at the level of individual nests were both done on 479 Great Tits in Wytham Woods (Cole et al. 2015; Hinks et al. 2015). Hinks et al. (2015) found a 480 significant correlation between laying date and oak budburst at a small scale (20m) in an area 481 showing strong spatial autocorrelation in oak budburst, as mentioned earlier. In the same 482 forest and at a somewhat lower resolution (240m), Cole et al. (2015) found that the degree of 483 phenological matching between Great Tits and tree budburst depended on local tree 484 composition, being particularly strong in areas dominated by pedunculate oak and hazel. 485 The absence of spatial synchronization in our study may be explained by the near absence of 486 spatial clustering in phenology; thus, most nestboxes were surrounded by a mixture of early 487 and late trees, and most breeding birds would be exposed to an intermediate environment in

- 488 terms of tree budburst. So both from a proximate (available cues) and ultimate (available trees
- 489 for foraging) perspective, we could expect synchronization to be low compared to a forest

490 with clustering of early and late trees. Nevertheless, even if trees are mixed more or less

- 491 randomly, there is still variation among territories in average timing of budburst, especially at
- 492 small scales. One could argue that birds may not adjust their timing to the average phenology
- but, for example, only to the earliest trees. However, we did not find any difference in our
- 494 results when replacing average budburst date with a measure of the proportion of early trees
- 495 in the territory.

496 An alternative explanation for the lack of spatial synchrony is that birds do not respond 497 directly to individual tree phenology but to other cues predicting local tree budburst, notably variation in microclimate. Aviary experiments have indeed shown that an increase in 498 499 temperature, but not exposure to leafing branches, induces Great Tits to lay earlier (Visser et 500 al. 2009; Schaper et al. 2011; Schaper et al. 2012). This would imply that the spatial matching 501 between bird laying and tree phenology found in other studies may be driven by 502 microclimatic factors that affect both bird and tree phenology. However, the spatial synchrony 503 between Great Tit laying and oak budburst in Marley Wood could not be explained by 504 temperature or elevational differences (Hinks et al. 2015). A study on Blue Tits in deciduous 505 forest also suggested that vegetation and/or insect phenology, rather than temperature itself, 506 set the cue for timing of breeding across years (Thomas et al. 2010). Thus, more research is 507 needed to elucidate the cues driving small-scale variation in phenology at individual nest

508 level.

509 One general implication of our finding is that within this particular study area, females appear

510 to have limited options in achieving phenological synchrony by shifting towards sites with

- 511 earlier or later phenology. Previous analyses in our population and others (Charmantier et al.
- 512 2008; Matthysen et al. 2011; Vedder et al. 2013) have shown considerable among-year
- 513 plasticity in individual females that allows them to track annual changes in tree and caterpillar
- 514 phenology. Clearly, females at this research plot will not be able to enhance their synchrony
- 515 by moving into earlier territories in warm springs.

516 Laying date and tree composition

517 Consistent variation in laying date among forest types has been very well documented in hole-

- 518 nesters, in relation to factors such as general tree species composition (Nilsson and Kallander
- 519 2006; Szulkin et al. 2015), evergreenness (Bourgault et al. 2010; Szulkin et al. 2015),
- 520 elevation (Nager and Van Noordwijk 1995) and urbanization (Dhondt et al. 1984;
- 521 Chamberlain et al. 2009; Vaugoyeau et al. 2016) but typically this variation has been studied

522 at the between-site rather than the within-site level. Some exceptions are the studies by 523 Szulkin et al. (2015) who found a correlation between Blue Tit laying and local variation in 524 evergreen versus deciduous trees, by Wilkin et al. (2007) who found that Great Tits laid 525 earlier in oak-dominated, more south-facing and low-altitude territories within Wytham 526 Woods, and by Arriero et al. (2006) showing later laying by Blue Tits in territories with more 527 immature or degraded oaks. Our finding that Blue Tits started laying earlier in nestboxes with 528 more mature oak trees nearby is in agreement with at least two other studies in mixed 529 temperate forests on Blue and Great Tits, respectively (Wilkin et al. 2007; Amininasab et al. 530 2016). Dekeukeleire et al. (2019) could not demonstrate such an effect of tree composition on 531 Great and Blue Tit laying dates in similar stands in Belgium as the ones we studied, 532 dominated by pedunculate oak, red oak and beech; however, their raw data show a similar 533 pattern as our study, with on average 4-5 days earlier laying in pure pedunculate oak stands 534 versus pure beech stands, and mixed stands being intermediate. The lack of significant 535 variation between their stand types may be explained by the inclusion of many different 536 intermediate stand types in their study (Dekeukeleire et al. 2019) which reduced the statistical 537 power. The effect size is somewhat larger than in our study, probably because our study are is 538 more mixed without stands of pure oak or beech.

539 We have no clear explanation of why laying is earlier in oak-dominated versus beech-540 dominated territories. This is clearly not explained by species-specific differences in budburst 541 (cf. Bourgault et al. 2010; Cole et al. 2015), since beech trees were earlier in most years. 542 Moreover, the difference between oak- and beech-dominated territories was unrelated to the 543 relative timing of oak versus beech trees in general. We also did not find that later laying in 544 beech-dominated areas would confer an advantage, since there was no interaction between 545 effects of laying date and tree composition on breeding success. Thus, we cannot show that 546 this variation in timing is adaptive.

547 An alternative hypothesis would be that Blue Tit females breed later in beech-dominated 548 territories because these are sites of intrinsically lower food quality, and therefore attract 549 females of lower phenotypic quality. This is supported to some extent by the lower failure rate 550 in territories with more oaks and fewer beech trees. This would agree with other studies 551 showing higher breeding success in forest stands or territories with a higher proportion of 552 oaks (Wilkin et al. 2009; Dekeukeleire et al. 2019). However, we note that the actual 553 difference in number of oak trees between successful and failed territories was rather small, 554 and that this difference seems to be driven more strongly by the number of beech rather than

555 oak trees. In addition, we found no association between female age and tree composition, nor 556 preferential shifts towards territories with more oaks by individual females. Moreover, 557 territories with more beech trees did not have broods with lower fledgling mass, which is 558 generally accepted as a proxy of food availability and a predictor of brood fitness (Tinbergen 559 and Boerlijst 1990; Matthysen et al. 2011). It is possible that females in territories with more 560 beech are constrained to lay early due to lower food availability early in the season, as 561 suggested by Wilkin et al. (2007) for Great Tits. Some support for this is given by the smaller 562 clutch size in Blue Tit territories with more beech trees, but it does not explain why such 563 territories had higher failures in the post- rather than the pre-hatching stage. By lack of data 564 on female foraging or condition in the egg-laying stage, we cannot test this hypothesis further. 565 We also have no clear explanation why Blue, but not Great Tits, delayed egg-laying in 566 territories with more beech trees. The two species are known to differ in foraging ecology 567 including tree species, tree parts and prey size (Gibb 1954; Betts 1955; Nour et al. 1998; 568 Slagsvold and Wiebe 2007). A recent study found Blue Tits to be more selective than Great 569 Tits in their choice of tree species in an urban environment, but neither species favoured oaks 570 in particular, and there was no clear link with breeding success (Mackenzie et al. 2014). There 571 is also no clear difference in home range size, although this has rarely been studied 572 (Naefdaenzer 1994), and if anything Blue Tits may have larger home ranges (Mackenzie et al. 573 2014).

574 Regardless of its causation, the fact that Great and Blue Tit respond differently to habitat 575 variation in terms of laying date has been reported earlier based on comparisons among 576 multiple forest types (Dhondt et al. 1984). Surprisingly, however, and despite the large 577 number of publications on breeding biology of each species, very few studies have addressed 578 these differences between Great and Blue Tits in recent decades. Dhondt et al. (1984) found 579 that Great and Blue Tit varied in the relative order of timing between habitats along a rural to 580 urban gradient. In an analysis across Europe, Vaugoyeau et al. (2016) also found habitat-581 related differences between the species, with Great Tits showing overall stronger variation 582 and being particularly late in coniferous habitat, while Blue Tits were even later in evergreen 583 habitat. It should be noted however, that this study included many sites where only one of the 584 species was reported. Nilsson & Källander (2006) found no differences between the two 585 species in their relative timing in oak/birch or coastal/inland sites. A few studies have also 586 reported differential responses over time of the two species in relation to spring temperature 587 (Wesolowski and Cholewa 2009; Vatka et al. 2014). In our own study population, however,

- 588 we showed earlier that Great and Blue Tits showed highly similar responses (Matthysen et al.
- 589 2011).

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- 598

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820

822 **Tables**

823 Table 1. Between-year repeatability of budburst date of individual trees ($r_s = Spearman rank$

824 correlation). Note that in 2009 and 2017 budburst scores were assessed on a single day, while

825 in 2010 multiple visits were made and used to estimate the date of reaching budscore 4. Thus,

826 negative correlation signs indicate that an earlier budburst in 2010 corresponds with a higher

- 827 budburst score on a fixed day in either 2009 or 2017.
- 828

	2009 vs 2010			2010 vs 2017			
	r _s	Р	Ν	r _s	Р	Ν	
All trees	-0.69	< 0.001	116	-0.42	< 0.001	100	
Oak	-0.81	< 0.001	89	-0.61	< 0.001	73	
Beech	-0.53	0.02	20	-0.47	0.03	20	

829

- 831 Table 2. Baseline breeding data of the two study species (first broods only). All values are
- based on annual means, subsequently averaged over the study period (N = 21 years). The
- range of annual means is given between parentheses. Data are given for all females, as well as
- 834 for known females only (data used for the actual analyses).

	G	reat Tit	Blue Tit		
	all females	known females only	all females	known females only	
Laying date (April)	16 (2 – 29)	16 (2 – 28)	13 (2 – 27)	13 (2 - 25)	
S.D. in laying date	4.92 (2.5 – 8.1)	4.88 (2.5-8.5)	4.72 (2.9 - 7.9)	4.41 (2.5 – 8.2)	
Clutch size	8.64 (6.8 – 11.0)	8.98 (7.5 - 11.0)	10.7 (8.4 – 12.0)	11.3 (9.9 – 12.8)	
Nest success rate	76.5 (38 – 100)	89.7 (60 - 100)	77.0 (61 – 91)	92.9 (81 - 100)	
Proportion 1 st year females	/	61.0 (34 - 79)	/	59.9 (39 - 81)	
Number of first clutches	29.0 (17 - 40)	24.2 (13 - 37)	32.6 (21 - 53)	26.3 (15 - 47)	

835

837 Table 3. Parameter estimates for fixed effects in the best model explaining laying date

- variation of Great and Blue Tits. In the Blue Tit model, tree composition represents the
- relative amount of oak versus beech trees in a 25m radius around the nestbox (first axis of a
- 840 Principal Component Analysis, see methods). Variance estimates for random effects were
- 841 41.7 (year), 14.5 (female identity), 1.02 (nestbox) and 12.0 (residual) for Great Tits. For Blue
- 842 Tits these were 32.1 (year), 10.8 (female identity), 0.8 (nestbox) and 9.8 (residual).
- 843

Effect	Estimate	Estimate refers to	DF	F	Р
Great Tit:					
Female age	1.12 <u>+</u> 0.43	First-year versus older	1; 281	6.86	<0.01
Blue Tit:					
Female age	1.66 <u>+</u> 0.38	First-year versus older	1; 411	19.5	< 0.001
Tree composition	-0.54 <u>+</u> 0.16		1; 99.7	12.1	< 0.001

845

846 Figure Legends

Figure 1. Overview of the study area. The left panel shows how the plot is divided into six

848 parcels bordered by lanes with predominantly oak trees. Symbols represent nestboxes as they

849 were present from 1997 to 2017, with either large (32mm; circles; N = 78) or small (26mm;

triangles; N =40) entrances. The right panel shows individual trees (N = 1558) by species

851 (green = oak, red = beech, orange = red oak, yellow = birch, grey = conifer, black = alder,

852 pink = other).

853

Figure 2. Repeatability of budburst for oak (black) and beech (white) trees between 2009 and

855 2010. Note that in 2009 all trees were scored on the same day, while in 2010 budburst date

856 (estimated time of reaching budburst score 4) was extrapolated from multiple observation

days. Bubble sizes reflect number of trees (from 1 to 7). Data points for beech are slighly

858 offset for clarity. Note that beech was markedly earlier in 2010, but much less so in 2009.

859

Figure 3. Relation between Blue Tit laying date and tree species composition (first principal

861 component axis at 25m scale; higher values mean an increasing number of oaks, and fewer

beech trees). The regression line and confidence interval are plotted through the raw data for

863 visual purposes, not taking into account random effects or age variation.

Figure 1













